# UNIVERSITÉ DU QUÉBEC EN ABITIBI-TÉMISCAMINGUE

## IMPACTS DE LA FRAGMENTATION SUR

# LA CAPACITÉ REPRODUCTRICE ET LA DIVERSITÉ GÉNÉTIQUE

# DE L'ÉRABLE À SUCRE (ACER SACCHARUM MARSHALL) AU QUÉBEC

THÈSE PRÉSENTÉE COMME EXIGENCE PARTIELLE DU DOCTORAT EN SCIENCES DE L'ENVIRONNEMENT

PAR

NOÉMIE GRAIGNIC

OCTOBRE 2014

Face à la roche, le ruisseau l'emporte toujours,

non pas par la force,

mais par la persévérance...

H. Jackson Brown

#### AVANT-PROPOS

L'objectif général de cette thèse visait à évaluer la capacité reproductrice et la structure génétique des populations d'érable à sucre au Québec; nous nous sommes intéressés à l'impact sur la diversité et structure génétique (1) de la fragmentation naturelle des populations situées à proximité de la limite nordique de l'aire de répartition de l'espèce et (2) de perturbations d'origine anthropique occasionnée par la coupe forestière telle que pratiquée actuellement au Québec. Le document est composé de six sections : La première partie est consacrée à introduire les thématiques de recherche à travers une revue de littérature et à définir les objectifs de travail. Dans la dernière section, nous avons discuté des principaux résultats et ouvert des perspectives sur des avenues de recherche. Les chapitres II à V forment le corps de cette thèse et ont été dédiés à la vérification et la discussion des hypothèses de travail. Ils ont été rédigés en langue anglaise sous la forme d'articles scientifiques. Les chapitres II et III ont été publiés et deux autres chapitres seront soumis prochainement à des revues avec comité de lecture.

Chapitre II: Graignic, N., Tremblay, F. & Bergeron, Y. (2014) Geographical variation in reproductive capacity of sugar maple (*Acer saccharum* Marshall) northern peripheral populations. *Journal of Biogeography*, **41**, 145–157 (Facteur d'impact (IF): 4,863)

Dans ce chapitre, nous avons comparé la capacité reproductrice de populations nordiques fragmentées situées dans l'aire discontinue de répartition de l'érable à sucre au Québec à celle de populations situées dans l'aire continue (non fragmentées). Chapitre III: Graignic, N., Tremblay, F. & Bergeron, Y. (2013) Development of polymorphic nuclear microsatellite markers in sugar maple (*Acer saccharum* Marsh.) using cross-species transfer and SSR-enriched shotgun pyrosequencing. *Conservation Genetics Resources*, **5**, 845–848 (IF: 0,708)

Nous avons développé vingt marqueurs microsatellites pour l'érable à sucre. Nous avons aussi mis au point les amplifications par multiplex (plusieurs paires d'amorces dans la même réaction).

Chapitre IV: Graignic, N., Tremblay, F. & Bergeron, Y. Genetic diversity and structure at the northern limit of a widespread North-American tree, sugar maple (*Acer saccharum* Marshall). (à soumettre à Molecular Ecology)

Dans ce chapitre, nous avons comparé la diversité et la structure génétique de populations nordiques fragmentées à celles de populations situées dans l'aire continue de répartition de l'espèce.

Chapitre V: Graignic, N., Tremblay, F. & Bergeron, Y. Genetic consequences of selection cutting on sugar maple (*Acer saccharum* Marshall). (soumis à Evolutionary Applications le 3 juillet 2014)

Nous avons étudié l'effet de la fragmentation anthropique sur la diversité et la structure génétique des populations situées dans l'aire continue de répartition. Nous avons comparé la structure génétique de peuplements traités avec la coupe de jardinage dans les années 1990 à celle de vieilles forêts.

La contribution des auteurs est la même pour tous les articles. Noémie Graignic, Francine Tremblay et Yves Bergeron ont défini la problématique de recherche et l'approche expérimentale. Noémie Graignic a fait l'échantillonnage et la collecte des données sur le terrain, les analyses au laboratoire, les analyses statistiques et a rédigé le manuscrit. Francine Tremblay a contribué aux analyses statistiques, révisé et corrigé les manuscrits. Yves Bergeron a apporté des commentaires sur les manuscrits.

#### REMERCIEMENTS

Je tiens en premier lieu à remercier ma directrice, Francine Tremblay, et mon codirecteur, Yves Bergeron, pour m'avoir accueillie au sein de leur laboratoire et de m'avoir donné la chance de travailler sur un sujet passionnant avec un « terrain de jeu » magnifique (tout le Québec). Je remercie Francine pour sa confiance, son soutien et toute l'aide qu'elle m'a apportés, surtout, pendant les corrections et pour répondre aux questions et exigences des réviseurs. Je remercie Yves pour sa porte toujours ouverte, sa présence perpétuelle virtuelle (toujours le mettre en copie conforme quand c'est important) et ses interventions ponctuelles, isolées, mais ô combien indispensables et pertinentes !

Je remercie aussi les membres du jury qui ont accepté d'évaluer ma thèse et pour leurs questions, corrections et suggestions: Annie DesRochers de l'Université du Québec en Abitibi-Témiscamingue (UQAT), présidente du jury; ainsi que les membres externes: Nathalie Isabel du Service Canadien des Forêts (SCF) et Antoine Kremer de l'Institut de Recherche en Agronomie (INRA) de Pierroton, France.

Je tiens à remercier le CRSNG (Conseil de recherches en sciences naturelles et en génie du Canada) d'avoir octroyé à Francine une subvention pour ce projet. Le CEF (Centre d'étude de la forêt) m'a permis de présenter lors de nombreuses conférences au Québec, de financer une conférence aux États-Unis (Botany 2012) et de me fournir une aide indispensable grâce aux professionnels des services aux membres, je remercie cette organisation. Je tiens particulièrement à remercier Mélanie Desrochers et Daniel Lesieur pour l'obtention de diverses bases de données, Marc J. Mazerolle sans qui mes articles n'auraient pas atteint ce niveau en statistique (voir pas le niveau du tout ! tu as été indispensable !) et enfin William F. J. Parsons pour les corrections en anglais de mon article dans *Journal of Biogeography*, ma note technique, le chapitre V et peut-être pour les articles suivants; ils sont tous professionnels au CEF et ils sont grandement appréciés pour leur travail et leur disponibilité.

Je remercie Marie-Hélène Longpré et Danielle Charron pour leur travail de logistique impeccable et leur dévouement, mon doctorat n'aurait pas été pareil sans vous. Je remercie mes aides de terrain, pour leur travail et pour m'avoir endurée un par un (jusqu'à 17 jours sans vraie pause) : Elizabeth Turcotte, Sylvain Larouche, Chelsea Archambault, Emilie Tarroux, Manuella Strukelj-Humphery, Hedi Kebli, Dave Gervais et pour finir la meilleure et la première : Geneviève Trudeau. Je remercie très sincèrement Arvin Beekharry d'être venu avec moi pour mes premiers sites et de m'avoir initiée à la forêt québécoise et à l'érable à sucre, sans toi j'aurais été perdue. Je remercie Norman Villeneuve et Claude Poulin du MRN (Ministère des Ressources naturelles) du Québec qui m'ont permis de trouver d'autres sites potentiels sur les cartes (la forêt non perturbée d'érable à sucre n'est pas si facile à trouver). Merci à eux et à Renaud Dostie de m'avoir permis d'aller avec Claude et Renaud sur mon site du Lac Pénobscot (2-D-A) qui est peut-être maintenant une EFE, ce fut très agréable. Michèle Bettez et Jean-Pierre Faust du Centre de semences forestières de Berthier du MRN du Québec ont fait les tests de germination, je les remercie.

Je remercie pour leur travail : Stéphanie Ferreira, Stéphanie Duthoy et toute l'équipe de GenoScreen en France (www.genoscreen.fr), ainsi que Valérie Catudal et Alexandre Belisle de Génome Québec (gqinnovationcenter.com).

Je tiens à remercier, tout particulièrement, Marie-Eve Sigouin de m'avoir très chaleureusement accueillie au Québec et à l'UQAT, tu fus la première Québécoise rencontrée, merci à toi pour tout ça ! Merci à Emilie Tarroux et Julien Moulinier de m'avoir grandement aidée au début de mes statistiques. Je remercie Hedi Kebli et Xavier Cavard pour la suite de l'aide en statistique. Je remercie aussi Véronique Paul

pour les données, les explications et le cours accéléré en ARCMAP, Ingrid Cea-Roa et Osvaldo Valeria pour leur aide ponctuelle sur ARCGIS et Igor Drobyshev pour BIOSIM. Je remercie Hedi Kebli, Mark Fox, Yi-Qin Zhou et leur directeur Pascal Drouin de l'URDAAT (Unité de recherche et de développement en agroalimentaire) de m'avoir permis d'utiliser leur matériel pour faire des gels d'agarose et pour les échanges d'autres matériels de biologie moléculaire. Je remercie Marie-Pier Ethier et l'URSTM (Unité de recherche et de service en technologie minérale) pour l'eau déminéralisée quand nous ne pouvions plus en faire nous-mêmes, et Mathieu Villeneuve et Janie Guimond Rousson de l'URSTM pour nous avoir prêté leur bonbonne d'azote liquide. Je remercie Françoise Bégin et Yves Poirier puis Nadine Neveu et Marc Boucher de l'UQAT, pour leur travail dans les commandes de produits. Je remercie Ann Gervais d'avoir simplifié la vie des étudiants étrangers à l'UQAT. J'en profite pour remercier tout le personnel de l'UQAT que j'ai fait travailler et que je n'ai pas cité. Je remercie Brian Harvey d'avoir discuté avec moi des types de coupes et Suzanne Brais pour son aide sur les types de sols (ce n'est pas dans cette thèse, mais qui sait? cet article verra peut-être le jour un jour). Je remercie Venceslas Goudiaby d'avoir répondu à mes diverses questions sur le doctorat à l'Université du Québec. Je remercie Marie-Pier Ethier pour m'avoir super bien logé pendant les quelques jours où je suis revenue à Rouyn pour présenter ma soutenance de thèse.

Je remercie mes amis d'Abitibi : Emilie Tarroux, Manuella Strukelj-Humphery, Marie-Pier Ethier, Julie Arseneault, Myriam Paquette, Samira Ouarmim, Marie-Eve Sigouin, Yi-Qin Zhou, Benoît Tendeng, Sylvie Gewehr, Xavier Cavard, Huaitong Xu, Mark Fox, Besma Bouslimi, Julien Moulinier, Caroline Trudeau, Véronique Paul, Joël Soucy, Jessica Smith... Je remercie aussi toutes les personnes avec qui j'ai discuté et qui m'ont encouragée : Hassan Bouzahzah, Thomas Genty, Aissa Zalagou Moussa Hamani, Virginie Derycke, Sandrine Picq... Je remercie Sandrine Picq, Mathieu Latutrie, Nahla Sellami et Simon Savinel pour leur vent de fraîcheur au labo. Je remercie mes amis de France qui sont venus me voir ici : Christelle et Johnny Hardelin et les autres pour leur soutien. Et enfin, le meilleur pour la fin, merci à Hedi Kebli de m'avoir soutenue, encouragée et surtout supportée !

Merci à ma famille de m'avoir permis de faire des études supérieures.

# TABLE DES MATIÈRES

AVA	NT-PROPOS	III
REMI	ERCIEMENTS	VI
LISTE	E DES FIGURES	XV
LISTE	E DES TABLEAUX	. XXI
RÉSU	JMÉ	XXVI
CHAF INTR	PITRE I ODUCTION GÉNÉRALE	2
1.1.	Effets de la fragmentation dans un contexte de changements climatiques	2
	1.1.1. Les populations nordiques marginales	2
	1.1.2. La coupe forestière	3
1.2.	L'érable à sucre (Acer saccharum Marshall)	4
	1.2.1. Description	4
	1.2.2. Aire de répartition	6
	1.2.3. Reproduction	8
	1.2.4. Reproduction végétative	10
	1.2.5. Croissance	10
	1.2.6. Les tests de provenances: premières observations des variations génétiques	11
	1.2.7. État des connaissances sur la génétique des populations naturelles	13
	1.2.8. Génétique et impacts anthropiques	17
	1.2.9. Migration post-glaciaire et phylogéographie	19

	1.2.10. Dépérissement, contraction ou expansion?	22
1.3.	Hypothèses, Objectifs et structure de la thèse	23
CHAF GEOC <i>SACC</i>	PITRE II GRAPHICAL VARIATION IN REPRODUCTIVE CAPACITY OF SUGAR MAPLE ( <i>ACER</i> HARUM MARSHALL) NORTHERN PERIPHERAL POPULATIONS	26
2.1.	Résumé	27
2.2.	Abstract.	29
2.3.	Introduction	31
2.4.	Materials and methods	33
	2.4.1. Study area	33
	2.4.2. Field sampling	35
	2.4.3. Data analysis	36
2.5.	Results	40
	2.5.1. Stand characteristics	40
	2.5.2. Sugar maple seed production	44
	2.5.3. Natural recruitment of sugar maple seedlings	45
	2.5.4. Relationships with climate	47
2.6.	Discussion	50
2.7.	Acknowledgements	54
2.8.	Biosketches	54
CHAF DEVE MAPI ENRIG	PITRE III ELOPMENT OF POLYMORPHIC NUCLEAR MICROSATELLITE MARKERS IN SUGAR LE ( <i>ACER SACCHARUM</i> MARSHALL) USING CROSS-SPECIES TRANSFER AND SSR- CHED SHOTGUN PYROSEQUENCING	55
3.1.	Résumé	56
3.2.	Abstract.	57
3.3.	Introduction	58
3.4.	Results & discussion	58
3.5.	Acknowledgments	62

-
---

GENE	TIC DIVERSITY AND STRUCTURE AT THE NORTHERN LIMIT OF A WIDESPREAD	
NORT	TH-AMERICAN TREE, SUGAR MAPLE ( <i>ACER SACCHARUM</i> MARSHALL)	. 64
4.1.	Résumé	. 65
4.2.	Abstract.	. 66
4.3.	Introduction	. 67
4.4.	Materials and methods	. 70
	4.4.1. Study area and sampling	. 70
	4.4.2. Molecular methods	. 72
	4.4.3. Markers genetic diversity	. 72
	4.4.4. Genetic diversity and differentiation between populations and groups.	. 73
	4.4.5. Quantitative relationship	. 73
	4.4.6. Correlation with sample size	. 74
	4.4.7. Proportions of rare alleles	. 74
	4.4.8. Bottleneck tests	. 75
	4.4.9. Isolation by distance	. 76
	4.4.10. Genetic structure	.77
	4.4.11. Gene flow and population sizes	. 78
4.5.	Results	. 79
	4.5.1. Nuclear microsatellite diversity statistics	. 79
	4.5.2. Genetic diversity and differentiation	. 79
	4.5.3. Quantitative relationship	. 82
	4.5.4. Correlation with sample size	. 82
	4.5.5. Proportions of rare alleles	. 85
	4.5.6. Bottlenecks	. 86
	4.5.7. Isolation by distance	. 88
	4.5.8. Genetic structure	. 88
	4.5.9. Gene flow and population sizes	. 89
4.6.	Discussion	. 90
	4.6.1. Genetic diversity levels and differentiation in sugar maple populations	90

	4.6.2. Genetic variation and the central-marginal hypothesis	94
	4.6.3. Special case of 1-D-C, Rémigny	96
	4.6.4. Mature tree and sapling cohorts	96
	4.6.5. Genetic signature of migration routes	97
4.7.	Conclusions	99
СНАН	PITRE V	
GENE	ETIC CONSEQUENCES OF SELECTION CUTTING ON SUGAR MAPLE (ACER SACCHAI	RUM
MAR	SHALL)	101
5.1.	Résumé	102
5.2.	Abstract.	103
5.3.	Introduction	104
5.4.	Materials and methods	107
	5.4.1. Study area and sampling	107
	5.4.2. Molecular methods	109
	5.4.3. Marker genetic diversity	109
	5.4.4. Genetic diversity and differentiation between cohorts and forest type	es110
	5.4.5. Partitioning of molecular variation	111
	5.4.6. Allele frequencies	111
	5.4.7. Bottlenecks	111
5.5.	Results	112
	5.5.1. Genetic variability of microsatellite markers	112
	5.5.2. Genetic diversity and differentiation between cohorts and forest type	es112
	5.5.3. Partitioning of molecular variation	114
	5.5.4. Allele frequencies	114
	5.5.5. Bottlenecks	118
5.6.	Discussion	118
	5.6.1. Genetic diversity in sugar maple populations	118
	5.6.2. Cohort differences	122
	5.6.3. Selection cutting influences and implications	123

5.7.	Conclusions	124
5.8.	Acknowledgements	125
CHAP	PITRE VI Slusion général e	126
CONC		120
6.1.	Effets des différents types de fragmentation	126
6.2.	Implications pour l'exploitation forestière et la conservation	128
	6.2.1. Implications de nos travaux de recherches pour l'aménagement forestier	128
	6.2.2. Intérêts pour les compagnies forestières	129
6.3.	Perspectives de recherche	130
	6.3.1. Élargir la zone d'étude & mariage à trois (écologie–génétique– palynologie)	130
	6.3.2. Quels marqueurs utiliser ? Vers la révolution du séquençage à haut débit	131
	6.3.3. Le séquençage à haut débit au service de la détection de gènes adaptatifs	133
ANNI SUPP	EXE A2 ORTING INFORMATION FOR CHAPTER II	135
ANNI SUPP	EXE A4 ORTING INFORMATION FOR CHAPTER IV	143
ANNI SUPP	EXE A5 ORTING INFORMATION FOR CHAPTER V	158
BIBLI	IOGRAPHIE	167

# LISTE DES FIGURES

Fig	Figure P	
1.1	Morphologie de l'érable à sucre ( <i>Acer saccharum</i> ). (a) érables à sucre adultes en forêt; (b) écorce grise et lisse d'un jeune érable à sucre d'environ 5 cm de dhp; (c) écorce d'un érable à sucre mature montrant les longues crêtes verticales profondes et irrégulières; (d) écorce écailleuse d'un érable à sucre de 50 cm de dhp; (e) feuilles d'érable à sucre au printemps; (f) schéma d'une feuille d'érable à sucre tiré de Farrar (1996). Photos: a, b, c et e, Noémie Graignic; d, Geneviève Trudeau.	6
1.2	Répartition de l'érable à sucre ( <i>Acer saccharum</i> ) (a) en Amérique du Nord (figure tirée de Godman <i>et al.</i> , 1990) et (b) au Québec avec les domaines bioclimatiques en légende, ceux de l'est sont représentés par une couleur plus foncée que ceux de l'ouest (Saucier <i>et al.</i> , 2003).	7
1.3	(a) fleurs d'érable à sucre ( <i>Acer saccharum</i> ) avec un pistil fonctionnel; (b) disamare: samares d'érable à sucre vertes jointes par deux; (c) samares d'érable à sucre mûres; (d) disamare d'érable à sucre mûre. a et b, photos tirées de Farrar (1996); c et d, photos de Noémie Graignic	8
1.4	Exemple de trappe à graine de $1 \text{ m}^2$ . (a) dispositif en Abitibi- Témiscamingue, fin octobre; (b) à la même période, il peut y avoir la première neige de l'année. Photos: Noémie Graignic.	9
1.5	Plantules d'érable à sucre ( <i>Acer saccharum</i> ) dans leur première année (issues de graines de l'année précédente) à la mi-juin 2009 avec (a) et sans (b) leurs cotylédons. Photos: Noémie Graignic.	10
1.6	Variation clinale du stade de coloration des feuilles d'érable à sucre ( <i>Acer saccharum</i> ) à l'automne selon leur provenance. Figure tirée de (Kriebel, 1957).	12
1.7	Localisation des études génétiques sur l'érable à sucre ( <i>Acer saccharum</i> ). L'aire de répartition de l'espèce est hachurée (Little, 1971). Les populations prélevées pour les tests de provenances sont représentées par un triangle; le site expérimental où le test a été effectué est représenté par une étoile; la couleur rouge a été utilisée pour l'étude de (Kriebel, 1957) et le bleu pour l'étude de (Ledig & Korbobo, 1983). Les études génétiques utilisant les allozymes et RAPD sont représentées par un cercle vert;	

- 2.2 Ordination diagram of species presence/absence (except sugar maple, Acer saccharum) by site for mature trees (M; n = 24) and saplings (S; n = 24) obtained by correspondence analysis (CA) in Québec. (a) By site: Circles, transect 1; squares, transect 2; triangles, transect 3; open symbols, discontinuous zone; black symbols, continuous zone; (b) by species. CA1, correspondence analysis axis 1; CA2, correspondence analysis axis 2. Species label abbreviations: Ab, balsam fir, Abies balsamea (L.) Mill.; Ap, striped maple, Acer pensylvanicum L.; Ar, red maple, Acer rubrum L.; Ba, yellow birch, Betula alleghaniensis Britt.; Bc, mountain paper birch, Betula cordifolia Regel; Bp, white birch, Betula papyrifera Marshall; Cc, bitternut hickory, Carya cordiformis (Wangenh.) K. Koch; Fg, American beech, Fagus grandifolia Ehrh.; Fn, black ash, Fraxinus nigra Marshall;

- 4.3 Migration pattern of four groups of Ouébec sugar maple (Acer saccharum) populations, using MIGRATE-N (Beerli, 2006). The groups (red lines for the northern group and green lines for other groups) are defined using the results of STRUCTURE. Mean mutation-scaled population sizes ( $\theta = 4N_{\rm e}\mu$ , where  $N_{\rm e}$  = effective population size and  $\mu$  = mutation rate per generation per locus) are given for each group. Arrows represent direction of migration, and the thicknesses of the arrows are proportional to mean mutation-scaled immigration rate ( $M = m/\mu$ , where m = migration rate; small arrow, M = 6.6-8.5; medium arrow, M = 15.0-17.8; big arrow, M =21-26). Locations of the 24 study sites (2-D-B was not included in genetics analysis) are given in red polygons for sites in discontinuous zone, and given in green polygons for sites in the continuous zone (see Graignic et al., 2014 for more details). Sugar maple-yellow birch (Betula alleghaniensis) bioclimatic domains are in gray hatch and the boundary of all bioclimatic domains are given in thin brown lines (Saucier et al.,

- A2.2 Sugar maple age structures of all 24 sites (n = 2880) in Québec. For each population, the linearized form of the power function is included, together with its coefficient of determination ( $r^2$ ).  $y_i = r M \times x_i$ , where  $y_i$  is ln(stems m<sup>-2</sup> + 1), r is intercept of the curve [ln(R), where R is initial recruitment], M is slope of the curve (mortality rate), and  $x_i$  is ln(age).........141

A2.3	Predicted means (± 95% confidence intervals) of sugar maple seedling	
	mortality rate $(n = 24)$ in Québec. Means with the same letter do not	
	differ at $\alpha = 0.05$ , but differ (with an asterisk) at $\alpha = 0.10$ . C, continuous	
	zone; D, discontinuous zone.	. 142

# LISTE DES TABLEAUX

# Tableau

1.1	Etudes génétiques sur l'érable à sucre ( <i>Acer saccharum</i> ) classées suivant l'ordre chronologique des publications et selon le type (1) de marqueur utilisé (M; Az, Allozyme; R, RAPD), (2) d'échantillon sur lequel l'étude a été effectuée et (3) de peuplement étudié (LN, limite nordique; PN, population naturelles; FCF, fragmentation du couvert forestier; CF, coupe forestière et A, autres types de peuplements). P, nombre de populations total; I, nombre d'individus total; $H_0$ , moyenne de l'hétérozygotie observée; $H_E$ , moyenne de l'hétérozygotie attendue; $F_{IS}$ , coefficient de consanguinité.	. 15
2.1	Parameter estimates and unconditional confidence intervals from multi- model inference of sugar maple ( <i>Acer saccharum</i> ) seedling density (ln(data + 1)) in Québec. Term abbreviations: m_BA, mature tree basal area; m_ers_BA, mature sugar maple basal area; s_BA, sapling basal area; s_ers_BA, sugar maple sapling basal area; July_T, July mean temperature; July_P, July mean precipitation.	.49
3.1	Primer sequences and characteristics of 23 microsatellite loci in two populations of <i>Acer saccharum</i>	.60
4.1	Genetic variability estimates of sugar maple ( <i>Acer saccharum</i> ) populations in Québec for mature trees, saplings or all individuals. IP, all individuals pooled; M, mature trees; Sa, saplings; N, sample size; <i>A</i> , mean number of alleles; <i>A</i> <sub>R</sub> , mean allelic richness; <i>H</i> <sub>O</sub> , mean observed heterozygosity; <i>H</i> <sub>E</sub> , mean expected heterozygosity; <i>F</i> <sub>IS</sub> , inbreeding coefficient; <i>F</i> <sub>ST</sub> , mean pairwise <i>F</i> <sub>ST</sub> . Mean was determined using AI except for $A_R^*$	. 80
4.2	Comparison of mean genetic variability estimates between populations of zones, transects and zones in each transect of sugar maple ( <i>Acer saccharum</i> ) in Québec for mature trees, saplings and all individuals	.83
4.3	Parameter estimates and unconditional confidence intervals from multi- model inference of sugar maple ( <i>Acer saccharum</i> ) $A_{\rm R}$ , $H_{\rm O}$ , $H_{\rm E}$ ( $\sqrt{\rm data}$ ), $F_{\rm IS}$ and $F_{\rm ST}$ in Québec. Asterisks identify parameters where confidence intervals excluded 0	. 84

4.4	Results of two-way ANOVA with interaction for percentage of allele frequency for the rare and private (frequency ( $f$ ) < 0.02; RP2; n = 23) category and three-way ANOVA with interaction for the percentage of allele frequency in the common and localized ( $f \ge 0.05$ ; CL5; n = 46) category prior to model simplification (model 1) and subsequent models (models 2 and 3).	86
4.5	Bottleneck results based on heterozygosity excess, Mode shift and <i>M</i> -ratio.	87
4.6	Isolation by distance analysis results performed separately for different geographic groups.	88
4.7	Comparison of genetic diversity using allozyme and microsatellite markers for old-growth, undisturbed or natural populations of tree species in northeastern North America, and <i>Acer</i> species around the world	92
5.1	Genetic variability estimates of sugar maple ( <i>Acer saccharum</i> ) populations in Outaouais, Québec, for mature trees (M), saplings (Sa), seedlings (S1), pooled individuals (PI) and only seedlings that emerged after logging (S2), separately. To compare allelic richness between each cohort (M, Sa, S1 and S2), we calculated allelic richness using those cohorts in the same database ( $A_R^*$ ). Populations were old-growth (OG) forest or had received a single selection cutting (SC) in end of 1990–beginning of 1991	113
5.2	Results of analysis of molecular variance (AMOVA) showing the partitioning of genetic variance among pairs, forest types and cohorts	116
5.3	Number of alleles per classes of frequency, per cohorts of the populations and grouping populations by forest types.	117
5.4	Bottlenecks results based on heterozygosity excess and mode shift	118
A2.	1 Summary of sample site coordinates, protection types, and site and stand characteristics. EFE, 'écosystèmes forestiers exceptionnels'; D, discontinuous; C, continuous; 1AY, undifferentiated till 50–100 cm thick; 1A, undifferentiated till; 1AM, undifferentiated till 25–50 cm thick; R1A, undifferentiated till < 50 cm thick. Data age class ecotone: VIN, old uneven-aged, more than three age classes and > 80-years-old; 70, 61–80 years old; 30, 21–40 years old.	136
A2.	2 Climatic data for the study area in the D (discontinuous) and C	

A2.2 Climatic data for the study area in the D (discontinuous) and C (continuous) zones along the three transects located in Québec, Canada. ..... 138

A2.3	Difference in Akaike information criterion ( $\Delta AIC_c$ ) and weights ( $\omega$ ) of age structure models (negative exponential and power function) for sugar maple seedling 5-year age-class frequency distributions in all sites ( $n = 2880$ ) located in Québec.	. 139
A2.4	Results of two-way ANOVA with interaction for mature trees $(n = 24)$ and linear mixed-effects models for saplings $(n = 111)$ in Québec, prior to model simplification. Num. d.f., numerator degrees of freedom; den. d.f., denominator degrees of freedom.	. 140
A2.5	Difference in Akaike information criterion ( $\Delta AIC_c$ ), weights ( $\omega$ ) and number of estimated parameters ( <i>K</i> ) for model comparisons of the relative importance of stand characteristics and climate conditions as predictors of sugar maple seedling density ( $n = 353$ ) in Québec. For brevity, only models with a cumulative $\omega \le 0.95$ were shown.	. 142
A4.1	Genetic variability estimates of microsatellite markers used in the study of sugar maple ( <i>Acer saccharum</i> Marsh.) in Québec.	. 143
A4.2	Summary of <i>P</i> -values for the Hardy–Weinberg equilibrium using GENEPOP (Marchov chain parameters: 10,000 dememorizations, followed by 500 batches of 5,000 iterations per batch).	. 144
A4.3	Summary of null allele frequencies for each pair of loci and populations using FREENA.	. 145
A4.4	Pairwise population $F_{ST}$ of (a) all individuals pooled, (b) mature trees and (c) saplings of sugar maple ( <i>Acer saccharum</i> ) for all 23 sites in Québec below the diagonal. <i>P</i> -values are found above the diagonal. Significant values after adjusted nominal level ( $\alpha = 0.05$ ) for multiple comparisons are given in bold.	. 146
A4.5	Comparison of mean genetic variability estimates between cohorts and cohorts in zones, transects and zones in each transect of sugar maple ( <i>Acer saccharum</i> ) in Québec.	. 149
A4.6	Difference in Akaike information criterion ( $\Delta AIC_C$ ), weights ( $\omega$ ) and number of estimated parameters ( <i>K</i> ) for model comparisons of the relative importance of stand characteristics and distance from northern limit as predictors of $A_R$ , $H_O$ , $H_E$ and $F_{IS}$ ( $n = 414$ ), and $F_{ST}$ ( $n = 23$ ) in Québec. For brevity, only models with a $\Delta AIC_C \leq 4.0$ and the next model were shown.	. 150

A4.7	Raw data (%) of allele classification based on the frequency and geographic distribution of all individuals from 23 Québec sugar maple ( <i>Acer saccharum</i> ) sites.	. 152
A4.8	Raw data (%) of allele classification based on the frequency and geographic distribution of each cohort for all 23 sugar maple ( <i>Acer saccharum</i> ) sites in Québec.	. 153
A4.9	Results of analysis of molecular variance (AMOVA) showing the partitioning of genetic variance among transects and zones.	. 156
A4.10	Mean mutation-scaled population size ( $\theta = 4N_e\mu$ , where $N_e$ = effective population size and $\mu$ = mutation rate per generation per locus) and mean mutation-scaled immigration rate ( $M = m/\mu$ , where $m$ = migration rate) between pairs of sugar maple ( <i>Acer saccharum</i> ) populations. Both parameters are estimated using MIGRATE-N (Beerli, 2006).	. 157
A5.1	Summary of sample site coordinates and protection types.	. 158
A5.2	Genetic variability estimates of microsatellite markers used in the Québec study of sugar maple ( <i>Acer saccharum</i> )	. 158
A5.3	Summary of <i>P</i> -values for Hardy–Weinberg equilibrium using GENEPOP (Markov chain parameters: 10,000 dememorizations, followed by 500 batches of 5,000 iterations per batch).	. 159
A5.4	Summary of null allele frequencies for each pair of loci and populations using FREENA.	. 160
A5.5	Pairwise population $F_{ST}$ (below the diagonal) of pooled individuals, mature trees, saplings, seedlings and post-harvest seedlings of sugar maple ( <i>Acer saccharum</i> ) for the 6 sites, in Québec	. 161
A5.6	Comparison of mean genetic variability estimates ( $A_R$ , $H_O$ , $H_E$ , $F_{ST}$ and $F_{IS}$ ) between old-growth (OG) and selection cutting stands (SC) of sugar maple ( <i>Acer saccharum</i> ) in Québec for mature trees (M), saplings (Sa), seedlings (S1), post-harvest seedlings (S2) and pooled individuals (PI) separately, between cohorts (M, Sa and S1, and M, Sa and S2) for PI, OG and SC separately, and between OG and SC for each pair of stands. Analyses were performed using FSTAT and tested for significance using 1,000 permutations.	. 162

xxiv

A5.7	Results of linear mixed-effects models for genetic variability estimates	
	(n = 432) in Québec, prior to model simplification.	. 164

RÉSUMÉ

Cette thèse visait à mieux comprendre la dynamique de répartition de l'érable à sucre (*Acer saccharum* Marshall) à sa limite nordique au Québec en évaluant la capacité reproductrice et la structure génétique des populations. L'érable à sucre atteint sa limite de répartition continue nordique dans le nord-est du Canada à la transition entre la forêt feuillue tempérée et la forêt boréale mixte. Nous avons étudié : (1) le potentiel de régénération et la diversité génétique des populations naturelles fragmentées situées à la limite nordique discontinue de répartition de l'érable à sucre et (2) la structure génétique des peuplements affectés par une perturbation de type anthropique, la coupe de jardinage.

Le territoire d'étude était localisé entre 45 ° 51 '-48° 59' N et 70 ° 21'-79 ° 27 'O et a été divisé en deux zones, continue et discontinue, sur la base de l'abondance des peuplements d'érable à sucre dans chacune des zones. La zone continue se situait dans le domaine bioclimatique de l'érablière à bouleau jaune et la zone discontinue se trouve dans les domaines bioclimatiques de la sapinière à bouleau jaune et de la sapinière à bouleau blanc. Vingt-quatre sites répartis le long de trois transects latitudinaux (situés dans l'ouest, le centre et l'est du Québec) ont été utilisés. Nous avons examiné la capacité reproductrice, et la diversité et structure génétique de ces populations. En ce qui a trait à la capacité reproductrice, nous avons analysé la structure d'âge des peuplements, la production et la germination des graines d'érable à sucre, de même que la densité et la structure d'âge des plantules. Les analyses génétiques ont été réalisées avec des marqueurs microsatellites développés spécifiquement pour cette espèce. Elles ont été réalisées sur 2 cohortes : jeunes arbres et arbres matures. L'impact de la coupe de jardinage sur la structure génétique des peuplements d'érable à sucre a été analysé dans trois paires de peuplements situés dans la zone de répartition continue du transect central. Chaque paire était composée d'une forêt ancienne et d'un peuplement coupé en 1990-1991. Nous avons comparé la diversité génétique de ces peuplements pour trois cohortes : plantules, jeunes arbres et arbres matures.

Les résultats ont montré que (1) la régénération de l'érable à sucre est bonne dans tous les sites étudiés et (2) que cette espèce possède une grande diversité génétique et une faible différenciation entre les populations, sur l'ensemble du Québec. Les peuplements d'érable à sucre avaient une structure d'âge inéquienne dans les zones continue et discontinue. Toutefois, nous avons observé moins de samares pleines et de plantules dans la zone discontinue par rapport à la zone continue sur le transect ouest. La limite nordique de l'érable à sucre est influencée en partie par le climat (années semencières et densité de plantules). La différence entre les zones se retrouvait aussi au niveau génétique. Ainsi, la diversité génétique était inférieure et la différenciation entre les populations était plus importante dans la zone discontinue par rapport à la zone continue sur l'ensemble du Québec et, plus particulièrement, sur le transect ouest (en Abitibi-Témiscamingue). Nous n'avons pas détecté de différences significatives d'un point de vue génétique entre les deux cohortes. Nous avons observé un motif de structure génétique particulier chez l'érable à sucre au Québec : les populations les plus nordiques formaient un groupe distinct et l'on a observé un flux génique plus important des populations du sud vers celles de nord. Les données génétiques et palynologiques indiquent que l'érable à sucre serait arrivé au Québec par deux voies migratoires, l'une en provenance de l'est et plus tardivement une seconde provenant du sud-ouest de la province. L'étude de la coupe de jardinage nous a révélé une différence significative entre les cohortes d'érable à sucre pour tous les types de peuplements : les arbres matures possédaient un taux d'hétérozygotie observée plus élevé comparativement aux jeunes arbres et aux plantules. Ceci pourrait indiquer une sélection naturelle en faveur des hétérozygotes dans le temps. La coupe de jardinage n'a pas eu d'effet important sur la structure génétique des peuplements. Toutefois on a détecté la présence d'un goulot d'étranglement dans tous les peuplements jardinés. Ceci pourrait refléter la perte de certains allèles à la suite du prélèvement d'une partie des arbres matures.

Dans un contexte de réchauffement climatique, il ressort clairement de cette étude que les populations nordiques d'érable à sucre au Québec ont le potentiel d'être à l'origine de l'expansion de cette espèce au nord de sa limite de répartition discontinue. Toutefois, avec leur patron distinct (capacité reproductrice et diversité génétique plus faibles), les populations au nord-ouest du Québec mériteraient une attention particulière. La détection d'un goulot d'étranglement dans les peuplements jardinés demanderait un suivi à plus long terme afin d'éviter une certaine érosion génétique suite à des coupes successives.

#### CHAPITRE I

#### INTRODUCTION GÉNÉRALE

#### **1.1.** Effets de la fragmentation dans un contexte de changements climatiques

### 1.1.1. Les populations nordiques marginales

Durant la dernière période interglaciaire de l'Holocène (à partir de 10 000 ans avant l'actuel (AA)), les essences forestières ont migré vers le nord, suivant le recul de la calotte glaciaire dans l'est du Canada (Liu, 1990; Prentice & Jolly, 2000; Williams *et al.*, 2000). La période Néoglaciaire (depuis environ 4 500 ans AA), qui a suivi, aurait été accompagnée d'un retrait vers le sud de certaines espèces méridionales pour atteindre leur limite de répartition actuelle (Liu, 1990; Richard, 1993). Ce retrait aurait laissé des populations résiduelles, isolées les unes des autres dans un couvert forestier formé d'espèces plus nordiques et dominé par les conifères. Le climat plus froid et sec en été à cette période, coïncide avec une augmentation des feux de forêt (vers 3 000 ans AA) et pourrait aussi avoir contribué à la fragmentation chez certaines espèces moins adaptées au feu de forêt (Carcaillet & Richard, 2000; Ali *et al.*, 2008). Les populations nordiques isolées et fragmentées d'espèces méridionales, pourraient également être issues d'évènements ponctuels de dispersion sur de longues distances (« sauts de puce ») lors de la migration vers le nord (Cain *et al.*, 2000). Nous sommes dans une période interglaciaire où un réchauffement climatique rapide du fait de l'activité humaine est établi (Hegerl *et al.*, 2007). Selon les prévisions de l'IPCC en 2007 (Intergovernmental Panel on Climatic Change : Groupement Intergouvernemental sur les Changements Climatiques), une augmentation de la température et des précipitations, respectivement de l'ordre de  $+2,1^{\circ}$ C à  $+5,4^{\circ}$ C (moyenne:  $+3.3^{\circ}$ C) et de -17% à +13% (moyenne: +1%), sont à prévoir dans l'est de l'Amérique du Nord durant la saison de croissance en 2080–2099 comparativement à 1980–1999 (Christensen *et al.*, 2007). Dans ce contexte, les arbres forestiers, et plus particulièrement l'érable à sucre, auront-ils la capacité de s'adapter à ce changement?

En Amérique du Nord, les modèles prédisent une migration vers le nord pour plusieurs espèces d'arbre, dont l'érable à sucre (Goldblum & Rigg, 2005; Iverson *et al.*, 2008). Afin d'affiner les modèles, il est important de comprendre l'influence de nombreuses variables comme le climat, les caractéristiques du peuplement, de l'espèce (capacité de reproduction et diversité génétique), le type de sol, la topographie, les autres espèces (compétition des autres végétaux, animaux se nourrissant de graines et les pathogènes) sur la répartition d'une espèce (McMahon *et al.*, 2011) et plus particulièrement sur les populations les plus nordiques (Parmesan *et al.*, 2005).

## 1.1.2. La coupe forestière

Le terme fragmentation est plus généralement associé à la fragmentation du couvert forestier et à la fragmentation de l'habitat. La fragmentation peut-être d'origine naturelle ou anthropique (Saunders *et al.*, 1991). De grandes ouvertures peuvent être créées à la suite de chablis ou de feux de forêt. La fragmentation d'origine anthropique, comme celle créée par l'urbanisation, l'agriculture ou l'exploitation forestière, est la plus étudiée pour l'érable à sucre (Young *et al*, 1993a; Young & Merriam, 1994).

Suivant la théorie, la fragmentation à long terme de petites populations entraînerait une réduction de la diversité génétique, une augmentation de la consanguinité intra-population et une augmentation de la différenciation génétique entre les populations comparativement aux grandes populations continues (Young *et al.*, 1996). De plus, dans les coupes forestières telles que pratiquées en Amérique du Nord, les arbres ont souvent été sélectionnés sur la base de certains critères qui mènent souvent à un appauvrissement de la qualité des tiges composant un peuplement (Majcen, 1994; Bouchard & Domon, 1997) et, éventuellement à un écrémage génétique (Nanson, 2004).

Les effets des coupes forestières se superposant aux changements climatiques, nous pouvons nous demander comment ces populations vont réagir. La capacité d'adaptation dépend en partie de la présence d'une grande diversité génétique (Hamrick, 2004).

#### 1.2. L'érable à sucre (Acer saccharum Marshall)

#### *1.2.1.* Description

L'érable à sucre fait partie du genre *Acer*, de la famille des *Aceraceae* et de l'ordre des *Sapindales* selon l'ancienne classification des angiospermes; et de la famille des *Sapindaceae* et de l'ordre des *Sapindales* selon les classifications phylogénétiques récentes de l'APG II (APG II, 2003). Le genre *Acer* compte 13 essences en Amérique du Nord dont 10 au Canada (Farrar, 1996).

L'érable à sucre, *Acer saccharum* Marsh., tire son nom de l'eau d'érable qui est sucrée et sert à faire le sirop d'érable. *Saccharum* vient du mot grec *sakcharon* qui signifie sucré ou doux et désigne la sève sucrée des noeuds de Bambou (Hosie, 1980). L'abréviation Marsh. vient de Humphry Marshall (1722–1801), botaniste américain rendu célèbre par son ouvrage intitulé « *Arbustrum americanum* », publié en 1785

(Hosie, 1980). L'érable à sucre est aussi appelé érable franc au Canada francophone (Farrar, 1996) et « sugar maple », « rock maple » et « hard maple » au Canada anglophone et aux Etats-Unis d'Amérique (USA) (Godman *et al.*, 1990).

Cette espèce est tolérante à l'ombre (Logan, 1965) et peut vivre jusqu'à 300–400 ans, atteindre jusqu'à 27–30 mètres de haut et 76–91 cm de diamètre à hauteur de poitrine (dhp) (Godman *et al.*, 1990) (Fig. 1.1a). L'écorce des jeunes tiges est grise et lisse (Fig. 1.1b). Elle devient ensuite gris foncé, crevassée, en longues crêtes verticales profondes et irrégulières (Fig. 1.1c) et parfois écailleuse (Fig. 1.1d). Les feuilles sont caduques, simples et opposées (Fig. 1.1e et 1.1f). Elles mesurent de 8 à 20 cm de longueur et sont un peu plus larges que longues. Elles ont cinq (parfois trois) lobes à longues pointes obtuses et ondulées de quelques dents irrégulières. Elles sont vertes et deviennent jaunes, orangé brillant ou rouge vif à l'automne avant de tomber ce qui donne le paysage de carte postale, si typique du Canada (Farrar, 1996). Sa sève est utilisée pour la fabrication du sirop d'érable et son bois pour la fabrication de planchers, pâte à papier et de bois de chauffage (Godman *et al.*, 1990).



**Figure 1.1** Morphologie de l'érable à sucre (*Acer saccharum*). (a) érables à sucre adultes en forêt; (b) écorce grise et lisse d'un jeune érable à sucre d'environ 5 cm de dhp; (c) écorce d'un érable à sucre mature montrant les longues crêtes verticales profondes et irrégulières; (d) écorce écailleuse d'un érable à sucre de 50 cm de dhp; (e) feuilles d'érable à sucre au printemps; (f) schéma d'une feuille d'érable à sucre tiré de Farrar (1996). Photos: a, b, c et e, Noémie Graignic; d, Geneviève Trudeau.

#### 1.2.2. Aire de répartition

L'érable à sucre est établi dans le monde seulement à l'est de l'Amérique du Nord. Son aire de répartition s'étend, pour sa limite nord, au Canada depuis l'extrême sud-est du Manitoba, jusqu'à la Nouvelle-Écosse. La limite sud traverse le centre du New Jersey, jusqu'à l'extrême sud du Tennessee, USA (Fig. 1.2a).

Plus particulièrement, au Québec, la limite nordique de l'érable à sucre se situe entre le domaine bioclimatique de la sapinière à bouleau jaune au sud et celui de la sapinière à bouleau blanc au nord (Saucier *et al.*, 2003) (Fig. 1.2b). Dans le domaine bioclimatique de l'érablière à bouleau jaune et plus au sud, l'érable à sucre est très fréquent. Dans le domaine bioclimatique de la sapinière à bouleau jaune, les

populations d'érable à sucre sont de moins en moins fréquentes, plus isolées les unes des autres jusqu'à la limite nordique; elles sont donc fragmentées.



**Figure 1.2** Répartition de l'érable à sucre (*Acer saccharum*) (a) en Amérique du Nord (figure tirée de Godman *et al.*, 1990) et (b) au Québec avec les domaines bioclimatiques en légende, ceux de l'est sont représentés par une couleur plus foncée que ceux de l'ouest (Saucier *et al.*, 2003).

#### 1.2.3. Reproduction

La floraison peut commencer après une vingtaine d'années. Cette espèce est monoïque (fleurs mâles et femelles sur le même arbre), les fleurs peuvent porter les deux sexes en même temps, mais généralement un seul des deux organes est fonctionnel (Fig. 1.3a). Les fleurs apparaissent à la fin mars – mi-mai selon la localisation géographique (Godman *et al.*, 1990). Le pollen est disséminé par le vent (anémophile) et les insectes (entomophile) avec une contribution plus importante du vent (Gabriel & Garrett, 1984). L'érable à sucre est dichogame (organes mâles fleurissent avant les femelles, protandrie et inversement, protogynie) créant ainsi une barrière à l'autofécondation (Gabriel, 1968). Dans le cas où une autofécondation a quand même lieu, un mécanisme d'avortement post-zygotique des ovules a été observé 2 semaines après l'autofécondation (Gabriel & Garrett, 1967).

Les fruits ou samares (Fig. 1.3b, c, d) se forment à la fin juin – fin juillet et ils commencent à tomber à la mi-août – fin septembre, habituellement juste avant que les feuilles ne tombent. Généralement, seulement une des deux samares jointes contient une graine (Godman *et al.*, 1990). Les ailes mesurent 30–35 mm de longueur (Farrar, 1996) et les graines, 7–9 mm (Godman *et al.*, 1990). Elles sont dispersées par le vent sur une distance d'une centaine de mètres (Johnson, 1988).



**Figure 1.3** (a) fleurs d'érable à sucre (*Acer saccharum*) avec un pistil fonctionnel; (b) disamare: samares d'érable à sucre vertes jointes par deux; (c) samares d'érable à sucre mûres; (d) disamare d'érable à sucre mûre. a et b, photos tirées de Farrar (1996); c et d, photos de Noémie Graignic.

Il existe des périodes de bonnes et de mauvaises années de fructification. Il y a environ une année semencière tous les 1 à 4 ans dans le centre nord du Wisconsin, 2 à 5 ans dans le reste des États-Unis et 3 à 8 ans au Canada (Wang, 1974; Godman *et al.*, 1990) (Fig. 1.4). Au nord du Michigan, jusqu'à 2 200 samares /m<sup>2</sup> ont été produites dans un vieux peuplement (Godman *et al.*, 1990). Dans une forêt du parc provincial du lac Supérieur en Ontario (limite nordique de l'érable à sucre), on a dénombré en moyenne de 16 à 277 graines pleines /m<sup>2</sup> entre 1992 et 1996, avec seulement une bonne année de fructification en 1996 (Kellman, 2004). Dans le centre nord du New Hampshire, des années semencières ont été identifiées en 1998, 2002 et 2006 pendant la période 1998–2008 (Cleavitt *et al.*, 2011).



**Figure 1.4** Exemple de trappe à graine de 1 m<sup>2</sup>. (a) dispositif en Abitibi-Témiscamingue, fin octobre; (b) à la même période, il peut y avoir la première neige de l'année. Photos: Noémie Graignic.

Les samares vont germer le printemps suivant leur dispersion. Il faut une période de froid de 1–3°C (Carl, 1983) et de l'humidité (Janerette, 1979) pour que cette germination soit optimale. Cette température est la plus basse enregistrée pour les feuillus d'Amérique du Nord (Godman *et al.*, 1990). Les tests de viabilités réalisés à l'automne montrent que 80–100% des graines sont viables (Kellman, 2004). Les tests de germination réalisés en laboratoire au printemps suivant la dispersion donnent jsuqu'à 77% de graines germées (Carl, 1983).

#### 1.2.4. Reproduction végétative

L'érable à sucre peut former des rejets de souche à l'occasion. Ce mode de régénération est plus fréquent dans les peuplements situés au nord de son aire de répartition (Godman *et al.*, 1990). Le drageonnement, depuis une racine, est rarement observé (Godman *et al.*, 1990).

## 1.2.5. Croissance

La période de croissance des plantules (Fig. 1.5a, b) commence à la mi-mai au nord du Michigan, avant que les feuilles de l'étage dominant sortent (Godman *et al.*, 1990). L'érable à sucre est une espèce tolérante à l'ombre (Logan, 1965), ce qui en fait une espèce compétitive. Environ 90% de la croissance en hauteur dans des peuplements denses et ouverts s'effectue respectivement en 18 et 24 jours. Il n'est pas rare d'avoir plus de 37 plantules m<sup>-2</sup>. Environ 50 % des plantules survivent à la première année, et une forte mortalité est observée les 4 années subséquentes (Godman *et al.*, 1990).



**Figure 1.5** Plantules d'érable à sucre (*Acer saccharum*) dans leur première année (issues de graines de l'année précédente) à la mi-juin 2009 avec (a) et sans (b) leurs cotylédons. Photos: Noémie Graignic.
La saison de croissance en hauteur et en diamètre de l'érable à sucre varie selon la localisation géographique et le climat (Godman *et al.*, 1990). Pour de nombreuses espèces arborescentes, la limite nord est modulée par un ensemble de conditions climatiques (principalement la température) et phénologiques qui convergent vers la réduction des chances de survie à différents stades du cycle de vie (Chuine & Beaubien, 2001). Par exemple, pour l'érable rouge (*Acer rubrum* L.), on observe une baisse du succès de reproduction sexuée qui entraîne une diminution de la capacité à se maintenir et à coloniser de nouveaux sites (Tremblay *et al.*, 2002).

### 1.2.6. Les tests de provenances: premières observations des variations génétiques

Un test de provenances consiste à comparer, dans des conditions environnementales similaires, le phénotype de plusieurs exemplaires d'arbres, issus de semences ou de boutures prélevées dans plusieurs régions couvrant l'aire naturelle de répartition de l'espèce (Wright, 1976). L'objectif d'un test de provenances est donc d'identifier les variations génétiques intra-spécifiques en relation avec les conditions environnementales.

Des tests de provenances ont été réalisés pour l'érable à sucre et ils ont montré une variation génétique pour de nombreuses caractéristiques. Kriebel (1957) a utilisé un total de 34 provenances (Fig.1.7), dont il a évalué les différences phénotypiques dans un test localisé en Ohio (USA). Plus précisément, l'auteur a montré qu'il existait (1) une variation clinale est-ouest de la sensibilité des feuilles aux dommages causés par une forte intensité lumineuse, (2) une plus faible résistance à la sécheresse pour les populations du nord par rapport à celles du centre et du sud, (3) une variation clinale nord-sud pour l'initiation de la saison de croissance avec les populations du nord qui débutent leur saison de croissance plus rapidement au printemps et (4) une variation clinale très marquée, nord-sud, avec la coloration des feuilles à l'automne qui est plus rapide pour les provenances du nord (Fig. 1.6). Ledig & Korbobo (1983) ont quant à eux travaillé sur des populations d'érable à sucre provenant d'un gradient altitudinal au New Hampshire (USA). Ils ont montré que les populations à la limite de répartition altitudinale, ont une photosynthèse et un taux de respiration plus élevés.



**Figure 1.6** Variation clinale du stade de coloration des feuilles d'érable à sucre (*Acer saccharum*) à l'automne selon leur provenance. Figure tirée de (Kriebel, 1957).

L'érable à sucre est diploïde avec 26 chromosomes (2n = 26) (Kriebel, 1957). Les espèces du genre *Acer* sont souvent diploïdes avec un nombre de chromosomes multiple de 13 (Darlington & Wylie, 1955). Il existe aussi des érables polyploïdes comme l'érable argenté en Amérique du Nord (*Acer saccharinum*) et un érable européen (*Acer pseudoplatanus*), qui sont tétraploïdes avec 52 chromosomes (Darlington & Wylie, 1955). L'érable à sucre et l'érable noir (*Acer nigrum*) sont compatibles et une hybridation est possible entre ces deux espèces (Kriebel & Gabriel, 1969). Grâce aux nouvelles techniques de génétique, il a pu être montré que les patrons génétiques (RAPD et ADN chloroplastique) sont très similaires entre les deux espèces (Skepner & Krane, 1997; Skepner, 1998).



**Figure 1.7** Localisation des études génétiques sur l'érable à sucre (*Acer saccharum*). L'aire de répartition de l'espèce est hachurée (Little, 1971). Les populations prélevées pour les tests de provenances sont représentées par un triangle; le site expérimental où le test a été effectué est représenté par une étoile; la couleur rouge a été utilisée pour l'étude de (Kriebel, 1957) et le bleu pour l'étude de (Ledig & Korbobo, 1983). Les études génétiques utilisant les allozymes et RAPD sont représentées par un cercle vert; associé à chaque cercle, le numéro de la référence (voir tableau 1.1 pour la correspondance entre le numéro et la référence) est indiqué et entre parenthèse le nombre de sites étudié à cette localisation géographique.

### 1.2.7. État des connaissances sur la génétique des populations naturelles

Les premières études sur la diversité et la structure génétique de l'érable à sucre, ont été réalisées dès le début des années 90 grâce à l'étude du polymorphisme des allozymes, un type de marqueur biochimique qui fait appel au polymorphisme des protéines (Tableau 1.1). Puis, depuis la fin des années 90, les marqueurs RAPD, qui analysent directement l'ADN, sont utilisés sur l'érable à sucre.

Beaucoup d'espèces d'arbres forestiers pollinisées par le vent ont une dispersion moyenne efficace ce qui maintient une diversité génétique importante dans les populations et une faible différenciation génétique entre elles (Hamrick *et al.*, 1992; Hamrick & Godt, 1996). La diversité génétique, basée sur l'analyse des allozymes, est modérée chez l'érable à sucre ( $H_E = 0,148$ ;  $H_O = 0,150$ ; Tableau 1.1) et est comparable aux autres espèces d'arbres (Foré *et al.*, 1992a).

Il existe une faible différenciation entre les régions dans l'aire de répartition de l'espèce. Elle représente moins de 2 % de la variation totale pour trois régions situées au nord (Wisconsin, USA), au centre (Illinois, USA) et au sud (Tennessee, USA) de l'aire de répartition (Gunter et al., 2000) (Fig. 1.7). Young et al. (1993b) ont étudié des régions très éloignées géographiquement au Canada, dont une à la limite nordique en Ontario. Ils n'ont trouvé qu'une faible distance génétique entre les populations. En théorie, les populations situées au centre de l'aire de répartition vont avoir la plus grande diversité génétique et les populations très proches des limites (nord et sud) vont avoir une plus faible diversité génétique (Hampe & Petit, 2005) (Fig. 1.8). Pour l'érable à sucre, la plus grande diversité génétique devrait, donc, se retrouver aux USA (états de l'Indiana, de l'Ohio et du Kentucky). L'étude sur des populations naturelles, basée sur l'analyse des allozymes, dans ce secteur est celle de Foré et al. (1992a) ( $H_E = 0,148$ ) et la diversité génétique n'était pas différente de populations situées plus au nord, au Canada, ( $H_E = 0,112$ ; Young et al., 1993b), et à proximité des limites nord ( $H_E = 0,110$ ; Perry & Knowles, 1989) et sud ( $H_E = 0,133$ ; Baucom et al., 2005, pour les vieux peuplements seulement).

**Tableau 1.1** Etudes génétiques sur l'érable à sucre (*Acer saccharum*) classées suivant l'ordre chronologique des publications et selon le type (1) de marqueur utilisé (M; Az, Allozyme; R, RAPD), (2) d'échantillon sur lequel l'étude a été effectuée et (3) de peuplement étudié (LN, limite nordique; PN, population naturelles; FCF, fragmentation du couvert forestier; CF, coupe forestière et A, autres types de peuplements). P, nombre de populations total; I, nombre d'individus total;  $H_0$ , moyenne de l'hétérozygotie observée;  $H_E$ , moyenne de l'hétérozygotie attendue;  $F_{IS}$ , coefficient de consanguinité

n°	М	Cohorte échantillonnée	Р	Ι	ΓN	NA	FCF	CF	Y	H <sub>0</sub>	$H_{\rm E}$	$F_{\rm IS}$	Référence
1	Az	Arbre	5	500	×	×				—	0,110	—	(Perry & Knowles, 1989)
2	Az	Arbre	3	302	×	×						—	(Perry & Knowles, 1991)
3	Az	Juvénile ( $\leq 1$ cm diamètre à la base) et canopée ( $\geq 30$ cm dhp)	15	755			×			0,169	0,171	_	(Foré et al., 1992b)
4	Az	Embryon, semis d'1 an, jeune arbre (dhp $\leq 2$ cm), sous-canopée (dhp = 15–25 cm) et canopée (dhp $\geq 40$ cm)	1	280		×				0,15	0,148	_	(Foré et al., 1992a)
5	Az	Utilisation de l'index de déclin pour chaque arbre	6	350					$\times$	0,141		-0,03	(Geburek & Knowles, 1992)
6	Az	Arbre	6	350		×				_		-0,027	(Geburek, 1993)
7	Az	Plantule d'1 an	16	384	×	×				—	0,112	0,042	(Young et al., 1993b)
8	Az	Plantule d'1 an	16	800			×			—	0,115	0,062	(Young et al., 1993a)
9	Az	Plantule d'1 an	8	968			×			—	—		(Young & Merriam, 1994)
10	Az	Embryon, juvénile (1 an à $\leq$ 1,0 cm de diamètre à la base) et canopée (dhp $\geq$ 30 cm)	4	572			×			0,208		0,126	(Ballal, 1994)
11	Az	Juvénile (6–15 ans) et canopée (arbres établi avant 1850)	1	152					×	0,136	0,148	0,077	(Simon et al., 1995)
12	R	Plantules issues d'une récolte de graines	6	547		×				—		—	(Gunter et al., 2000)
13	R	Plantules	2	49	×	×				_	_		(Diochon <i>et al.</i> , 2003)
14	Az	Plantules (dhp $\leq$ 1 cm), jeunes arbres (dhp $>$ 1–10 cm) et adultes (dhp $>$ 10 cm)	16	567				×		0,113	0,116	0,025	(Baucom <i>et al.</i> , 2005)



**Figure 1.8** Caractéristiques des populations situées au front de migration ("leading edge") et à la zone arrière ("rear edge") de l'aire de répartition des espèces. La largeur des barres grises, à gauche de la figure, indique la variation de chacune des caractéristiques mesurées correspondant à la position des populations dans l'aire de répartition. Figure tirée de Hampe & Petit (2005).

La différenciation génétique entre les populations d'une même région est faible (Young et al., 1993b; Gunter et al., 2000), alors qu'elle est plus importante entre deux populations à la limite sud de l'aire de répartition comparativement à celles situées au centre et au nord de l'aire de répartition (Gunter et al., 2000). En bref, la grande partie de la diversité génétique se retrouverait intra-populations (Foré et al., 1992a; Young et al., 1993b). Les études génétiques sur l'érable à sucre à l'échelle du peuplement ont été effectuées à plusieurs niveaux suivant les familles (graines issues d'un même arbre), l'individu, la topographie (face sud, nord et haut de pente) et selon différents critères (indice de déclin, cohortes de classes d'âge et dhp). La distance génétique est plus grande entre les familles qu'entre les peuplements ou les régions (Gunter et al., 2000). Selon la position de chaque individu dans le peuplement, une structure très faible a été trouvée pour des populations proches de la limite nordique (Perry & Knowles, 1991) et d'autres populations situées un peu plus au sud (Geburek, 1993). Du point de vue topographique, la division en sous-groupe selon la position des individus sur la pente (face sud, nord et haut de pente) n'a pas montré de motif d'association génétique particulier en utilisant la RAPD (Diochon et al., 2003). En comparant des sous-groupes d'individus en santé et en déclin, aucune différence génétique (allozymes) n'a été détectée (Geburek & Knowles, 1992). Des analyses de la structure allélique de sous-groupes classés suivant des classes d'âge et de dhp n'ont révélé qu'une faible différenciation entre cinq cohortes (Foré *et al.*, 1992a). Une différence (fréquences génotypiques et alléliques) a été observée entre les semis d'un an et les autres cohortes (Tableau 1.1).

### 1.2.8. Génétique et impacts anthropiques

Plusieurs études se sont intéressées à l'effet, sur la diversité et structure génétique, de la fragmentation du couvert forestier par l'agriculture, l'urbanisation et la récolte du bois qui a ouvert de grands espaces particulièrement depuis les 200 dernières années en Amérique du Nord. L'érable à sucre est utilisé comme bois de chauffage (Godman *et al.*, 1990) et pour la fabrication des planchers, des feuilles de placages et du contreplaqué (Farrar, 1996; Bowyer *et al.*, 2005). Il est très recherché pour la fabrication de meubles haut de gamme, et compte parmi les espèces utilisées pour les dormants de chemin de fer (Bowyer *et al.*, 2005). Une utilisation de l'érable à sucre qui lui confère une valeur économique très importante est la fabrication du sirop d'érable à partir de l'eau d'érable. À cette fin, la culture de cette espèce (l'acériculture) utilise l'éclaircie pour permettre aux érables à sucre du d'érable (Robitaille & Pardé, 1979). Ces différentes utilisations de l'érable à sucre ouvrent le couvert forestier plus ou moins intensivement, elles ont donc été spécifiquement étudiées pour leur impact sur la génétique de cette espèce.

À la suite de l'ouverture du couvert forestier par l'agriculture, une différenciation génétique moins importante de l'érable à sucre se trouvant dans des massifs forestiers résiduels a été observée (Young & Merriam, 1994), de même qu'une augmentation de l'hétérozygotie et du polymorphisme alléliques (Young *et al.*, 1993a) comparativement à ceux se trouvant dans une matrice forestière relativement

continue. Ces résultats s'expliqueraient par l'absence de barrières physiques à la dispersion du pollen et des graines entre des massifs forestiers dans un paysage agricole et l'augmentation conséquente du flux génique entre des populations fragmentées (Saunders et al., 1991; Kramer et al., 2008). Cependant, une perte d'allèles présents à une faible fréquence dans les populations témoins a été détectée dans les peuplements résiduels (Young et al., 1993a). Ces deux études ont été réalisées sur des plantules âgées d'un an. Toutefois, les études utilisant des cohortes d'érable à sucre de différentes classes de diamètres et d'âge ont montré qu'on pouvait observer, occasionnellement, des différences génétiques. Par exemple, la diversité génétique n'était pas différente selon la cohorte analysée dans l'étude de Foré et al. (1992b), mais elle était plus faible pour les embryons selon Ballal (1994). La différenciation génétique entre les peuplements fragmentés pour les érables à sucre matures (dhp  $\geq$  30 cm) était plus importante que pour les juvéniles ( $\leq$  1 cm de diamètre à la base; établis après la fragmentation) (Foré et al., 1992b) (Tableau 1.1). Il faut toutefois noter que ces deux dernières études ont été réalisées dans des forêts fragmentées et aucune comparaison n'avait été faite avec une forêt continue proche des peuplements fragmentés. Cette différence entre cohortes pourrait donc être dû à un processus naturel et non l'effet de la fragmentation.

Baucom *et al.* (2005) ont analysé les effets de la coupe forestière (réalisée en 1934) de peuplements d'érable à sucre et non la fragmentation du couvert forestier. Ils ont montré, avec les allozymes, que les peuplements coupés avaient une perte de 40 % de diversité génétique (pourcentage de sites polymorphes) par rapport à une forêt ancienne (non perturbée). Cette perte était moindre pour les classes d'âges des jeunes arbres et arbres matures de ces populations (Tableau 1.1). La différenciation génétique entre les peuplements était faible et elle était plus élevée entre les peuplements coupés ( $\theta_p = 0,093$ ;  $\theta_p$  est similaire au  $F_{ST}$ ) comparativement à celle des forêts anciennes ( $\theta_p = 0,060$ ).

Un autre impact de l'activité humaine moins visible à l'échelle du paysage est le problème des pluies acides apparu au cours des années 80. Simon *et al.* (1995) ont étudié l'effet de cette perturbation sur une forêt ancienne d'érable à sucre. L'analyse des isoenzymes a révélé la présence d'une faible différenciation génétique entre le sous-groupe des arbres juvéniles (6–15 ans) et celui des arbres pré-perturbation (établis avant 1850; Tableau 1.1). Certains allèles peu fréquents dans le sous-groupe pré-perturbation étaient encore moins fréquents, voire absents, dans le sous-groupe juvénile. La différence étant peu importante et il en ressort que les pluies acides ne semblent pas avoir eu d'effet sur la structure génétique de l'érable à sucre.

# 1.2.9. Migration post-glaciaire et phylogéographie

Actuellement, l'aire de répartition de l'érable à sucre s'étend du Tennessee, au sud jusqu'au Québec, au nord (section 1.2.2). Jackson et al. (2000) ont montré, par l'étude du pollen fossile, que l'érable (genre Acer incluant l'érable à sucre) était localisé dans deux refuges aux USA (refuge 1: environ les états du Mississipi, d'Alabama et du Tennessee; refuge 2: Caroline du Sud et du Nord) pendant la période du maximum glaciaire, soit vers 21 000 ±1 500 ans AA (Fig. 1.9). Les analyses palynologiques des macrorestes d'érable ont permis de confirmer la limite sud du plus grand refuge, le refuge 1 (Fig. 1.9). Une hypothèse suggère que l'expansion de l'érable à sucre vers le nord en suivant le retrait des glaces a été faite suivant deux voies principales (Braun, 1950) soit une direction: (1) nord-est le long des Appalaches et (2) directement vers le nord. L'expansion directement au nord, aurait été retardée en Ohio et en Indiana (USA) par un climat post-glaciaire chaud et sec. L'érable à sucre se trouvant dans la région des Grands Lacs pourrait être issu d'une migration directement vers le nord ou en provenance de l'est (Braun, 1950). L'érable à sucre s'est établi au sud du Québec vers 8 500 ans AA et des analyses polliniques indiquent que dès cette période, des populations disjointes étaient présentes près du lac St-Jean (Richard & Grondin, 2009). D'autres populations disjointes seraient

également apparues au nord de la Gaspésie vers 4 000 ans AA (Labelle & Richard, 1984). L'érable à sucre, comme d'autres espèces (section 1.1.1), aurait subi un recul vers 4 500–3 000 AA des populations les plus nordiques laissant sur place quelques populations qui sont actuellement à proximité de la limite nordique de l'aire de répartition de l'espèce (dans la sapinière à bouleau jaune). Cette possibilité n'a pas encore été validée pour l'érable à sucre (Richard & Grondin, 2009) (preuves trouvées pour le sapin baumier; Ali *et al.*, 2008), laissant toujours plausible l'hypothèse de « sauts de puce » importants (Cain *et al.*, 2000).



**Figure 1.9** Carte Isopollinique et de macrorestes de l'érable durant le maximum glaciaire (21 000  $\pm 1500$  ans AA). Cercles pleins = présence; vides = absence. Figure modifiée de Jackson *et al.* (2000).

La phylogéographie étudie les liens de parenté, entre des populations d'une même espèce ou des espèces très proches, intégrés dans une dimension spatiale (Avise *et al.*, 1987). Elle permet donc de reconstruire les voies empruntées par les espèces lors de leur migration. Une étude de l'ADN chloroplastique d'une autre espèce d'érable, l'érable rouge, confirme l'hypothèse de deux refuges pour l'érable rouge lors du maximum glaciaire (McLachlan *et al.*, 2005). Le premier refuge serait situé entre les états de Louisiane et Georgie et l'autre, est au sud des Appalaches (McLachlan *et al.*, 2005). Ce sont à peu près les mêmes refuges retrouvés dans l'étude de Jackson *et al.* (2000). De plus, l'analyse de ces haplotypes démontrent que les individus du refuge au sud-ouest (premier refuge) ont moins contribué à la diffusion de l'érable rouge dans les régions plus au nord contrairement au refuge situé au sudest (Fig. 1.10). Dans leur revue de littérature, Soltis *et al.* (2006) ont comparé la phylogéographie d'espèces animales et végétales à l'est de l'Amérique du Nord. Les auteurs ont trouvé que la migration ne répondait pas à un patron simple pour tous les organismes. Cependant, deux espèces ayant à peu près la même répartition nordique pourraient avoir rencontré les mêmes barrières géographiques lors de leur migration, comme les Appalaches ou les Grands Lacs dans l'est de l'Amérique du Nord. Dans la perspective où deux voies de migration existeraiennt pour l'érable à sucre et rouge, nous pourrions émettre l'hypothèse que celles-ci se sont rencontrées au nord au Québec ou en Ontario.



**Figure 1.10** Répartition actuelle (pointillé rouge) et des haplotypes d'ADN chloroplastique de l'érable rouge (schéma des clades au centre). La ligne bleue représente le maximum glaciaire. La ligne noire représente en haut à gauche la répartition de l'espèce (pollen) il y a 15 000 ans et en haut à droite il y a 16 000 ans. Figure tirée de McLachlan *et al.* (2005).

#### 1.2.10. Dépérissement, contraction ou expansion?

Au Québec, on observe un certain dépérissement des érablières depuis le début des années 80. Les symptômes sont une perte de coloration des feuilles, une diminution de la vigueur, et de la croissance, une perte prématurée du feuillage et une diminution de la capacité à cicatriser les blessures causées par les entailles pratiquées pour l'acériculture. Différentes études indiquent que ce dépérissement pourrait être dû (1) à la sécheresse et aux insectes défoliateurs (Payette *et al.*, 1996), (2) à la sécheresse et à des déficiences en potassium (K), phosphate (P) et magnésium (Mg) (Bernier *et al.*, 1989) ou (3) encore, plus probablement, à une déficience en nutriments et acidification du sol (Moore *et al.*, 2000; Ouimet *et al.*, 2001; Duchesne *et al.*, 2005).

Dans la région de la ville de Québec, durant le refroidissement du petit âge glaciaire (environ quatre siècles) l'abondance de l'érable à sucre a fortement diminué au profit de l'épinette rouge qui a atteint son maximum au 18<sup>ième</sup> siècle (Richard & Grondin, 2009). Lors du réchauffement qui a suivi, l'érable à sucre a rapidement repris la même importance qu'avant le petit âge glaciaire. Des travaux réalisés dans la région du Bas St-Laurent à l'est du Québec ont montré que l'érable à sucre et le peuplier faux-tremble (*Populus tremuloides*) étaient moins abondants que les résineux (cèdre blanc de l'est, *Thuja occidentalis* et de l'épinette, *Picea* spp.) avant la colonisation par les Européens (1846–1949) comparativement à l'actuel (1980–2009) (Dupuis *et al.*, 2011). Les auteurs attribuent cet enfeuillement à l'augmentation des coupes forestières (coupes à blanc) et des feux après 1900 (Dupuis *et al.*, 2011). Plus au sud, dans les forêts dominées par le chêne (*Quercus*) et le caryer (*Carya*), on observe également une augmentation de l'érable à sucre et de l'érable rouge qui serait due au contrôle des feux depuis le début des années 20 (Abrams, 1992; Nowacki & Abrams, 2008; Pierce *et al.*, 2008). Par exemple, dans une réserve biologique de

l'Indiana (USA), la valeur d'importance de l'érable à sucre est passée du 12<sup>ième</sup> rang en 1960 (6,3 tiges/ha) au 2<sup>ième</sup> rang en 2000 (68,9 tiges/ha) (Pierce et al., 2008).

Dans le futur et dans le contexte des changements climatiques, la plupart des modèles prédisent une migration vers le nord pour l'érable à sucre, comme pour la plupart des espèces, qui serait accompagnée d'une diminution de leur abondance au sud (Goldblum & Rigg, 2005; Iverson *et al.*, 2008).

#### 1.3. Hypothèses, Objectifs et structure de la thèse

Nos travaux s'articulent autour de l'hypothèse que les populations nordiques, fragmentées sont les plus susceptibles d'être à l'origine de l'expansion, vers le nord, de l'aire de répartition de l'érable à sucre dans un contexte de changements climatiques.

Dans cette thèse, nous avons étudié les populations nordiques, situées dans l'aire continue (non fragmentées) et discontinue (fragmentées) de répartition de l'érable à sucre au Québec. Nous avons abordé à la fois les aspects démographiques et génétiques de la dynamique des populations, à l'aide d'une approche multidisciplinaire.

Pour se propager vers le nord, les populations nordiques fragmentées doivent déjà, actuellement, pouvoir se régénérer. Dans le chapitre II, nous avons abordé l'aspect écologique en étudiant la capacité reproductrice des populations nordiques d'érable à sucre dans trois régions (ouest, centre et est) du Québec. Nous avons comparé leur capacité reproductrice à celle des populations qui se situent plus au sud. Pour plus de clarté, nous avons délimité, au Québec, une zone nordique où les populations d'érable à sucre sont fragmentées naturellement, cette zone a été appelée « zone discontinue » et les populations sont ainsi appelées « populations discontinues ». Cette zone discontinue se situe dans les domaines bioclimatiques de la sapinière à bouleau jaune et de la sapinière à bouleau blanc (Fig. 1.2b). Les populations plus au sud sont incluses dans une zone appelée « zone continue ». Cette zone est comprise exclusivement dans le domaine bioclimatique de l'érablière à bouleau jaune, pour être proche écologiquement de la zone discontinue (Fig. 1.2b). De plus, nous avons comparé les régions entre elles pour savoir si des variations existaient longitudinalement au Québec.

Une population ayant une diversité importante d'un point de vue génétique, aura plus de chance d'avoir, parmi ses individus, des individus adaptés à différentes conditions environnementales. Dans l'hypothèse que ces populations nordiques puissent se propager, il faut aussi qu'elles aient une grande diversité génétique, afin de pouvoir prendre de l'expansion vers des nouveaux milieux. Dans le chapitre IV de cette thèse, nous avons étudié, pour les mêmes populations que celles du chapitre II, le niveau de diversité génétique et le type de structure génétique. Nous avons fait les mêmes comparaisons entre les populations discontinues (nordiques, fragmentées) et continues (sud), et entre les régions.

Dans l'éventualité où les populations nordiques auraient une très faible capacité reproductrice et très peu de diversité génétique, l'expansion nordique pourrait difficilement provenir de ces populations et devrait provenir des populations situées plus au sud. Dans la zone continue (au sud), l'érable à sucre est exploité pour son bois. Cette perturbation d'origine anthropique a-t-elle un effet négatif sur la diversité génétique des populations et sur leur potentiel d'adaptation et d'expansion? Nous avons abordé cette question dans le chapitre V.

Deux points importants sont à mentionner et constituent l'aspect novateur de cette thèse :

L'un des aspects concerne les chapitres III à V; Les études en génétique des populations utilisent de plus de plus un type de marqueur très polymorphe: les marqueurs microsatellites (Li *et al.*, 2002). De tels marqueurs n'étant pas développés

pour l'érable à sucre, nous avons d'abord utilisé des microsatellites développés pour d'autres espèces d'érables. Ce transfert ne s'est pas avéré très fructueux et nous avons utilisé la technique du pyrosequençage pour le développement de marqueurs microsatellites spécifiques à l'érable à sucre. Un article a été publié à partir de ses travaux et fait l'objet du chapitre III.

Un second aspect est le nombre de populations étudiées couvrant une large zone géographique et le fait qu'elles soient étudiées sous les aspects démographique et génétique (chapitres II et IV). Comme nous l'avons vu (Fig. 1.7), les études génétiques sur les populations proches de la limite nordique et sur leur potentiel de régénération sont limitées tant du point de vue du nombre de populations étudiées que de leur répartition. La présente étude couvre trois régions au Québec où les populations d'érable à sucre peuvent prendre de l'expansion au nord.

### CHAPITRE II

# GEOGRAPHICAL VARIATION IN REPRODUCTIVE CAPACITY OF

# SUGAR MAPLE (*ACER SACCHARUM* MARSHALL)

# NORTHERN PERIPHERAL POPULATIONS

Noémie Graignic, Francine Tremblay & Yves Bergeron

Article publié en 2014 dans Journal of Biogeography, vol. 41, p. 145–157

## 2.1. Résumé

**Objectif** • L'érable à sucre (*Acer saccharum* Marshall) atteint sa limite de répartition continue nordique dans le nord-est de l'Amérique du Nord au niveau de la transition entre la forêt boréale mixte et la forêt feuillue tempérée. Avec les changements climatiques, plusieurs modèles prédisent la migration vers le nord de cette espèce. Notre objectif était d'évaluer si le potentiel de recrutement de l'érable à sucre diminuait au nord suivant la transition progressive entre les zones continue et discontinue de l'aire de répartition de l'espèce.

Localisation • Le nord-est du Canada, à la limite nordique de l'érable à sucre.

**Méthodes** • Nous avons analysé la capacité reproductrice des populations d'érable à sucre le long de trois transects latitudinaux (situés dans l'ouest, le centre et l'est du Québec) dans 24 sites situés entre  $45^{\circ}51' - 48^{\circ}59'$  N et  $70^{\circ}21' - 79^{\circ}27'$  O. Le territoire étudié a été divisé en deux zones, continue et discontinue, sur la base de l'abondance des peuplements d'érable à sucre dans chacune des zones. Nous avons examiné la structure d'âge du peuplement, la production et la germination des graines d'érable à sucre, la densité et la structure d'âge des plantules d'érable à sucre.

**Résultats** • La régénération de l'érable à sucre était inéquienne et similaire entre la zone continue (sud) et discontinue (nord). Nous avons observé plus de samares pleines et de plantules dans la zone continue par rapport à la zone discontinue sur le transect ouest. La densité des plantules d'érable à sucre était positivement influencée par (1) la surface terrière des érables à sucre matures et celle des jeunes arbres de toutes les espèces, et (2) la température et les précipitations moyennes de juillet. Quatre années semencières ont été identifiées et étaient bien synchronisées dans l'ensemble des sites; les bonnes années semencières variaient significativement avec les températures moyennes et les précipitations moyennes de juillet de l'année précédente. **Conclusions principales** • Notre étude a clairement démontré un effet des variables climatiques et des caractéristiques du peuplement sur la régénération de l'érable à sucre. Cependant, ces facteurs n'expliquaient pas le passage de la zone de répartition continue à discontinue pour cette espèce. La plupart de nos sites nordiques montraient un recrutement constant de l'érable à sucre dans le temps. Ces résultats soulignent l'importance d'inclure des facteurs non climatiques dans les modèles visant à prédire des changements dans l'abondance des espèces.

#### 2.2. Abstract

**Aim** • Several models have predicted that, with climate change, Northern Hemisphere species will migrate northwards from their present distribution ranges. Sugar maple (*Acer saccharum* Marshall) reaches its northern continuous distributional limit in north-eastern North America at the transition between boreal mixed-wood and temperate deciduous forest. Our objective was to determine whether lower sugar maple recruitment potential accounts for this gradual transition between the continuous and discontinuous zones of the distribution.

Location • The northern limit of sugar maple in eastern Canada.

**Methods** • We analysed the reproductive capacity of sugar maple populations along three latitudinal transects (located in the west, centre and east of Québec) in 24 sites located between 45°51′–48°59′ N and 70°21′–79°27′ W. The study area was divided into two zones, continuous and discontinuous, based on sugar maple stand abundance. We examined stand structure, sugar maple seed abundance and germination, and sugar maple seedling density and age structure.

**Results** • Sugar maple regeneration was uneven-aged and similar between continuous (south) and discontinuous (north) zones. For the western transect, more filled seeds and more seedlings were recorded in the continuous zone than in the discontinuous zone. Sugar maple seedling density was positively influenced by (1) basal area of mature sugar maple and saplings of all species, and (2) July mean temperature and precipitation. Four mast seed years were identified that were well synchronized across all sites; mast seeding covaried significantly with July mean temperature and July mean precipitation of the previous year.

Main conclusions • Our study clearly demonstrated an effect of climatic variables and stand characteristics on sugar maple regeneration. However, these

factors did not explain the transition from a continuous to a discontinuous distribution for this species. Most of our northern sites exhibited constant sugar maple recruitment over time. These results highlight the importance of including non-climatic factors in models predicting species change in abundance.

#### 2.3. Introduction

Several models suggest that tree populations are capable of rapid migration in the face of climatic variation (Iverson et al., 2008; Morin & Thuiller, 2009). Accordingly, the reconstruction of post-glacial vegetation in North America shows several forest tree species reaching their maximum extent in the middle Holocene, followed by contractions that could have been caused by a cooling climate (McLachlan et al., 2005). Given ongoing climate changes, many species are expected to expand their ranges, which could be particularly marked for populations located at the limits of their distribution ranges (Iverson et al., 2004). For example, the northern limits of many tree species are modulated primarily by climatic factors, mainly temperature and light availability, which constrain the chances of survival at various stages of the life cycle (Woodward, 1987). A progressive decline in reproductive success results in a reduced capacity of these species to sustain themselves and to colonize new sites (Gaston, 2009). Among various natural processes, disturbances (e.g. fire, insect outbreaks and disease) play a major role in regulating species distributions at their latitudinal limits by altering population densities and limiting reproduction capacity (Ali et al., 2008). Human activities, such as agriculture and forest exploitation, are superimposed upon these factors and can contribute to habitat fragmentation, thereby accentuating the isolation of populations at a landscape scale (Vranckx *et al.*, 2012).

To date, only a few studies have specifically addressed climate effects on species that are present within transition zones between forested areas. Most of these studies have focused on species that are present in northern tree-line ecosystems (i.e. the transition zone between boreal forest and the tundra) that are very sensitive to changes in climatic conditions, and which have shown recent species expansions (Lloyd & Fastie, 2003; Caccianiga & Payette, 2006). However, recent empirical observations in North America have shown that range contraction can also be observed (Zhu *et al.*, 2012).

Sugar maple (*Acer saccharum* Marshall) is a widespread and abundant tree in north-eastern North America that reaches its northern continuous distribution range at the transition between boreal mixed-wood and temperate deciduous forests (Saucier *et al.*, 2003). It is a deciduous, shade-tolerant species (Logan, 1965) that forms uneven-aged stands (Majcen *et al.*, 1984) and has major ecological and economic value in eastern North America (Godman *et al.*, 1990). Like other tree species in the Northern Hemisphere, sugar maple is predicted to migrate northwards from its current range. In the United States, models predict decreases in abundance at the southern edge of this species' range (Iverson *et al.*, 2008) and a northward expansion that will eventually lead to an increase in sugar maple abundance towards its northern limits in Canada (Goldblum & Rigg, 2005).

Climate controls species distribution in part by affecting recruitment at different phases of sexual reproduction (Walck *et al.*, 2011). Sugar maple seeds require high soil moisture levels during germination (Janerette, 1979) and a period of stratification at low temperatures, between 1 and 5 °C, to break embryo dormancy and stimulate germination (Shih *et al.*, 1985; Godman *et al.*, 1990). Germination of northern seed sources begins one week earlier at 1 °C than at 7 °C, but the cumulative proportion of germination after 90 days is 20% higher at 7 °C than at 1 °C (McCarragher *et al.*, 2011). Sugar maple seedlings and mature trees may be affected by early leaf senescence and eventually die due to frost damage (Pilon *et al.*, 1994). Sufficient understorey light is limiting during the growing season for seedlings at northern latitudes and may lead to decreased seedling growth because the time period between seedling leaf emergence and canopy closure by mature trees is shorter in northern stands (Kwit *et al.*, 2010).

In the present study, we analysed the reproductive capacity of sugar maple populations along a climatic gradient. Our objective was to determine whether or not sugar maple recruitment differs between the discontinuous and continuous zones within its northern range. We hypothesized that sugar maple had lower recruitment in the discontinuous than in the continuous part of its range, because seed production and seedling survival are reduced by low temperatures. To test this hypothesis, we examined stand structure, seed abundance and germination, seedling density and seedling age structure in populations along latitudinal transects ranging from the southern limit of the continuous sugar maple distribution in the sugar maple–yellow birch bioclimatic domain to its northern limit in the balsam fir–white birch bioclimatic domain.

#### 2.4. Materials and methods

#### 2.4.1. Study area

The study was located at the northern range limit of sugar maple in Québec, eastern Canada (Fig. 2.1). The presence of sugar maple stands along the latitudinal gradient was estimated from the analysis of large inventory databases ('points d'observations écologiques', 'placettes permanentes' and 'placettes temporaires') of the Ministère des Ressources Naturelles du Québec (MRNQ, 2013a). The gradient was divided into two zones based on the proportion of sugar maple stands in the continuous and discontinuous zones (8.5% and 2.2%, respectively).



**Figure 2.1** Map of the study area at the northern limit of sugar maple (*Acer saccharum*) distribution in Québec, showing the three transects (lines), locations of the 24 study sites (polygons; see Table A2.1 for more details), sugar maple–yellow birch (*Betula alleghaniensis*) bioclimatic domain (hatched), and boundary of all bioclimatic domain limits (thin line) (Saucier *et al.*, 2003). In the top middle map panel, points represents sugar maple detected in databases of the Ministère des Ressources Naturelles du Québec ('points d'observation ecologiques', 'placettes permanentes' and 'placettes temporaires').

The continuous zone lay within the sugar maple–yellow birch (*A. saccharum* and *Betula alleghaniensis* Britton) bioclimatic domain where sugar maple is abundant. The discontinuous zone was in the balsam fir–yellow birch (*Abies balsamea* (L.) Miller and *B. alleghaniensis*) bioclimatic domain, and some northern sites were located in the transition with the balsam fir–white birch (*A. balsamea* and *Betula papyrifera* Marshall) bioclimatic domain (Saucier *et al.*, 2003). Old-growth, unevenaged sugar maple stands were selected to be as similar as possible after an analysis of the databases (the ecoforestry maps and tessellation-based forest information system

(Système d'Information FORestière par Tesselle, SIFORT) obtained from MRNQ (2013a,c). The ecoforestry maps are a result of three aerial photographic campaigns and field surveys. SIFORT uses a 14-ha grid overlaid on the ecoforestry maps. In the SIFORT database, sites that had been severely disturbed by fires, insect outbreaks or wind-throw since 1965 were identified (MRNQ, 2013c). In all sites, the diameter structures of sugar maple trees formed a reverse-J shape that is typical of unevenaged stands (Fig. A2.1).

Sites were distributed along three north–south transects (eight sites per transect) and four sites per transect per zone (Fig. 2.1). Eleven of those sites are old-growth or rare forests that were classified as Exceptional Forest Ecosystems (EFE; MRNQ, 2013b). Data were collected at 24 sites between 45°51' N and 48°59' N latitude, and between 70°21' W and 79°27' W longitude (Table A2.1), at elevations ranging between 157 and 493 m a.s.l. Climatic data for each zone and transect (1971–2010) are summarized in Table A2.2 and were estimated from the meteorological station nearest to each study site (Environment Canada, 2009).

### 2.4.2. Field sampling

Data were collected in 2008 and 2009. Quadrats of 0.16 ha (40 m × 40 m) were established at each site. Within each quadrat, the diameter at breast height (d.b.h. at 1.3 m) of every mature tree (d.b.h.  $\geq$  10 cm) was measured. The five largest (d.b.h.) sugar maple trees were cored in each study plot. Cores were mounted on wooden strips and sanded, and the number of rings was counted to determine their age. The ages of individuals with missing piths were estimated using correlations with trees of the same size of the missing portion of the cores. Data on age and size of trees from the same areas were available in Majcen *et al.* (1984).

The d.b.h. of every sapling (1 cm  $\leq$  d.b.h. < 10 cm) was measured within five circular plots (2 m radius) that were randomly distributed in the larger study plot. The

age of sugar maple seedlings (d.b.h. < 1 cm) was recorded in 15 circular plots of 1 m<sup>2</sup> that were randomly distributed across the study site. Seedlings were excavated to count all scars because sugar maple frequently develops adventitious roots along prostrate portions of stems. Age was estimated for 8822 sugar maple seedlings (total across all sites) using scars that were left each year by the loss of the scales protecting the bud of the terminal leader. Dating may be not precise for older seedlings, but this imprecision should be constant across sites. Five-year age classes were used in subsequent age structure models. When it was not possible to count the number of scars (i.e. due to stem breakage or larger basal diameter), the affected individuals were used to estimate seedling density ('undetermined' class) but were excluded from the age structure analysis.

Three seed traps, which covered a total area of 3 m<sup>2</sup>, were randomly distributed within the 0.16-ha sample plot in two sites per zone in the western transect (transect 1 – Lac Labelle, Rémigny, Kipawa and Lac Six Milles; Table A2.1). Traps were installed in July and were removed at the end of October in 2008, 2009, 2010 and 2011. Filled seeds (seeds with kernels) were checked by firmly pressing each seed between the fingers during counting. Unfilled seeds were discarded and the remaining filled seeds were cold moist-stratified at 3 °C (for a maximum of 42 days) and germinated. Germination tests were performed at the Centre de Semences Forestières de Berthier of MRNQ.

#### 2.4.3. Data analysis

Data were analysed using the statistical software package R version 2.13.1 (R Development Core Team, 2011).

### 2.4.3.1. Forest composition

Tree species composition (presence/absence) in the 24 sites was analysed using correspondence analysis (CA, using the *cca* function in the VEGAN library; Oksanen

*et al.*, 2011). Mature sugar maple trees and sugar maple saplings were excluded from this analysis because they were present in all sites.

#### 2.4.3.2. Recruitment between zones

Variation in mature tree density and basal area was tested for every tree species (across all plots) and for sugar maple alone, using two-way analysis of variance (ANOVA) with zone, transect and their interaction as fixed effects. Density and basal area for all tree species (total) and sugar maple saplings, seed abundance and sugar maple seedling density were tested using a linear mixed-model analysis (LMM, using the *lme* function in the NLME library; Pinheiro *et al.*, 2011). For saplings, zone, transect and their interaction were treated as fixed effects in the analysis. For seed abundance, zone, year and their interaction were the fixed effects. For sugar maple seedling density responses, the fixed effects were seedling emergence year, zone, transect and their three-way interaction (seedling emergence year  $\times$  zone  $\times$  transect). Random effects included site for saplings and seed abundance, and site and circular plot for sugar maple seedling density. Year of seedling emergence was estimated on the basis of seedling age for the period from 1970 to 2008, plus one 'undetermined' class. The assumption of normality and homoscedasticity was verified graphically; seed abundance and seedling density data were log-transformed to respect these assumptions. ANOVA and linear mixed models were simplified by stepwise backward elimination of non-significant fixed-effect terms to produce the most parsimonious models.

## 2.4.3.3. Filled seeds and germination

We used generalized linear mixed-effects models (GLMM, using the *glmer* function in the LME4 library; Bates et al., 2011) with a logit link function and binomial error distribution to analyse the filled seed ratio (the proportion of filled seeds to total seeds) and the germination ratio (the proportion of germinated seeds to

filled seeds). In order to avoid problems in estimation, year and zone were considered to be fixed effects in two different models for each seed ratio analysis, and in one model for germination ratio; seed trap was considered a random effect. The germination ratio model was simplified by backward elimination.

#### 2.4.3.4. Age structure

To compare sugar maple recruitment across the study area, seedlings (1–40 years old) were grouped into 5-year age classes. Sugar maple seedling age structures can be described either by a negative exponential function with constant mortality, or by a power function with a higher mortality rate in early years (Hett & Loucks, 1971). We used the Akaike information criterion (AIC; *aictab* in the AICCMODAVG library; Mazerolle, 2011), corrected for small sample sizes (AIC<sub>c</sub>), and AIC<sub>c</sub> weights ( $\omega$ ) for model evaluation (Burnham & Anderson, 2004). The power function was the best model ( $\Delta$ AIC<sub>c</sub>  $\leq$  4) for most of the 24 sites (Table A2.3). It was used to estimate *R* (initial recruitment) and *M* (mortality rate) with the function ln(stem m<sup>-2</sup> + 1) = ln (*R* + 1) – *M* × ln(age). We then used *R* and *M* to assess the effects of zones on sugar maple abundance.

The age-structure data set (n = 24) was not large enough to include all stand and climate variables in one model. Sugar maple seedling ln(R) was analysed using oneway, two-way or three-way ANOVA. We determined five candidate models: model 1 considered the effect of mature sugar maple basal area; model 2, mature sugar maple density; model 3, transect, zone and transect × zone; model 4 included models 1 and 3; and model 5 included models 2 and 3. We ranked each model based on AIC<sub>c</sub>. We then computed  $\Delta$ AIC<sub>c</sub> and  $\omega$  to determine the strength of evidence for each model (Burnham & Anderson, 2004). Following model fitting, we performed multimodal inference (*modavg* function in AICCMODAVG library) when required ( $\Delta$ AIC<sub>c</sub> ≤ 4) to assess variable effects on the initial recruitment of sugar maple. Variation in M for sugar maple seedlings was tested using two-way ANOVA with zone, transect and their interaction as fixed effects. Assumptions of normality and homoscedasticity were verified prior to analysis of  $\ln(R)$  and M.

### 2.4.3.5. Relationships with climate

Sugar maple seedling density was compared using LMM analysis for transect  $\times$ zone, or stand and/or meteorological fixed effects, with site as a random effect. Meteorological data for all sites (monthly means for 1930–2010) were obtained using the simulation software BIOSIM 9 (Régnière, 1996). We constructed different models to test the following effects: transect × zone; stand characteristics, which included total density and basal area (density and basal area included mature tree, mature sugar maple, total saplings and sugar maple saplings); climatic conditions, which included average annual temperature and precipitation, monthly average temperature and precipitation from April to October; and global models with a combination of one model of stand characteristics and one model of climate conditions. Correlations between variables were examined; total sapling density was therefore excluded from the model, because it was strongly correlated with sugar maple sapling density ( $r^2 =$ 0.80). Assumptions of normality and homoscedasticity were tested on all 16 global models and seedling density was log-transformed. Fits of the 27 models were compared using AIC<sub>c</sub>. Model averaging was performed to calculate parameter estimates and unconditional 95% confidence intervals (CI).

Mast years were compared for transect 1 using GLMM. We constructed a model based on the hypothesis that high seed production was related to warm, dry conditions in the spring of the previous year (Houle, 1999).

# 2.5. Results

## 2.5.1. Stand characteristics

The first and second axes of the correspondence analysis (CA) accounted for 35.0% of the variation in tree species composition among sites. The first CA axis explained 22.8% of the variance (eigenvalue  $\lambda_1 = 0.68$ ) and clearly highlighted a difference in tree stand composition between zones (Fig. 2.2). Towards the left-hand end of CA axis 1, most species were found only in the continuous zone (transects 1 and 2), except for mature ironwood (Ostrya virginiana (Mill.) K. Koch), which was also present in one site from the southern discontinuous zone (2-D-D', Lac des Polonais). Towards the right-hand end of the first CA axis, species that were observed in the discontinuous zone are characteristic of the balsam fir-white birch bioclimatic domain (i.e. white or paper birch and trembling aspen, Populus tremuloides Michx.). One exception was white or American elm (Ulmus americana L.), which was also observed in the continuous zone. A second CA axis of variation  $(\lambda_2 = 0.37, 12.2\%$  of the variance) distinguished sites within zones. Variation in species composition was greater within the continuous zone. For example, in continuous zone transect 2, variation was due to one species being present only once in three of the four sites sampled; this included saplings of bitternut hickory (Carya cordiformis (Wangenh.) K. Koch) at site 2-C-D (Lac de l'Ecluse), mature red spruce (Picea rubens Sarg.) at site 2-C-A (Montagne du Diable), and mature white elm at site 2-C-B (Lac Ecuyer).



**Figure 2.2** Ordination diagram of species presence/absence (except sugar maple, *Acer saccharum*) by site for mature trees (M; n = 24) and saplings (S; n = 24) obtained by correspondence analysis (CA) in Québec. (a) By site: Circles, transect 1; squares, transect 2; triangles, transect 3; open symbols, discontinuous zone; black symbols, continuous zone; (b) by species. CA1, correspondence analysis axis 1; CA2, correspondence analysis axis 2. Species label abbreviations: Ab, balsam fir, *Abies balsamea* (L.) Mill.; Ap, striped maple, *Acer pensylvanicum* L.; Ar, red maple, *Acer rubrum* L.; Ba, yellow birch, *Betula alleghaniensis* Britt.; Bc, mountain paper birch, *Betula cordifolia* Regel; Bp, white birch, *Betula papyrifera* Marshall; Cc, bitternut hickory, *Carya cordiformis* (Wangenh.) K. Koch; Fg, American beech, *Fagus grandifolia* Ehrh.; Fn, black ash, *Fraxinus nigra* Marshall; Qr, red oak, *Quercus rubra* L.; Ov, ironwood, *Ostrya virginiana* (Mill.) K. Koch; Pg, white spruce, *Picea glauca* (Moench) Voss; Pr, red spruce, *Picea rubens* Sarg.; Pt, trembling aspen, *Populus tremuloides* Michx.; Ta, basswood, *Tilia americana* L.; To, eastern white-cedar, *Thuja occidentalis* L.; Ua, white elm, *Ulmus americana* L.

Total mature tree density was higher along all transects ( $F_{1,18} = 8.9691$ , P = 0.0078) in the discontinuous zone than in the continuous zone (means of 519 stems ha<sup>-1</sup> and 390 stems ha<sup>-1</sup>, respectively; Fig. 2.3a). Total basal area was not significantly different (zones, transects or interaction zone × transect), except for total sapling basal area (Table A2.4).

Significantly or marginally significant differences were apparent among transects but not between zones (Table A2.4) for total sapling density ( $F_{2,99} = 5.8395$ , P = 0.0040), total sapling basal area ( $F_{2,99} = 2.9039$ , P = 0.0595) and mature sugar maple density ( $F_{1,18} = 3.9380$ , P = 0.0381). Indeed, sapling densities and sapling basal area were significantly higher along transect 1 (Fig. 2.3c, 3209 stems ha<sup>-1</sup>; Fig. 2.3e, 7.02 m<sup>2</sup> ha<sup>-1</sup>) than along transect 3 (Fig. 2.3c, 1691 stems ha<sup>-1</sup>; Fig. 2.3e, 3.29 m<sup>2</sup> ha<sup>-1</sup>). Mature sugar maple density was lower in transect 1 (223 stems ha<sup>-1</sup>) than in transect 2 (356 stems ha<sup>-1</sup>). No differences were observed between transect 3 (305 stems ha<sup>-1</sup>) and transects 1 and 2 (Fig. 2.3b). Sugar maple sapling densities differed among transects, but this response depended upon zone (zone × transect interaction:  $F_{2,99} = 3.1813$ , P = 0.0458; Table A2.4). The highest sugar maple sapling density was encountered in transect 1 of the continuous zone (Fig. 2.3d, 3024 stems ha<sup>-1</sup>), while no differences were found between zones within each transect.



**Figure 2.3** Predicted means ( $\pm$  95% confidence intervals) of (a) total mature tree density (stems ha<sup>-1</sup>; n = 24), (b) mature sugar maple (*Acer saccharum*) density (stems ha<sup>-1</sup>; n = 24), (c) total sapling density (stems ha<sup>-1</sup>; n = 111), (d) sugar maple sapling density (stems ha<sup>-1</sup>; n = 111) and (e) total sapling basal area (m<sup>2</sup> ha<sup>-1</sup>; n = 111) in two zones (C: continuous and D: discontinuous) or three transects (1, 2 and 3) or zones and transects in Québec. Means with the same letter do not differ at  $\alpha = 0.05$  but differ (with an asterisk) at  $\alpha = 0.10$ .

### 2.5.2. Sugar maple seed production

Sugar maple seed production differed between years, depending upon zone (zone × year interaction:  $F_{3,30} = 24.1483$ , P < 0.0001). In 2008, seed production was higher in the continuous zone (mean 34.0 seeds m<sup>-2</sup>) than in the discontinuous zone, where almost no seeds were produced (Fig. 2.4a). In 2009 and 2010, the seed crop was very low in both zones (mean range 0–1.8 seeds m<sup>-2</sup>). In 2011, very high mean seed production was observed within the continuous zone (471 seeds m<sup>-2</sup>) but did not significantly differ from that of the discontinuous zone (211 seeds m<sup>-2</sup>).



**Figure 2.4** Predicted mean ( $\pm$  95% confidence intervals) of (a) sugar maple (*Acer saccharum*) seed production (seeds m<sup>-2</sup>) in C (continuous) and D (discontinuous) zones over four years (2008 to 2011; n = 48); (b) sugar maple filled seed ratio in two zones (n = 5234); (c) sugar maple filled seed ratio over 4 years (n = 5234); (d) sugar maple germination ratio over 3 years (2008, 2010 and 2011; n = 1866). Seed study was conducted only in transect 1 (Abitibi-Témiscamingue, Québec). Means with the same letter do not differ at  $\alpha = 0.05$  but differ (with an asterisk) at  $\alpha = 0.10$ .

The filled seed ratio varied significantly between zones and years (Fig. 2.4b,c). The mean ratio was higher in the continuous (0.42) than in the discontinuous zone (0.26). It was also higher in 2010 and 2011 (0.46 and 0.35, respectively) than in 2008 (0.11) or 2009 (0.00). Significant differences in germination ratios were apparent between years (2008, 0.30; 2011, 0.18) but not between zones (Fig. 2.4d).

# 2.5.3. Natural recruitment of sugar maple seedlings

Sugar maple seedling density varied significantly across years, zones and transects (seedling emergence year × zone × transect interaction:  $F_{78,13761} = 12.23615$ , P < 0.001). The age structure exhibited an inverse J shape, suggesting sustained seedling recruitment, especially along transect 1 (Fig. 2.5). In transect 1, the years 1996, 2002 and 2006 were identified as mast seed years in both the continuous and discontinuous zones (here, we assumed that higher seedling recruitment was synchronized with higher seed production during the previous year; Kelly & Sork, 2002). For each mast seed year in this transect, the continuous zone had higher seedling densities. For transect 2, 2002 and 2003 were mast years, whereas only 2003 was a mast year in transect 3. Mast seed years 1996 and 2006 were apparent in transects 2 and 3, but 1996 was not observed as a mast year in the continuous zone of transect 2.

Using a power function (Fig. 2.6; Fig. A2.2), estimated initial recruitment (*R*) ranged from 4.6 to 119.6 seedlings m<sup>-2</sup> in the discontinuous zone, and from 2.9 to 436.0 seedlings m<sup>-2</sup> in the continuous zone. As with recruitment, the range for mortality rate (*M*) was lower in the discontinuous zone  $[0.47-1.41 \ln(\text{stems m}^{-2} + 1) \ln(\text{years}^{-1})]$ , but it overlapped estimates for the continuous zone  $[0.31-1.70 \ln(\text{stems m}^{-2} + 1) \ln(\text{years}^{-1})]$  (Fig. 2.6; Fig. A2.2). The model that tested the effects of transect and zone upon *R* was not the best model. Models that tested the effects of mature sugar maple basal area (model 1) or mature sugar maple density (model 2) on *R* ranked the highest among our candidate models ( $\Delta \text{AIC}_c \leq 4.0$ ), with  $\omega$  of 0.81 and

0.16, respectively. Nonetheless, multi-model inference showed no influence on *R* of either sugar maple basal area (model-averaged estimate: 0.043; 95% CI: -0.003, 0.089) or density (model-averaged estimate: -0.001; 95% CI: -0.006, 0.004). In brief, none of the explanatory variables that we considered could explain the variation in *R*. *M* was weakly associated with geographical location (zone × transect interaction:  $F_{2,18} = 3.1340$ , P = 0.0679), with a marginally significant difference between transects 1 and 3 in the continuous zone (Fig. A2.3).



**Figure 2.5** Predicted mean ( $\pm$  95% confidence intervals) of sugar maple (*Acer saccharum*) seedling density (stems m<sup>-2</sup>) for each year (Und. (undetermined) and 1970 to 2008) of C (continuous) and D (discontinuous) zones (n = 14119) in Québec. Three transects were established in each zone, e.g. 1D = discontinuous zone in transect 1. \* Potential mast seed year assigned one year before seedling emergence.


**Figure 2.6** Sugar maple (*Acer saccharum*) age structures of the most northern and southern sites of each transect [n = 720; for all 24 sites, see Fig. A2.2 (n = 2880)]. For each population, the linearized form of power function is included, together with its coefficient of determination.  $y_i = r - M \times x_i$ ;  $y_i$  is ln(stems m<sup>-2</sup> + 1); r is the intercept of the curve, ln(R), where R is initial recruitment; M is the slope of the curve, mortality rate and  $x_i$  is ln(age).

# 2.5.4. Relationships with climate

Sugar maple seedling density ranged from 0.0 to 191.8 seedlings m<sup>-2</sup>, with a mean of 25.5 seedlings m<sup>-2</sup> in the 1-m<sup>2</sup> circular plots. Only 6% of the plots lacked sugar maple seedlings (4% in the discontinuous zone; 9% in the continuous zone). According to the AIC<sub>c</sub>, the best model for predicting sugar maple seedling density included basal area, July mean temperature, and July mean precipitation ( $\Delta AIC_c \leq 4.0$ ; Table A2.5). The model including the zone × transect interaction ranked lowest.

Parameter estimates derived from multi-model inference suggested that sugar maple seedling density increased with average July temperature and, to a lesser extent, with sapling basal area, mature sugar maple basal area and mean July precipitation (Fig. 2.7a,b,d,e, Table 2.1). Sugar maple sapling basal area had a negative effect on sugar maple seedling density (Fig. 2.7c, Table 2.1), while total mature tree basal area exerted no influence (Table 2.1).



**Figure 2.7** Predicted sugar maple (*Acer saccharum*) seedling density (stems  $m^{-2}$ ) in response to all explanatory variables in the best supported model (model 21), based on multi-model averaging of all candidate models (n = 353) in Québec. Dashed lines show 95% confidence intervals.

**Table 2.1** Parameter estimates and unconditional confidence intervals from multi-model inference of sugar maple (*Acer saccharum*) seedling density (ln(data + 1)) in Québec. Term abbreviations: m\_BA, mature tree basal area; m\_ers\_BA, mature sugar maple basal area; s\_BA, sapling basal area; s\_ers\_BA, sugar maple sapling basal area; July\_T, July mean temperature; July\_P, July mean precipitation.

Explanatory variables	Model-averaged estimate	95% confidence interval				
Explanatory variables		Lower	Upper			
$m_BA [ln(m^2 ha^{-1} + 1)]$	0.023	-0.008	0.055			
$m_{ers}BA [ln(m^2 ha^{-1} + 1)]$	0.035	0.015	0.055			
$s_BA [ln(m^2 ha^{-1} + 1)]$	0.103	0.037	0.169			
$s_{ers}BA [ln(m^2 ha^{-1} + 1)]$	-0.139	-0.250	-0.028			
July_T $[ln(^{\circ}C + 1)]$	0.578	0.211	0.946			
July_P [ln(mm + 1)]	0.023	0.004	0.041			

The relationship between mast years and climatic variables was tested for transect 1. Mast seeding varied significantly with July mean temperature (Fig. 2.8a) and July mean precipitation (Fig. 2.8b) of the previous year. Predictions were valid and more than 50% probability of masting was observed when July mean temperature was > 19.5 °C and July mean precipitation was < 35 mm.



**Figure 2.8** Predicted probability of mast seeding in sugar maple (*Acer saccharum*) for transect 1 (Abitibi-Témiscamingue, Québec; n = 136) based on (a) July mean temperature of the previous year and (b) July mean precipitation of the previous year. Dashed lines show 95% confidence intervals.

# 2.6. Discussion

Our results indicated that sugar maple regeneration was globally similar across the discontinuous and continuous parts of its distributional range. Seedling age structures was also similar across its range with high recruitment, higher mortality at early stages, and decreasing mortality rates over time in most sampled sites. However, seedling density varied significantly among year (seedling emergence year), transects (west–east), and bioclimatic domains (zones). Variation in sugar maple seedling density was best explained by stand characteristics (mature tree species, mature sugar maple, tree species saplings, and sugar maple sapling basal areas) and climate conditions (July average temperature and precipitation), rather than by zonation.

Sugar maple seedling density was estimated to be 25.5 stems  $m^{-2}$ , when averaged over the 24 stands, and 32.6 stems  $m^{-2}$ , when we considered only the three northernmost sites (1-D-A', Lac Labelle; 2-D-A', Lac Pénobscot; 3-D-A', Lac Patrick). These values were higher than those reported by Caspersen & Saprunoff (2005) for central Ontario (i.e. 11.1 stems m<sup>-2</sup>; Haliburton Forest and Wildlife Reserve, located south of our study area), and the range of values (14.7-29.9 stems m<sup>-2</sup>) that are reported by Goldblum & Rigg (2002, 2009) on the north-eastern shore of Lake Superior (Lake Superior Provincial Park; 47°45' N, 84°42' W). Variation in recruitment was low in our study area, in that sugar maple seedlings were absent from only 9% of circular plots in the continuous zone, which was consistent with a value of 20% for all quadrats measured by Caspersen & Saprunoff (2005). Multi-model inference indicated that five of our six first models included mature sugar maple basal area and that this variable was positively related to sugar maple seedling density (Table 2.1, Fig. 2.7, Table A2.5). This result was also consistent with Caspersen & Saprunoff (2005), who showed that sugar maple seedling density is correlated with higher sugar maple basal area, and with Garrett & Graber (1995), who showed that

larger sugar maple trees generally producing more seeds than did smaller individuals. In contrast, Houle (1992) reported a negative relationship between sugar maple basal area versus sugar maple seed and seedling abundance, within a stand located southeast of our study area. This difference could be related to a different bioclimatic domain (sugar maple–basswood; Houle, 1992) or to stand characteristics that are not representative of a general pattern for sugar maple (Houle, 1992).

Like seedling density, basal area for mature trees (total and sugar maple) did not vary among transects and between zones. Mature sugar maple trees in transect 1 (western region, Abitibi-Témiscamingue) were larger but occurred at lower densities than those in the central region of Québec (transect 2; Fig. 2.3). We had expected to find larger sugar maple trees in the continuous zone where climatic conditions are more favourable to the growth of this species. Our results showed a negative influence of sugar maple sapling basal area on seedling density. The negative effect of saplings may be attributed to the shade produced by sugar maple leaves. This closed shade-cover (compared to mature trees) may induce higher sugar maple seedling mortality, even though sugar maple is a shade-tolerant species (Logan, 1965). This hypothesis of light limitation is supported by Kellman (2004), who found that sugar maple seedling mortality is lower in boreal stands than in sugar maple stands at the same latitude.

In transect 1 (west), differences between the continuous and discontinuous zones were very clear with respect to seed abundance and seedling density. In 2008, seeds were produced only in the continuous zone (34.0 seeds m<sup>-2</sup>; Fig. 2.4). However, this trend was not observed in 2009 and 2010, where very few seeds were produced in either zone, or in 2011, where high production was observed in both zones. Only 2.5 seeds m<sup>-2</sup> germinated in 2008, which was not consistent for a mast seed year, while 43.0 seeds m<sup>-2</sup> germinated in 2011. On the basis of seedling density structure (Fig. 2.5) and seed collections, we identified four mast seed years in transect 1 (1996,

2002, 2006 and 2011). Overall, mast years were well synchronized in all sites, if we excluded 2002–2003. There was a mast year in 2002 in transect 1 (west), 2002 and 2003 in transect 2 (centre), and 2003 in transect 3 (east), which suggested a 'west–east gradient'. Four to 6 years separated each mast, which is consistent with the 3–8 years reported for Canada (Wang, 1974). Cleavitt *et al.* (2011) identified 1998, 2002 and 2006 as mast years in north-central New Hampshire. This synchronized and intermittent reproduction across the range of sugar maple could be due to endogenous rhythms (Kelly, 1994), climate conditions (Houle, 1999; Kelly & Sork, 2002), or seed predators (Tachiki & Iwasa, 2010). With respect to climate conditions, our data suggested that a warm and dry July in the previous year can induce mast seeding (Fig. 2.8). If some seedlings survived for several years, the presence of a seedling bank on the forest floor could compensate for the disadvantage of intermittent reproduction.

Model selection identified July mean temperature and precipitation (Table A2.5) as variables positively influencing sugar maple seedling density (Fig. 2.7). Sugar maple seeds have high soil moisture requirements during germination (Janerette, 1979). Our results showed that precipitation during April and May did not influence seedling density. This suggested that moisture during germination was not a limiting factor. We could hypothesize that warm July temperature, combined with high precipitation, might reduce seedling mortality. However, more studies are needed to identify the direct effect of climate conditions on seedling survival in sugar maple.

Climatic models for eastern North America have predicted a mean increase of +3.3 °C (minimum, 2.1 °C; maximum, 5.4 °C) and +1% in precipitation (min. -17%, max. +13%) in June, July and August for the decade 2080–2099 compared to 1980–1999 (Christensen *et al.*, 2007). Applying these mean values with our model for the northernmost site of transect 1 (1-D-A', Lac Labelle), our model predicted a mean increase of 5.3% (min. 2.1%, max. 18.4%) in sugar maple seedling densities for 2080–2099 compared to 1980–1999. However, the uncertainties around these

estimates progressively increased as temperature and precipitation increased (Fig. 2.7). Therefore, our predictions are limited by these uncertainties, but an increase in sugar maple seedling density is predicted to occur with climate change.

Species distributional range-shift predictions are based on niche models, climatic envelopes, and process-based models (McKenney et al., 2007; Morin & Thuiller, 2009). All of these approaches have more or less emphasized the influence of climate in predicting future species distributions, in part because other predictors are frequently unavailable. Our sugar maple seedling density model and habitat distribution models showed that the sugar maple limit was not controlled exclusively by the effect of tested climatic variables on regeneration. Therefore, northward expansion could possibly occur under different climatic scenarios (Iverson et al., 2008; Morin & Thuiller, 2009). All of the sites from the discontinuous zone showed very good sugar maple regeneration. For example, the northernmost site in transect 2 (2-D-A', Lac Pénobscot; Fig. 2.6) showed high sugar maple seedling densities and a typical J-shaped age structure, which indicates continuous recruitment over time. It is possible that soil characteristics (type and nutrients) could serve as explanatory factors. For example, a lack of rotten wood and leaf litter can represent a barrier to seedling establishment (Caspersen & Saprunoff, 2005), while an increase in calcium availability and a thinner litter layer are important for improving early sugar maple seedling survival (Cleavitt et al., 2011). The inclusion of soil characteristics may increase the predictive power of our sugar maple seedling density model. Sugar maple regeneration could also be influenced by factors such as herbivore grazing (Salk et al., 2011), seed predation (Hsia, 2009), disease (Cleavitt et al., 2011), and interspecific competition (Gravel et al., 2011). Accumulating evidence has shown that important non-climatic factors must be included to increase model predictive power for actual and future tree species range shifts (McMahon et al., 2011).

#### 2.7. Acknowledgements

We thank Normand Villeneuve, Claude Poulin and Renaud Dostie of MRNF du Québec who found many potential sites. Michèle Bettez and Jean-Pierre Faust of the Centre de Semences Forestières de Berthier (MRNF du Québec) performed the seed germination tests. Geneviève Trudeau, Elizabeth Turcotte, Sylvain Larouche, Chelsea Archambault, Emilie Tarroux, Manuella Strukelj-Humphery, Hedi Kebli and Dave Gervais provided invaluable field assistance. We are grateful to Marc J. Mazerolle for statistical advice, Igor Drobyshev for his help with BIOSIM, Osvaldo Valeria for his help with geographic analysis, and William F. J. Parsons for careful language editing. We are thankful for the referees' comments, which greatly improved the manuscript. This work was funded by an NSERC (Natural Sciences and Engineering Research Council of Canada) strategic grant (STPGP 336871) to Francine Tremblay.

# 2.8. Biosketches

Noémie Graignic is a PhD student in environmental sciences at the University of Québec in Abitibi-Témiscamingue (Canada). Her project consists of identifying factors that are influencing northward migration of sugar maple.

Francine Tremblay is a professor at the University of Québec in Abitibi-Témiscamingue (Canada). Her laboratory research interest focuses on population and conservation genetics of forest ecosystems.

Yves Bergeron is a professor of forest ecology at the University of Québec in Abitibi-Témiscamingue and University of Québec at Montréal (Canada). He is interested in forest dynamics at both the stand and landscape scales.

Author contributions: N.G., F.T. and Y.B. conceived the ideas; N.G. collected and analysed the data and led the writing of this paper.

# CHAPITRE III

#### DEVELOPMENT OF POLYMORPHIC NUCLEAR MICROSATELLITE MARKERS

# IN SUGAR MAPLE (ACER SACCHARUM MARSHALL)

# USING CROSS-SPECIES TRANSFER AND SSR-ENRICHED SHOTGUN PYROSEQUENCING

Noémie Graignic, Francine Tremblay et Yves Bergeron

Note technique publiée en 2013 dans Conservation Genetics Ressources, vol. 5, p. 845-848

# 3.1. Résumé

Un ensemble de 23 microsatellites polymorphes a été testé avec succès pour *Acer saccharum*. Dix-huit de ces nouveaux marqueurs polymorphes ont été développés en utilisant le séquençage de nouvelle génération et cinq ont été transférés à partir d'autres espèces d'*Acer*. Un total de 68 individus de deux populations a été génotypé pour tous les marqueurs à l'aide de multiplex. La diversité allélique moyenne observée était de 7,7, le polymorphisme variait de 2 à 16 allèles par locus. L'hétérozygotie globale était élevée, avec une hétérozygotie moyenne observée de 0,644. Deux loci, obtenus à partir de transfert entre espèces, et un locus obtenu par pyroséquençage, ont montré un déficit d'hétérozygotes, une déviation de l'équilibre d'Hardy-Weinberg, et une fréquence importante d'allèles nuls pour une ou deux populations. Aucune preuve de déséquilibre de liaison n'a été détectée. Enfin, 20 des 90 marqueurs pourraient être utilisés dans le futur pour des études de génétique des populations d'*Acer saccharum*.

# 3.2. Abstract

A set of 23 polymorphic microsatellites was successfully tested for *Acer saccharum*. Eighteen of these new polymorphic markers were developed using nextgeneration sequencing and five were transferred from other *Acer* species. A total of 68 individuals from two populations were genotyped at all markers using a multiplex approach. Mean allelic diversity was 7.7, polymorphism ranged from 2 to 16 alleles per locus. Overall heterozygosity was high, with an average observed heterozygosity of 0.644. Two loci that were obtained from cross-species transfer and one from shotgun pyrosequencing showed heterozygote deficiencies, deviations from Hardy–Weinberg equilibrium, and high frequencies of null alleles for one or both populations. No evidence for linkage disequilibrium was detected. Finally, 20 of the 90 markers could be used in future population genetics studies of *A. saccharum*.

#### **3.3.** Introduction

Sugar maple (*Acer saccharum* Marsh.) is a widespread and abundant tree in Northeastern North America (Godman *et al.* 1990). With anticipated climate change, models are predicting a decrease in abundance at the southern edge of its range (Iverson *et al.* 2008). This range contraction could lead to population fragmentation, progressive genetic erosion, and loss of genetic diversity in the most southern populations. Consequently, preservation of the original genetic diversity in marginal populations is a major issue in conservation biology (Hampe & Petit 2005). The aim of the present study was to develop nuclear microsatellite markers for sugar maple to enable future population genetic studies. Here, we report the isolation and characterization of novel microsatellite markers in *A. saccharum* using both cross-species transfer and next-generation sequencing.

# 3.4. Results & discussion

Total genomic DNA was extracted from dry leaf tissues of sugar maple from different sites in Québec, Canada, using DNeasy Plant Mini Kit (QIAGEN). A set of 9 polymorphic microsatellites has been developed for *Acer yangbiense* (Zhao *et al.* 2011), 8 for *Acer opalus* (Segarra-Moragues *et al.* 2008), 13 for *Acer mono* Maxim. (Kikuchi & Shibata 2008), and 8 for *Acer pseudoplatanus* (Pandey *et al.* 2004). Thirty of these markers were tested with *A. saccharum* and only five were transferable.

One DNA sample of sugar maple was sent to Genoscreen (Lille, France) for the construction of an enriched library. The library was sequenced using pyrosequencing on 454 GS-FLX Titanium (Roche Applied Science) and automated primer pairs were subsequently designed as follows Malausa *et al.* (2011). This provided 103,331 raw reads, 14,718 raw reads with microsatellite motifs, and 9,322 primer pairs, which

flanked 1,269 unique microsatellites. A total of 63 primers pairs were tested. Markers with sequences size that were evenly distributed between 90 and 320 bp, were selected to optimize chances of homogeneously distributed PCR product sizes among the selections that were available for PCR multiplex design. To ensure that we had polymorphic markers, we selected mostly di- and tetra-nucleotide microsatellite motifs and loci with the highest number of repeats (10–25). Since, only few primer pairs were designed for 200–300 bp size products, we downsized our selections to loci with six repeats.

Primer pairs were initially tested for amplification with unlabelled primers on five geographically separated individuals. Amplifications used 10 µL PCRs that contained 1× QIAGEN Multiplex Master Mix, 0.1  $\mu$ M of each primer, and 2–5 ng of genomic DNA. The thermal profile began with initial denaturation at 95 °C (15 min), followed by 35 cycles of denaturation at 94 °C (1 min), annealing at 54 °C (1 min 30 s) and extension at 72 °C (1 min), and ended with final extension at 72 °C for 45 min. Amplification products were visualized on agarose gel. Twenty-six loci were successfully amplified. They were further tested for polymorphisms using 11 individuals from 7 populations, following the same protocol with labelled primers (Table 3.1). Fragments were separated using an ABI 3730xl Sequencer (Applied Biosystems) at McGill University and Génome Québec Innovation Centre (Montréal, Canada). The resulting microsatellite profiles were examined using GENEMAPPER 3.7 (Applied Biosystems). A total of 23 polymorphic and unambiguous genotype profile markers were then combined into six PCR multiplex sets and two simplex PCR reactions, which were subsequently tested on 28 and 40 individuals from two populations located in western Québec (WQc; 46°45'0.40"N, 78°54'20.16"W) and eastern Québec (EQc; 46°36'34.34"N, 73 2'27.67"W), respectively (Table 3.1).

Locus	Drimor acquerect $(5! 2!)$	Damaat atmusture	Size renge (hp)	51 days	W W		WQc $(n = 28)$			c(n = 40)	GenBank	
Locus	Primer sequences (5 - 5 )	Repeat structure	Size range (op)	5 dye	IVI	$N_{\rm A}$	Ho	$H_{\rm E}$	$N_{\rm A}$	H <sub>0</sub>	$H_{\rm E}$	Accession no.
Am096	F: TAAGCTTCATACGCCATCAACCT R: GGCATCACCAAATCCAGACAC	(AG)9	133–152	6-FAM	1	5	0.192‡	0.305	4	0.205*§	0.417	AB303348
Am116	F: AACGCTACCGACTTCGCCAACT R: TGGAGGTCAAGTGCTGGAAACAA	(AG)6	230-273	PET	1	8	0.679	0.644	12	0.744	0.712	AB303350
Am607	F: CACACATGGGCTTCTCTATGAGT R: CATCCGCCAGTTGGTGAAT	(AG)11	133–160	PET	а	9	0.696‡	0.828	12	0.675‡	0.848	AB303355
Aop122	F: TCCCTCGTTAGATTTACGGTGGTTT R: TTCTTGATGACGATGATGACGATG	(TC)12	184–213	6-FAM	а	8	0.227*§	0.645	14	0.436*§	0.844	EF531293
Aop943	F: ACTGTGTAGGAGAGTGAGTGTGAA R: CTTCCCAAAGGTAGGAACCA	(GA)11	143–158	VIC	1	4	0.385	0.456	5	0.641	0.600	EF531298
SM11	F: AAGTTGCAGGAGGAGATTGC R: CACAATAACATGACCTTATGCCA	(CA)17	181–191	NED	4	5	0.667‡	0.622	3	0.575	0.581	KC731552
SM14	F: TTTTATGTAGAGCAACTCAACCCA R: TATCTGCTGCTTGACATGAACTT	(GA)15	72–112	NED	6	14	0.846	0.877	15	0.769	0.886	KC751436
SM21A	F: TAGTTGTGCACCAACCATGC R: TCCATCAAAACGCTGCTATG	(GAT)14	179–243	VIC	6	12	0.889	0.882	15	0.923	0.836	KC751437
SM22	F: CCAGAGCTTGAAGAAAATGTACG R: GGTAAGGGGTTGTTTATGCAAT	(GA)13	295-323	PET	3	14	0.917	0.895	12	0.872	0.888	KC751438
SM26	F: AAATCAAACAATTTCGCGCC R: GATTGATGCAAAGGCAGTGA	(GA)11	277-302	VIC	2	6	0.370*§	0.707	8	0.275*§	0.759	KC751439
SM27	F: TTTCTACTTTAGAGATGGAAACGG R: CCCTAAATCCCAAATCAGTGAA	(TC)10	242-260	PET	6	7	0.760	0.745	6	0.769	0.717	KC751440
SM29	F: TTAAACAAGCTGAACAACCCAA R: TCACACGGAAAGACATCAGC	(CTT)10	281-301	NED	6	6	0.815	0.730	5	0.692	0.667	KC751441
SM34	F: TGTGAAAATTCATAGATTTCTCAGTC R: ATACCACTTCAAAGCAACAAAACA	(GA)14	118–171	6-FAM	4	13	0.926	0.857	16	0.875	0.834	KC751442
SM36	F: GATGCTGATAAAGAAATGGAAACA R: GGCCCTCCCTGTTCTAAAGT	(AG)14	150-176	PET	2	9	0.963	0.814	10	0.950	0.833	KC751443
SM37	F: TGGGAGAAAATTAACATAGGATTT R: AGACTTGACTCTCCTAATCTTGGTG	(AG)14	174–198	NED	3	8	0.731	0.611	9	0.744	0.722	KC751444
SM42	F: ATTGCAGACGCACTCGTAAG R: CAACTCACTTGTGTTGGAATGC	(CT)12	96–128	NED	5	6	0.593	0.761	7	0.650	0.735	KC751445

Table 3.1 Primer sequences and characteristics of 23 microsatellite loci in two populations of Acer saccharum.

Locus	Primer sequences (5'-3')	Repeat structure	Size range (bp)	5' dye	Μ	WÇ	Qc (n = 2)	8)	EQc (	(n = 40)		GenBank Accession no.
	• · · /					$N_{\rm A}$	$H_0$	$H_{\rm E}$	$N_{\rm A}$	$H_0$	$H_{\rm E}$	-
SM47	F: TAAAGAGAAAACCTAGAAATCACTCTG R: CAACCAACTCGCATCAAGAA	(GA)10	201–219	PET	3	6	0.529‡	0.664	6	0.538‡	0.709	KC751446
SM51	F: GCCAAGTAACCACAGGCAAT R: CAAGAACCAGATCACAGATGTC	(TC)9	169–290	NED	5	5	0.481	0.474	6	0.625	0.609	KC751447
SM53	F: TTGGAAAGGATATTTGCATACG R: ATTTTTCTCCGAGCCATTGA	(AG)9	294–301	6-FAM	2	2	0.556	0.497	2	0.600	0.499	KC751448
SM55	F: ATGAGGGAGTAGAGGAGGCG R: TCAACAGATCCAAAACACGC	(TC)8	255-274	6-FAM	3	7	0.769	0.636	11	0.641	0.748	KC751449
SM56	F: GTTGTGGTCGTGAGACTTTAGG R: TAAACCGCTTGAGGACAACTT	(AG)8	287–299	PET	4	6	0.560	0.696	4	0.700	0.648	KC751450
SM57	F: ACTGCCACCAAAAGCAAAAG R: CCAAATCGAACACAAAACCCT	(AAG)7	203-218	NED	2	6	0.667	0.663	3	0.425	0.535	KC751451
SM60	F: CTTTAGAGCGGCCCAAGTTA R: GAGGGCCATTTTCAGTTGAG	(AAC)6	231-237	VIC	4	2	0.519	0.494	3	0.575	0.493	KC751452

Repeat structures of Am096, Am116, Am607, Aop122 and Aop943 were obtained from one individual over six sequenced individuals of *Acer* saccharum.

M, letter for simplexes and number for multiplexes; n, number of individuals analyzed;  $N_A$ , number of alleles observed;  $H_O$ , observed proportion of heterozygous individuals;  $H_E$ , expected heterozygosity; PHW, Hardy-Weinberg *P*-value.

\* Loci significantly deviating from HWE following Bonferroni correction (adjusted critical  $P \le 0.001$ )

‡ Loci in HWE following Bonferroni correction

§ Presence of null alleles

 Table 3.1 (to continued)

The number of alleles per locus ( $N_A$ ), and observed ( $H_O$ ) and expected heterozygosities ( $H_E$ ) were calculated in GENALEX 6.5 (Peakall & Smouse 2006). Deviations from Hardy–Weinberg equilibrium (HWE), linkage disequilibrium (LD) and heterozygote deficiencies were estimated using Genepop v. 4.2 (Rousset 2008). A sequential Bonferroni test for multiple comparisons was used to adjust *P* values to a predetermined experiment-wise error rate of 0.05 (Rice 1989). Null allele frequencies were estimated using FREENA (Chapuis and Estoup 2007).

Mean  $N_A$  was 7.7, with a range of 2–16 alleles per locus. Mean  $H_O$  and  $H_E$  were 0.644 and 0.687, respectively (ranging from 0.192 to 0.963, and from 0.305 to 0.895, respectively). Two of the 23 loci (Aop122 and SM26) in both populations and one locus (Am096) in one population failed to meet HWE (Table 3.1). All deviations from HWE were due to heterozygote deficiencies. Similar results have been previously reported for two loci (Aop122, Am096) in *A. opalus* and *A. mono* (Segarra-Moragues *et al.* 2008; Kikuchi & Shibata 2008). FREENA estimated zero to low (< 10 %) frequencies of null alleles for all of the loci, except Aop122, SM26 and Am096. There was no evidence of linkage disequilibrium between pairs of loci. Finally, 20 highly polymorphic microsatellite markers are now valuable resources for future landscape population genetics studies of *A. saccharum*.

# 3.5. Acknowledgments

This research was carried out as part of the doctoral research of the first author, with support from an NSERC (Natural Sciences and Engineering Research Council of Canada) strategic grant (STPGP 336871) to Francine Tremblay. We would like to thank Stephanie Ferreira, Stéphanie Duthoy and all team at GenoScreen, France (www.genoscreen.fr), and Valérie Catudal and Alexandre Belisle from Génome Québec (gqinnovationcenter.com) for their work. We are grateful to William F.J.

Parsons from the Centre for Forest Research (www.cef-cfr.ca) for careful language editing.

# CHAPITRE IV

# GENETIC DIVERSITY AND STRUCTURE AT THE NORTHERN LIMIT OF A WIDESPREAD

# NORTH-AMERICAN TREE,

# SUGAR MAPLE (*ACER SACCHARUM* MARSHALL)

Noémie Graignic, Francine Tremblay et Yves Bergeron

## 4.1. Résumé

Les modèles prédisent une expansion de nombreuses espèces d'arbres nordaméricains au nord de leur limite actuelle de répartition avec l'influence des changements climatiques. L'érable à sucre (Acer saccharum Marshall) atteint sa limite de répartition continue nordique dans le nord-est de l'Amérique du Nord, à la transition entre la forêt boréale mixte et la forêt feuillue tempérée. Notre travail s'articule autour de l'hypothèse que les populations nordiques fragmentées et marginales de la zone discontinue de répartition ont un patron distinct de structure et de diversité génétique. Nous avons analysé 18 marqueurs microsatellites dans 23 populations réparties le long de trois transects latitudinaux (situés à l'ouest, le centre et l'est du Québec). Chaque transect a été divisé en deux zones, continue et discontinue, sur la base de l'abondance des populations d'érable à sucre. Dans l'ensemble du Québec, la diversité génétique de cette espèce est élevée. Les populations d'érable à sucre situées dans la zone discontinue (nord) ont montré une plus faible diversité génétique comparativement aux populations de la zone continue (sud) - notamment pour les populations du transect ouest. L'analyse de la structure génétique révèle un patron défini avec un groupe au nord formé des populations d'érable à sucre les plus nordiques (1-D-A, 1-D-B, 2-D-A et 3-D-A). Nous avons observé un flux génétique asymétrique entre les populations, avec les populations du nord qui recoivent un flux génétique plus élevé en provenance des populations du sud. Les populations du sud du Québec semblent provenir de deux voies de migration distinctes ; la première venant de l'est et la seconde, plus tardive, venant de l'ouest. Les populations du nord-ouest ont montré une diversité génétique réduite et une différenciation entre les populations plus élevée. Nous avons par ailleurs observé une baisse de recrutement dans ces mêmes populations lors d'une précédente étude. Nous en arrivons à la conclusion que les populations d'érable à sucre du nord-ouest du Québec pourraient être plus à risque dans un contexte de changements climatiques.

#### 4.2. Abstract

Due to climate change, the ranges of many North-American tree species are expected to shift northward. Sugar maple (Acer saccharum Marshall) reaches its northern continuous distributional limit in north-eastern North America at the transition between boreal mixed-wood and temperate deciduous forest. We hypothesized that marginal fragmented populations from the discontinuous zone (northern populations) would have a distinct pattern of genetic structure and diversity. We analyzed 18 microsatellite loci variation in 23 sites distributed along three latitudinal transects (located in the west, centre and east of Québec). Each transect was divided into two zones, continuous and discontinuous, based on sugar maple stand abundance. Sugar maples showed a high level of genetic diversity in Québec, and a lower diversity in populations from the discontinuous zone (north) compared to those from continuous zone (south) - especially in populations in the western transect. North-western populations showed lower genetic diversity and higher population differentiation. These populations may be more at risk in a context of climate changes. The STRUCTURE analysis revealed the presence of 4 clusters with the presence of a high level of admixture. The four most northerly populations received a higher level of gene flow from the southern populations. Southern populations in Québec may have originated from two migration routes (from eastern first and then later from western routes). We concluded that north-western sugar maple populations in Québec need to be protected, due to there lower level of genetic diversity.

#### 4.3. Introduction

The Last Glacial Maximum (LGM) occurred approximately 21 000 years before present (-yr BP) and was the coldest period in recent climatic history (Jackson *et al.*, 2000). At that time, North American species were located south of the Laurentian Ice Sheet. After the ice retreated, postglacial vegetation ranges expanded and contracted repeatedly following climatic oscillations (McLachlan *et al.*, 2005). Given anticipated climate changes, we foresee the northern range expansion of several species (Iverson *et al.*, 2008).

This shift in range could be particularly marked in peripheral populations located along their distribution edge (Iverson et al., 2004). Generally isolated and smaller in size (effective population sizes) than core populations (Vucetich & Waite, 2003), these peripheral populations are expected to have a lower level of genetic diversity and a higher level of genetic differentiation among populations following genetic drift or bottlenecks (Eckert et al., 2008; Waters et al., 2013). In theory, bottlenecked populations are expected to experience a reduction of allelic richness and a limited decrease in heterozygosity. The consequence of this drift would be a reduction in the number of rare alleles (Nei et al., 1975). The central-marginal hypothesis expected that those patterns of diversity structure were link to the ecological marginality of the populations found at the periphery of the species distribution compared to core populations (Eckert et al., 2008). Whereas, "range shift following the last glacial maximum" hypothesis according more link to the influences of the migration during the last glacial maximum (Hampe & Petit, 2005). This range-shift hypothesis expected a lower within population diversity in peripheral populations located along the colonizing front (leading edge) and the low-latitude limit (rear edge) of the species' range, and contrary to the central-marginal hypothesis, a higher regional diversity at rear edge compare to core and leading edge (Hampe & Petit, 2005).

In the Northern Hemisphere, northern peripheral populations often exhibit local adaptations to cold stress (cold hardiness and earlier bud set in fall), and can potentially be maladapted to climate warming. Lower chilling requirements can result in early bud flush in spring under milder environments and higher susceptibility to spring frost damages (Howe *et al.*, 2003). On the other hand, climate warming could also improve their survival and increase the success of sexual reproduction (Alberto *et al.*, 2013). Wind-pollinated tree species may also be buffered against the effect of fragmentation on the genetic structure of peripheral populations due to long-distance gene flow that contributes to a decrease in differentiation among populations (Kremer *et al.*, 2012).

Sugar maple (Acer saccharum Marshall) is a widespread, abundant and longlived (300-400-yr) deciduous north-eastern American tree (Godman et al., 1990). This species is a monoecious, dichogamous, wind-pollinated and shade-tolerant broadleaf tree (Logan, 1965; Gabriel, 1968; Gabriel & Garrett, 1984) that forms uneven-aged stands (Majcen et al. 1984). It has major economic value for saw timber and syrup production in Canada and the United States (Godman *et al.*, 1990). Sugar maple reaches its northern range at the transition between the temperate deciduous and boreal mixed-wood forests (Saucier et al., 2003). Stands present in the boreal mixed-wood forest may possibly be remnants of formerly larger populations. Palynological reconstruction supports the idea that many species had migrated farther north of their current range after the retreat of ice sheet (Liu, 1990; Richard, 1993). Richard & Grondin, (2009) reported the establishment of disjointed sugar maple stands around 8500-yr BP in northern Québec. The decline of several tree species in the coniferous forest was concurrent with a climatic shift observed at the beginning of the Neoglacial period. This period was characterized by cooler and drier summers and an increase in fire frequency around 3000-yr BP (Carcaillet & Richard, 2000; Ali et al., 2008). Like many other tree species in the Northern Hemisphere, sugar maple is

predicted to migrate north of its current range (Goldblum & Rigg, 2005; Iverson *et al.*, 2008).

Earlier studies of sugar maple genetic variation were conducted with provenance tests. These studies revealed clinal variations in response to temperature and moisture stresses (Kriebel, 1957). Sugar maple provenances also showed an adaptation to altitudinal gradients for respiration, photosynthesis and specific leaf weight (Ledig & Korbobo, 1983). Analysis of neutral markers (allozymes and RAPD) revealed weak or no genetic differentiation among sugar maple populations in Canada (Perry & Knowles, 1989; Young *et al.*, 1993b) and across its range (Gunter *et al.*, 2000). On a regional scale (20 km in southern Québec), Young *et al.* (1993b) reported the presence of moderate levels of genetic structure among populations, while Diochon *et al.* (2003) found no structure among two populations (distance of 45 km in Ontario, Canada). No relationship between genetic and geographic distances was found between five northern populations (Perry & Knowles, 1989).

In this work, highly polymorphic microsatellite makers recently developed for sugar maple (Graignic *et al.*, 2013) were used to study the genetic diversity and structure of 24 sugar maple populations. Sugar maple populations were selected along three transects (west, central and east) that encompass the continuous-discontinuous species' range at the transition between the temperate deciduous and boreal mixed-wood forests in Québec. We hypothesized that marginal fragmented populations from the discontinuous zone (northern populations) would have; (1) a lower level of genetic diversity and higher genetic structure than populations from the continuous zone (southern populations), (2) the northernmost peripheral populations would have a lower number of rare alleles, and we predicted (3) a higher correlation between genetic and geographic distances in marginal populations. Two cohorts of mature trees ( $\geq 10$  cm d.b.h.) and saplings ( $1 \leq d.b.h. < 10$  cm), were sampled because temporal variations in genetic differentiation may be present between different

cohorts of regeneration (Mulcahy, 1975; Foré *et al.*, 1992). We tested the hypothesis that (4) mature trees would be more diverse than saplings; particularly for populations sampled in the discontinuous zone. Sugar maple is wind-pollinated and earlier studies reported a low level of genetic structure between populations, therefore we expected to find (5) high gene flow between populations and a weak signal of migration routes in Québec.

#### 4.4. Materials and methods

## 4.4.1. Study area and sampling

The study area was located at the northern range limit of sugar maple in Québec (eastern Canada; Fig. 4.1). The climatic gradient was divided into two zones based on the proportion of sugar maple stands in the continuous and discontinuous zones. The presence of sugar maple stands along the latitudinal gradient was estimated from an analysis of large inventory databases (for more details see Graignic *et al.*, 2014). The continuous zone lies within the sugar maple–yellow birch (*A. saccharum–Betula alleghaniensis* Britton) bioclimatic domain where sugar maple is abundant. The discontinuous zone was in the balsam fir–yellow birch (*Abies balsamea* (L.) Miller–*B. alleghaniensis*) bioclimatic domain and some northern sites were located in the transitional area of the balsam fir–white birch (*A. balsamea–Betula papyrifera* Marshall) bioclimatic domain (Saucier *et al.*, 2003).

Sites were distributed along three north–south transects (1, western; 2, center; 3, eastern) with eight sites per transect (four sites per transect per zone) (Fig. 4.1). Eleven of the sites selected were old-growth or rare forests that are classified as Exceptional Forest Ecosystems (EFE) by the ministère des Ressources naturelles du Québec (MRNQ, Québec Ministry of Natural Resources). All the sites were old-growth, uneven-aged stands, selected to be as similar as possible (for more details see Graignic *et al.*, 2014). Data were collected at 24 sites located between 45°51'N and

48°59' N latitude, and 70°21'W and 79°27' W longitude, at elevations ranging between 157 m and 493 m a.s.l.

Tissue samples (usually leaves or bark) were collected from 953 individual sugar maples in 2008 and 2009. Two cohorts were sampled per site: mature trees ( $\geq 10$  cm d.b.h.; *N*=12–22) and saplings (1  $\leq$  d.b.h. < 10 cm; *N*=20). Samples were frozen (-80° C) or dried (using silica-gel) until needed for genetic analyses.



**Figure 4.1** Map of the study area at the northern limit of sugar maple (*Acer saccharum* Marsh.) distribution in Québec showing the three transects (red lines), locations of the 23 study sites (2-D-B was remove) that show genetic results (red polygons, sites in discontinuous zone, and green polygons, sites in continuous zone; see Graignic *et al.* (2014) for more details on site characteristics), sugar maple–yellow birch (*Betula alleghaniensis*) bioclimatic domain (gray hatch), boundary of all bioclimatic domain limits (thin brown lines) (Saucier *et al.* 2003), and proportional membership of each sugar maple sample in clusters for five genetic groups (K = 4) inferred by STUCTURE analysis. For structure results, each individual was represented as a vertical line segment, and bottom-to-top clusters were: cluster 1, orange; cluster 2, blue; cluster 3, yellow and cluster 4, pink respectively.

#### *4.4.2. Molecular methods*

DNA was extracted using Extract-N-Amp<sup>™</sup> Plant PCR Kits (Sigma-Aldrich, Oakville, Ontario, Canada). All samples were genotyped for 18 variable microsatellite loci using PCR and genotyping protocols as previously described by Graignic *et al.* (2013) with the following modifications. Five different multiplex PCR sets and 37 cycles in the PCR amplification were used (see Table A4.1).

### *4.4.3. Markers genetic diversity*

The 2-D-B population, Lac aux Cèdres, had to be removed from the analysis due the low number of successful amplifications likely caused by poor sample conservation. Accordingly, a total of 913 individuals were selected for the subsequent analysis. For each locus, the total number of alleles  $(A_{\rm T})$ , mean number of alleles (A), mean observed  $(H_0)$  and expected  $(H_E)$  heterozygosity, and inbreeding coefficient (F<sub>IS</sub>) were estimated using FSTAT 2.9.3.2 (Goudet, 2001). Departure from Hardy-Weinberg equilibrium (HWE) per locus in each population, and linkage equilibrium between all pairs of loci in each population were calculated using an exact test implemented in GENEPOP 4.2.1 (Rousset, 2008). Markov chain parameters for HWE were 10 000 dememorizations, followed by 500 batches of 5000 iterations per batch. We corrected for multiple comparisons using a sequential Bonferroni adjustment of P-values to a predetermined experiment-wise error rate of 0.05 (Rice, 1989). Null allele frequencies were estimated using FREENA (10 000 replicates; Chapuis & Estoup, 2007). This software was chosen because it uses the Dempster et al. (1977) algorithm which provided the most accurate estimate of several algorithms tested in Chapuis & Estoup (2007). When populations are significantly differentiated the presence of null alleles overestimated fixation index  $(F_{ST})$  and genetic distances (Chapuis & Estoup, 2007). We performed a Mantel test (1000 permutations; using the mantel function in the VEGAN library from the software R 2.13.1; Oksanen et al.,

2011; R Development Core Team, 2011) between pairwise  $F_{ST}$  values with and without correction for null alleles calculated with FREENA.

# 4.4.4. Genetic diversity and differentiation between populations and groups

The mean number of alleles per locus (A), mean allelic richness ( $A_R$ ), mean observed heterozygosity ( $H_O$ ), mean expected heterozygosity ( $H_E$ ), pairwise  $F_{ST}$ , mean pairwise  $F_{ST}$  and inbreeding coefficient ( $F_{IS}$ ) were estimated using FSTAT. The indices were estimated for each population and on mature trees and saplings for each population separately. To assess whether population diversity parameters differed between zones and transects,  $A_R$ ,  $H_O$ ,  $H_E$ ,  $F_{ST}$  and  $F_{IS}$  were calculated by population (all individuals pooled, mature trees and saplings) within each zone (discontinuous and continuous), transects (1, 2 and 3) and zone in each transect separately using FSTAT (tested for significance using 1000 permutations).

# 4.4.5. Quantitative relationship

Data were analysed using version 2.13.1 of R package statistical software (R Development Core Team, 2011). Genetic indices ( $A_R$ ,  $H_O$ ,  $H_E$  and  $F_{IS}$ ) were compared using a linear mixed-model analysis (LMM, using the *lme* function in the NLME library; Pinheiro *et al.*, 2011), and a linear model for  $F_{ST}$ . Random effect included microsatellite marker for LMM. Different models were run to test the effects of: (1) stand characteristics such as basal area of mature sugar maple or/and sugar maple sapling ( $m^2$  ha<sup>-1</sup>), density of mature sugar maple or/and sugar maple sapling (stems ha<sup>-1</sup>), population size of mature sugar maple or/and sugar maple sapling (stems), (2) distance of each population to the northern limit (km) (see below), and (3) global models with a combination of one model of stand characteristics and its distance from the northern limit. Basal area and density were calculated in a quadrat of 0.16 ha at each site (see Graignic *et al.*, 2014). The population size of sugar maple was calculated separately for mature tree and sapling using their densities within the

quadrat and stand surface areas. For stand surface areas, we used data from EFE and ecoforestry maps that were obtained from MRNQ (MRNQ, 2013a, b). We fixed the northern limit between balsam fir-white birch (A. balsamea-B. papyrifera) and balsam fir-yellow birch (A. balsamea-B. alleghaniensis) bioclimatic domains and the distance of each sampled population to this limit was estimated with ARCMAP<sup>™</sup> 10.0 (Environmental Systems Research Institute (Esri), Toronto, Ontario, Canada); negative values were assigned to sites located to the north of the northern limit, i.e. in the balsam fir-white birch (A. balsamea-B. papyrifera) bioclimatic domain. Correlations between variables were examined, and assumptions of normality and homoscedasticity were graphically tested on global models. Where needed, the data were squared-root transformed to fit the assumptions. Fits for the 19 models were compared using the Akaike information criterion (AIC; *aictab* in the AIC<sub>C</sub>MODAVG library; Mazerolle, 2011), corrected for small sample sizes (AIC<sub>C</sub>). We computed AIC<sub>C</sub> and  $\Delta$ AIC<sub>C</sub> weights ( $\omega$ ) to determine the strength of evidence for each model (Burnham & Anderson, 2004). We then performed multimodal inference (modavg function in AIC<sub>C</sub>MODAVG library), where required ( $\Delta AIC_C \leq 4$ ), to calculate the parameter estimates and unconditional 95% confidence intervals (CI).

# 4.4.6. Correlation with sample size

Spearman's rank correlation coefficients were calculated between the sample size of each population and  $H_0$ ,  $H_E$ ,  $F_{IS}$  and  $F_{ST}$  using the *rcorr* function in the HMISC library from R (Harrell, 2013). As  $A_R$  was calculated using rarefaction method, a method that standardizes A on the basis of the size of the smallest number of samples, this parameter was already adjusted for sample size.

# 4.4.7. Proportions of rare alleles

We classified alleles, for each population with all individuals, on the basis of frequency (both common and rare, with two rare allele frequencies (f) < 0.05 and

0.02) and distribution (widespread and localized; with localized restricted to < 25% of the number of population that fixed localized at 5 populations and widespread at  $\geq$  6). This classification was based on Marshall & Brown (1975). Eight classes of alleles emerged from this classification and we considered four additional classes with private alleles (restricted to a single population). The same classification was used for each population for each cohort separately, except that only f < 0.05 was used because the lower allele frequency (f < 0.02; RP2), common and localized ( $f \geq$  0.05; CL5) categories was compared in the R software using two-way analysis of variance (ANOVA) with zone, transect and their interaction considered as fixed effects respectively. The assumption of normality and homoscedasticity was verified graphically; CL5 data was log-transformed. Global models were simplified by stepwise backward elimination of non-significant fixed-effect terms to produce the most parsimonious models.

# 4.4.8. Bottleneck tests

Evidence of bottlenecks in each population was evaluated using three tests: heterozygote excess, allele frequency mode shift and *M*-ratio. For the first test, heterozygote excess, we used BOTTLENECK 1.2.02 (Cornuet & Luikart, 1996) and each population was tested for three different mutation models: infinite allele mutation (IAM), stepwise-mutation model (SMM) and two-phase mutational model (TPM). In TPM, we choose 70, 90, 95 and 99% SMM, and we assumed a 12% variance in multi-step mutations (Piry *et al.*, 1999). The significance of the test under all models was performed using one-tailed Wilcoxon sign-rank test with 1000 permutations. The second approach, the allele frequency mode shift, was also conducted in each population using BOTTLENECK. The principle being that, under recent bottlenecks, rare alleles may occurred at low frequencies leading to a mode-

shift distortion of the typical L-shaped allele frequency distribution (Luikart & Cornuet, 1998).

We also calculated the *M*-ratio, the mean ratio of the number of alleles (*k*) to the range in allele size (*r*), using M\_P\_VAL software (Garza & Williamson, 2001). This method assumes that rare alleles are lost faster during a bottleneck, reducing the number of observed allelic states (*k*) faster than the size range of those alleles (*r*) resulting in a reduced *M*-ratio (M = k/r). To determine the significance of a ratio we compared our observed value to critical values ( $M_c$ ) that were established using the CRITICAL\_M program. When the observed *M*-value was significantly (P < 0.05) below  $M_c$  we assumed that a population had experienced a significant reduction in size (Garza & Williamson, 2001). To calculate  $M_c$  and *P*-values we used three different  $\theta$  values ( $\theta = 4N_e\mu = 1$ , 5 and 10;  $N_e$ , the effective population size; the mutation rate,  $\mu$ ) and the parameters were set as recommended by Garza & Williamson (2001) ( $\mu = 5.0 \times 10^{-4}$  mutants/generation/locus; probability of changes greater than one step,  $p_g = 0.10$ ; the size of non-one-step changes,  $\Delta_g = 3.5$ ). Each set of simulations consisted of 10 000 iterations.

#### 4.4.9. Isolation by distance

Isolation by distance was investigated using a Mantel test (1000 permutations; Pearson's correlation; using the *mantel* function in the VEGAN library from R) between the genetic  $[F_{ST}/(1-F_{ST})]$  or  $[D_S/(1-D_S)]$  and geographic distances (km). Pairwise  $F_{ST}$  and  $D_S$  (Nei, 1972) were calculated respectively using FSTAT and POPULATIONS 1.2.32 (Langella, 1999). Geographic distances were estimated using the *earth.dist* function in the FOSSIL library from R (Vavrek, 2012). We performed isolation by distance analysis for all populations, populations for each zone and each transect separately.

#### *4.4.10. Genetic structure*

Neighbor-joining trees were constructed based on  $D_S$  and using 1000 bootstraps over loci in POPULATIONS, one tree for all populations and one tree using cohort separate populations in two parts. The tree was visualized using TREEVIEW (Page, 1996).

We used the Bayesian clustering method implemented in STRUCTURE 2.3.4 (Pritchard *et al.*, 2000) to assess the number of populations (*K*), assign individuals to each *K* cluster, detect the population structure and show whether geographic populations are assigned in genetic populations. Our analyses were based on an admixture ancestral model with correlated allele frequencies and a priori sampling locations was used as prior information (LOCPRIOR setting). LOCPRIOR was used to detect any further structures unidentified by standard settings (Hubisz *et al.*, 2009). Twenty independent runs were performed for each value of *K* (1–26) with a burn-in of 100 000 following by 200 000 MCMC iterations. The most likely value of K was determined using the  $\Delta K$  criterion (Evanno *et al.*, 2005). STRUCTURE HARVESTER version 0.6.93 was used to extract the results and created a graphic of the  $\Delta K$  criterion (Earl & Vonholdt, 2012). The results were visualized for the best *K*, with DISTRUCT version 1.1. (Rosenberg, 2004).

Genetic variation was examined using the hierarchical analysis of molecular variance (AMOVA) based on  $\Phi_{PT}$  statistic and conducted in GENALEX 6.5b3 (Peakall & Smouse, 2006). We examined the significant difference (9999 permutations) between transects (1, 2 and 3), between populations within transect and within populations, and between zones (C and D), between populations within zone and within populations for all populations and for each transect separately. We also examined the significant difference between groups, between populations within the group and within populations for grouping A and B. Grouping A included group 1 (1-D-A, 1-D-B, 2-D-A, 3-D-A), group 2 (1-D-D, 1-C-A, 1-C-B, 1-C-C, 1-C-D), group 3

(2-D-C, 2-D-D, 2-C-A, 2-C-B, 2-C-C, 2-C-D), group 4 (3-D-B, 3-D-C, 3-D-D, 3-C-A, 3-C-B, 3-C-C, 3-C-D) and group 5 (1-D-C). We took into account the four clusters identified by STRUCTURE to create the grouping A (see Results and Fig. 4.1). To simplify grouping A, grouping B was created with removing group 5 and we have included the population 1-D-C in group 2.

# 4.4.11. Gene flow and population sizes

We used the Bayesian approach implemented in MIGRATE-N version 3.6 (Beerli & Felsenstein, 2001; Beerli, 2006) to assess the direction and amount of gene flow among populations. We used this software to calculate the posterior probability distribution of the mutation-scaled effective population size ( $\theta = 4N_{\rm e}\mu$ , where  $N_{\rm e} =$ effective population size and  $\mu$  = mutation rate per generation per locus) and the mutation-scaled past immigration rates ( $M = m/\mu$ , where m = migration rate) using a Metropolis-Hastings algorithm to explore all possible genealogies. The mutationscaled population size  $(\theta)$  for each population and the mutation-scaled immigration rates (M) between each pair of populations were estimated using the Brownian motion mutation model for microsatellites. A coalescent simulation explored the likelihood space for  $\theta$  and M. To limit the number of parameters, the amount of gene flow was estimated between four populations: from the transect 1 (1-D-C, 1-D-D, 1-C-A, 1-C-B, 1-C-C, 1-C-D), 2 (2-D-C, 2-D-D, 2-C-A, 2-C-B, 2-C-C, 2-C-D), 3 (3-D-B, 3-D-C, 3-D-D, 3-C-A, 3-C-B, 3-C-C, 3-C-D) and the most northern populations (1-D-A, 1-D-B, 2-D-A, 3-D-A). The starting values of  $\theta$  and M were generated from the  $F_{ST}$  estimates, as well as an exponential window prior set for both parameters (min = 0, mean = 5, max = 50 and  $\Delta$  = 5 for  $\theta$ ; min = 0, mean = 5, max = 50 and  $\Delta$ = 5 for M). We replicated two long chains of 10 000 000 genealogies recorded every 10 steps after a burn-in period of 10 000. The static heating scheme was set to 4 chains with temperatures of 100 000.0, 3.0, 1.5, and 1.0 and the swapping interval set at one. The program was run several times adding more sampling schemes and replications.

The starting parameters and the resulting estimates were compared until the results were congruent.

# 4.5. Results

# 4.5.1. Nuclear microsatellite diversity statistics

The total and mean number of alleles per locus ranged between 4 (SM60) and 31 (SM21A) and from 2.8 (SM60) to 14.8 (SM21A), respectively (Table A4.1). Significant departure from HWE was observed in 37 (9%) of the 414 possible population-locus combinations (Table A4.2). Four of the 18 loci (SM22, SM27, SM47 and SM56) failed to meet HWE in  $\geq 6$  (26%) populations. For those loci, deviations from HWE were due to heterozygote deficiencies ( $F_{IS} \geq 0.227$ ; Table A4.1). Significant linkage disequilibrium between pairs of loci (within populations) was detected only in one out of 3519 tests after Bonferroni correction (SM14 × SM34 in population 3-C-A, Lac Paul). The loci were therefore all considered genetically independent. Markers that showed a departure from HWE, also had high frequencies of null alleles ( $\geq 0.10$ ) in most populations (Table A4.3). Pairwise  $F_{ST}$  estimates obtained before and after corrections for null alleles were similar (r = 0.98, P < 0.001). In consequence, all loci (18) were retained in the analysis.

## 4.5.2. Genetic diversity and differentiation

Population 1-D-C (Rémigny) from the discontinuous zone of transect 1 had the lowest mean number of alleles per locus (A), mean allelic richness ( $A_R$ ), mean expected heterozygosity ( $H_E$ ) and inbreeding coefficient ( $F_{IS}$ ) for all individuals and both cohorts (mature trees and saplings) (Table 4.1). This population had the highest pairwise and global population differentiation ( $F_{ST}$ ) for saplings, mature trees and all individuals pooled (Table 4.1, A4.4). Pairwise  $F_{ST}$  were significant between 176, 73 and 51 pairs on 253 for the individuals pooled, saplings and matures trees, respectively after adjustment for multiple comparisons (Table A4.4).

**Table 4.1** Genetic variability estimates of sugar maple (*Acer saccharum*) populations in Québec for mature trees, saplings or all individuals. IP, all individuals pooled; M, mature trees; Sa, saplings; N, sample size; A, mean number of alleles;  $A_R$ , mean allelic richness;  $H_O$ , mean observed heterozygosity;  $H_E$ , mean expected heterozygosity;  $F_{IS}$ , inbreeding coefficient;  $F_{ST}$ , mean pairwise  $F_{ST}$ . Mean was determined using AI except for  $A_R^*$ .

Population	Cohort	Ν	A	$A_{\rm R}$	$A_{R}^{*}$	$H_0$	$H_{\rm E}$	$F_{\rm IS}$	$F_{\rm ST}$
1-D-A	IP	40	7.8	6.6		0.532	0.684	0.222	0.020
	М	20	6.2	5.1	4.9	0.566	0.668	0.152	0.019
	Sa	20	6.8	5.3	5.3	0.500	0.702	0.287	0.020
1-D-B	IP	40	8.1	7.0		0.590	0.693	0.149	0.015
	М	20	6.9	5.6	5.4	0.595	0.699	0.150	0.013
	Sa	20	6.9	5.4	5.4	0.587	0.690	0.150	0.016
1-D-C	IP	32	6.6	5.8	—	0.670	0.637	-0.051	0.041
	М	12	4.6	4.4	4.3	0.668	0.612	-0.090	0.051
	Sa	20	5.7	4.5	4.5	0.671	0.644	-0.042	0.039
1-D-D	IP	40	7.9	6.9	—	0.608	0.700	0.131	0.017
	М	20	6.8	5.4	5.2	0.591	0.700	0.156	0.020
	Sa	20	6.6	5.2	5.2	0.625	0.693	0.099	0.020
1-C-A	IP	40	8.7	7.5		0.620	0.704	0.120	0.011
	М	20	6.9	5.8	5.6	0.638	0.702	0.092	0.011
	Sa	20	7.3	5.5	5.5	0.604	0.709	0.147	0.009
1-С-В	IP	39	8.0	7.2		0.697	0.702	0.007	0.013
	М	19	6.8	5.8	5.6	0.705	0.689	-0.023	0.016
	Sa	20	7.1	5.5	5.5	0.687	0.707	0.029	0.012
1-C-C	IP	40	8.6	7.4	—	0.664	0.698	0.049	0.018
	М	20	7.2	5.7	5.5	0.648	0.679	0.045	0.022
	Sa	20	6.9	5.5	5.5	0.679	0.714	0.050	0.017
1-C-D	IP	40	8.5	7.3	_	0.546	0.696	0.216	0.011
	М	20	6.6	5.8	5.5	0.547	0.715	0.207	0.011
	Sa	20	7.2	5.4	5.4	0.535	0.686	0.220	0.009
2-D-A	IP	40	8.1	7.1	_	0.623	0.709	0.121	0.019
	М	20	6.8	5.5	5.3	0.617	0.700	0.119	0.025
	Sa	20	6.8	5.5	5.5	0.630	0.713	0.117	0.018
2-D-C	IP	40	8.5	7.4	—	0.548	0.715	0.234	0.014
	М	20	7.6	6.2	5.9	0.577	0.726	0.205	0.005
	Sa	20	6.4	5.3	5.3	0.520	0.706	0.264	0.022
2-D-D	IP	40	8.6	7.5		0.516	0.684	0.246	0.011
	М	20	7.0	5.8	5.6	0.493	0.683	0.279	0.008
	Sa	20	7.1	5.5	5.5	0.539	0.693	0.223	0.008
2-C-A	IP	40	8.2	7.1		0.596	0.684	0.129	0.012
	М	20	6.9	5.5	5.3	0.614	0.655	0.062	0.017
	Sa	20	6.8	5.3	5.3	0.579	0.705	0.180	0.010
2-C-B	IP	40	8.6	7.3		0.505	0.698	0.276	0.012
	М	20	6.9	5.7	5.4	0.495	0.692	0.284	0.009
	Sa	20	7.0	5.4	5.4	0.519	0.707	0.266	0.013
2-C-C	IP	42	8.6	7.2		0.589	0.691	0.147	0.010
	М	22	7.6	5.8	5.5	0.629	0.695	0.095	0.009
	Sa	20	6.7	5.2	5.2	0.544	0.688	0.209	0.012

Population	Cohort	Ν	A	$A_{\rm R}$	$A_{R}^{*}$	$H_0$	$H_{\rm E}$	$F_{\rm IS}$	$F_{\rm ST}$
2-C-D	IP	40	8.6	7.2		0.569	0.687	0.172	0.009
	М	20	6.9	5.6	5.3	0.617	0.696	0.114	0.009
	Sa	20	6.9	5.4	5.4	0.520	0.676	0.230	0.011
3-D-A	IP	40	8.0	7.0	_	0.496	0.710	0.302	0.021
	М	20	7.0	5.8	5.6	0.478	0.718	0.335	0.022
	Sa	20	6.2	5.0	5.0	0.514	0.706	0.272	0.018
3-D-В	IP	40	7.8	7.0	_	0.681	0.687	0.008	0.017
	М	20	6.9	5.7	5.5	0.696	0.687	-0.012	0.015
	Sa	20	6.9	5.5	5.5	0.667	0.690	0.034	0.018
3-D-C	IP	40	7.4	6.3	—	0.639	0.661	0.033	0.026
	М	20	6.4	5.2	5.0	0.630	0.659	0.045	0.020
	Sa	20	6.2	5.0	5.0	0.649	0.667	0.027	0.027
3-D-D	IP	40	8.3	6.5		0.656	0.694	0.054	0.019
	М	20	6.5	5.5	5.3	0.681	0.705	0.035	0.017
	Sa	20	6.2	5.1	5.1	0.613	0.681	0.074	0.021
3-C-A	IP	40	8.4	7.1	—	0.548	0.699	0.215	0.019
	М	20	6.9	5.7	5.4	0.574	0.701	0.181	0.016
	Sa	20	6.6	5.2	5.2	0.524	0.701	0.252	0.017
3-С-В	IP	40	9.0	7.6	—	0.618	0.700	0.117	0.011
	М	20	7.6	6.0	5.8	0.651	0.715	0.090	0.011
	Sa	20	6.9	5.3	5.3	0.585	0.681	0.140	0.014
3-C-C	IP	40	8.0	6.9	—	0.499	0.691	0.278	0.012
	М	20	6.8	5.6	5.4	0.479	0.691	0.307	0.008
	Sa	20	6.4	5.1	5.1	0.515	0.694	0.258	0.014
3-C-D	IP	40	8.2	7.0	—	0.716	0.715	-0.000	0.012
	М	20	6.9	5.6	5.4	0.740	0.720	-0.027	0.011
	Sa	20	6.9	5.4	5.4	0.691	0.713	0.031	0.011
Mean		40	8.2	7.0	5.3	0.597	0.693	0.138	0.016

 Table 4.1 (to continued)

Mean  $A_R$  was significantly lower in populations from the discontinuous vs. the continuous zone (6.8 vs. 7.2; Table 4.2). This difference in mean  $A_R$  among zones was significant only in transect 1 (Table 4.2) for the individuals pooled (6.6 vs. 7.3), mature trees (5.1 vs. 5.8) and saplings (5.1 vs. 5.5). The mean  $F_{ST}$  was higher in populations from the discontinuous vs. the continuous zone (0.025 vs. 0.009) and in transect 1 (0.030 vs. 0.004) but not transects 2 and 3 (Table 4.2). Population differentiation ( $F_{ST}$ ) was different both for mature trees (0.034 vs. 0.007) and saplings (0.029 vs. 0.003) in transect 1 (Table 4.2). No significant differences for  $A_R$  and  $F_{ST}$ , were found between transects (Table 4.2), between zones for transect 2 and 3 (Table

4.2) and between cohorts (matures trees and saplings; Table A4.5), or for all other genetic parameters.

## 4.5.3. Quantitative relationship

There were 10, 7, 19, 11 and 10 best models ( $\Delta AIC_C \le 4.0$ ; Table A4.6) for  $A_R$ ,  $H_O$ ,  $H_E$ ,  $F_{IS}$  and  $F_{ST}$  respectively. Parameter estimates derived from multi-model inference showed no influence on  $H_E$  and  $F_{IS}$  for all variables tested (Table 4.3). A positive and negative relationship was found between the distance of each site to the northern limit (D\_north), and  $A_R$  and  $F_{ST}$ , respectively. A positive relationship between mature sugar maple basal area and  $H_O$  was also observed (Table 4.3, Fig. 4.2).

# 4.5.4. Correlation with sample size

No significant correlations were found between sample size for each population and  $H_{\rm E}$  (r = 0.03; P = 0.502) or  $F_{\rm ST}$  (r = -0.38; P = 0.078). Significant correlations were observed between sample sizes for each population and  $H_{\rm O}$  (r = -0.11; P = 0.021), and  $F_{\rm IS}$  (r = 0.23; P = 0.000).
Cohort	Genetic	Zone			Transe	ect											
	Indice	D	С	<i>P</i> -value	1	2	3	P-value	1			2			3		
									Zone			Zone			Zone		
									D	С	P-value	D	С	P-value	D	С	P-value
ID	$A_{\rm R}$	6.840	7.246	0.0080	6.970	7.272	6.941	0.2450	6.601	7.339	0.0100	7.338	7.222	0.7380	6.707	7.176	0.0880
11	$H_0$	0.600	0.600	0.9950	0.616	0.567	0.611	0.2610	0.597	0.634	0.4460	0.566	0.567	0.9870	0.623	0.599	0.5800
	$H_{\rm E}$	0.689	0.698	0.2110	0.691	0.695	0.694	0.8610	0.680	0.701	0.0780	0.703	0.690	0.2630	0.687	0.701	0.1960
	$F_{\rm IS}$	0.129	0.140	0.7730	0.108	0.184	0.120	0.2520	0.122	0.095	0.6760	0.194	0.178	0.8380	0.093	0.146	0.4230
	$F_{\rm ST}$	0.025	0.009	0.0010	0.017	0.009	0.016	0.4630	0.030	0.004	0.0050	0.014	0.005	0.3910	0.022	0.011	0.2490
м	$A_{\rm R}$	5.464	5.707	0.1000	5.458	5.708	5.621	0.4100	5.138	5.779	0.0060	5.816	5.626	0.4660	5.526	5.716	0.4430
11/1	$H_0$	0.603	0.616	0.6670	0.620	0.583	0.622	0.4900	0.599	0.641	0.4030	0.568	0.593	0.6630	0.627	0.616	0.8430
	$H_{\rm E}$	0.698	0.697	0.4020	0.687	0.693	0.699	0.4730	0.676	0.696	0.1990	0.703	0.686	0.3010	0.692	0.707	0.2940
	$F_{\rm IS}$	0.126	0.116	0.8260	0.097	0.158	0.111	0.5220	0.114	0.079	0.6400	0.193	0.135	0.4870	0.094	0.129	0.6410
	$F_{\rm ST}$	0.023	0.009	0.0250	0.021	0.007	0.013	0.2540	0.034	0.007	0.0280	0.013	0.003	0.4560	0.019	0.006	0.3290
Sa	$A_{\rm R}$	5.200	5.343	0.1730	5.271	5.382	5.185	0.2730	5.078	5.464	0.0200	5.457	5.327	0.4790	5.130	5.240	0.5410
	$H_0$	0.598	0.585	0.6370	0.612	0.551	0.601	0.1610	0.596	0.627	0.5320	0.565	0.541	0.6510	0.620	0.582	0.4260
	$H_{\rm E}$	0.689	0.698	0.1960	0.692	0.697	0.692	0.7790	0.682	0.703	0.0650	0.703	0.693	0.4270	0.685	0.698	0.2950
	$F_{\rm IS}$	0.132	0.163	0.4690	0.116	0.210	0.131	0.1440	0.126	0.109	0.8110	0.197	0.219	0.7770	0.096	0.166	0.3310
	$F_{\rm ST}$	0.027	0.009	0.0010	0.016	0.011	0.016	0.7640	0.029	0.003	0.0220	0.018	0.008	0.4340	0.022	0.010	0.2480

**Table 4.2** Comparison of mean genetic variability estimates between populations of zones, transects and zones in each transect of sugar maple (*Acer saccharum*) in Québec for mature trees, saplings and all individuals.

IP, all individuals pooled; M, mature trees; Sa, saplings;  $A_{\rm R}$ , mean allelic richness;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm ST}$ , mean pairwise  $F_{\rm ST}$ ;  $F_{\rm IS}$ , inbreeding coefficient; D, discontinuous zone; C, continuous zone. Significant values ( $\alpha = 0.05$ ) given in bold.

Explained	Explanatory	Model-averaged	Unconditional	95% confidence	e interval
variable	variable	estimate	SE	Lower	Upper
$A_{\rm R}$	m_ers_BA	0.00605	0.00569	-0.0051	0.01721
	s ers BA	0.00155	0.0194	-0.03647	0.03957
	m ers d	0.0006	0.00048	-0.00034	0.00155
	s ers d	0.00002	0.00005	-0.00007	0.00011
	m ers PS	0	0	-0.00005	0
	s ers PS	0	0	0	0
	D_north*	0.00378	0.00119	0.00146	0.00611
H <sub>0</sub>	m_ers_BA*	0.00154	0.00062	0.00031	0.00276
	s_ers_BA	-0.00438	0.0023	-0.00889	0.00013
	m_ers_d	0.00011	0.00006	0	0.00022
	m_ers_PS	0	0	0	0
	s_ers_PS	0	0	0	0
	D_north	0.00011	0.00019	-0.00027	0.00049
$H_{\rm E}$	m_ers_BA	0.0002	0.0002	-0.0002	0.00059
	s_ers_BA	-0.00051	0.00076	-0.002	0.00097
	m_ers_d	0.00002	0.00002	-0.00002	0.00006
	s_ers_d	0	0	0	0
	m_ers_PS	0	0	0	0
	s_ers_PS	0	0	0	0
	D_north	0.00004	0.00004	-0.00004	0.00013
$F_{\rm IS}$	m_ers_BA	-0.00171	0.00094	-0.00355	0.00012
	s_ers_BA	0.00566	0.00347	-0.00115	0.01247
	m_ers_d	-0.0001	0.00008	-0.00026	0.00006
	s_ers_d	0	0.00001	-0.00002	0.00001
	m_ers_PS	0	0	0	0
	s_ers_PS	0	0	0	0
	D_north	-0.00003	0.00022	-0.00046	0.00041
F <sub>ST</sub>	m_ers_BA	-0.0002	0.00015	-0.00049	0.0001
	s_ers_BA	-0.00037	0.00052	-0.00139	0.00064
	m_ers_d	-0.00002	0.00001	-0.00004	0.00001
	s_ers_d	0	0	0	0
	m_ers_PS	0	0	0	0
	s_ers_PS	0	0	0	0
	D north*	-0.00006	0.00003	-0.00011	-0.00001

**Table 4.3** Parameter estimates and unconditional confidence intervals from multi-model inference of sugar maple (*Acer saccharum*)  $A_{\rm R}$ ,  $H_{\rm O}$ ,  $H_{\rm E}$  ( $\sqrt{\rm data}$ ),  $F_{\rm IS}$  and  $F_{\rm ST}$  in Québec. Asterisks identify parameters where confidence intervals excluded 0.

Term abbreviations: m\_ers\_BA, mature sugar maple basal area (m<sup>2</sup> ha<sup>-1</sup>); s\_ers\_BA, sugar maple sapling basal area (m<sup>2</sup> ha<sup>-1</sup>); m\_ers\_d, mature sugar maple density (stems ha<sup>-1</sup>); s\_ers\_d, sugar maple sapling density (stems ha<sup>-1</sup>); m\_ers\_PS, mature sugar maple population size (stems); s\_ers\_PS, sugar maple sapling population size (stems); D\_north, distance of each site to the northern limit (km); SE, standard error.



**Figure 4.2** Predicted sugar maple (*Acer saccharum*) genetic indices ( $A_R$ ,  $H_O$  and  $F_{ST}$ ) in response to all influential explanatory variables in best supported models based on multi-model averaging of all candidate models (n = 414 for  $A_R$  and  $H_O$ , and n = 23 for  $F_{ST}$ ) in Québec. Dashed lines indicate 95% confidence intervals.  $A_R$ , allelic richness;  $H_O$ , observed heterozygosity.

#### 4.5.5. Proportions of rare alleles

Rare alleles were detected in all categories with the individuals pooled (Table A4.7). The same percentages were detected in all populations except for populations in the discontinuous zone of transect 1, where rare and private alleles (f < 0.02; RP2 category) were not detected (Table A4.7). However, a two-way ANOVA showed that this category differed among zones but not between transects (Table 4.4) with slightly fewer alleles in the discontinuous zone (mean: 0.25%; 95% confidence intervals (CI): -0.14, 0.65) than in the continuous zone (mean: 0.98%; 95% CI: 0.60, 1.36).

Rare alleles at f < 0.05 were detected in all categories for mature trees and saplings (Table A4.8). The same percentages were detected for both mature trees and saplings in all populations, except in the common and localized category ( $f \ge 0.05$ ; CL5) where more alleles were present in saplings than in mature trees for transects 1 and 2 (Table A4.8). These differences were marginally significant (Table 4.4).

Response Num. Den.										
variable	Model	Explanatory variable	Num. Af	d f	F	D				
variable	Widdei	Explanatory variable	<b>u</b> .1.	u.1.	Г	Г				
RP2	1	zone	1	17	6.057	0.0248				
		transect	2	17	1.255	0.3101				
		zone × transect	2	17	0.117	0.8905				
RP2	2	zone	1	21	6.279	0.0215				
		transect	2	20	1.584	0.2312				
RP2	3	zone	1	21	6.158	0.0216				
CL5	1	zone	1	34	1.619	0.2119				
		transect	2	34	0.181	0.8350				
		cohort	1	34	3.807	0.0593				
		$zone \times transect \times cohort$	7	34	0.800	0.5931				
CL5	2	zone	1	42	1.884	0.1772				
		cohort	1	42	4.192	0.0469				
		$zone \times cohort$	1	42	1.891	0.1763				
CL5	3	cohort	1	45	4.029	0.0509				

**Table 4.4** Results of two-way ANOVA with interaction for percentage of allele frequency for the rare and private (frequency (f) < 0.02; RP2; n = 23) category and three-way ANOVA with interaction for the percentage of allele frequency in the common and localized ( $f \ge 0.05$ ; CL5; n = 46) category prior to model simplification (model 1) and subsequent models (models 2 and 3).

Num. d.f., numerator degrees of freedom; den. d.f., denominator degrees of freedom.

## 4.5.6. Bottlenecks

Inferences to be drawn from the heterozygosity excess tests were influenced by the mutational model. Under the IAM most of the populations were in excess of heterozygotes except for populations 1-C-D, 2-C-D and 2-D-D (Table 4.5). The mode shift model (SMM) showed no population in bottleneck.

For the *M*-ratio, no bottleneck was detected excepting three populations from the discontinuous zone (1-D-C, 1-D-D and 3-D-A) for  $\theta = 1$  (Table 4.5). No significant bottlenecks were found for the other  $\theta$  values (Table 4.5).

			Heterozy	gosity exc	ess				Mode	M-ratio						
Transect	Zone	Site	IAM	TMM				SMM	shift		$\theta = 1$		$\theta = 5$		$\theta = 10$	
				70%	90%	95%	99%				$M_{\rm c}$	P-value	$M_{\rm c}$	P-value	$M_{\rm c}$	P-value
1	D	1DA	0.03327	0.81539	0.98288	0.99088	0.99671	0.99860	Ν	0.805	0.806	0.434	0.737	0.942	0.715	0.987
		1DB	0.01041	0.71008	0.94065	0.96316	0.99203	0.99552	Ν	0.792	0.806	0.342	0.737	0.345	0.715	0.577
		1DC	0.01518	0.56748	0.91632	0.95063	0.98658	0.99306	Ν	0.713	0.806	0.033	0.737	0.377	0.715	0.596
		1DD	0.00032	0.86774	0.99800	0.99987	0.99999	0.99999	Ν	0.717	0.806	0.037	0.737	0.345	0.715	0.517
	С	1CA	0.02158	0.95929	0.99615	0.99903	0.99979	0.99992	Ν	0.768	0.806	0.196	0.737	0.734	0.715	0.902
		1CB	0.00200	0.43252	0.85814	0.92924	0.97586	0.98482	Ν	0.761	0.806	0.210	0.737	0.531	0.715	0.811
		1CC	0.01342	0.89393	0.99088	0.99968	0.99997	0.99999	Ν	0.810	0.806	0.478	0.737	0.499	0.715	0.744
		1CD	0.14186	0.90927	0.97842	0.98959	0.99671	0.99800	Ν	0.768	0.806	0.203	0.737	0.797	0.715	0.927
2	D	2DA	0.00021	0.48306	0.68015	0.72456	0.86774	0.91632	Ν	0.778	0.806	0.251	0.737	0.828	0.715	0.936
		2DC	0.00329	0.75246	0.98816	0.99763	0.99979	0.99990	Ν	0.780	0.806	0.270	0.737	0.856	0.715	0.954
		2DD	0.18461	0.94065	0.99398	0.99615	0.99883	0.99961	Ν	0.792	0.806	0.350	0.737	0.921	0.715	0.820
	С	2CA	0.01041	0.92298	0.98959	0.99800	0.99860	0.99968	Ν	0.805	0.806	0.432	0.737	0.943	0.715	0.986
		2CB	0.04488	0.93513	0.99088	0.99615	0.99883	0.99961	Ν	0.798	0.806	0.394	0.737	0.930	0.715	0.983
		2CC	0.04937	0.95063	0.99480	0.99720	0.99961	0.99979	Ν	0.775	0.806	0.233	0.737	0.817	0.715	0.915
		2CD	0.10607	0.95063	0.99398	0.99860	0.99997	0.99998	N	0.776	0.806	0.239	0.737	0.627	0.715	0.932
3	D	3DA	0.01184	0.38301	0.56748	0.58414	0.79144	0.85814	Ν	0.811	0.806	0.484	0.737	0.961	0.715	0.994
		3DB	0.00520	0.35094	0.82673	0.94581	0.97586	0.97842	Ν	0.755	0.806	0.142	0.737	0.665	0.715	0.820
		3DC	0.00010	0.63314	0.96673	0.99398	0.99860	0.99903	Ν	0.756	0.806	0.148	0.737	0.675	0.715	0.834
		3DD	0.00013	0.26131	0.61699	0.76585	0.91632	0.91632	Ν	0.769	0.806	0.264	0.737	0.773	0.715	0.900
	С	3CA	0.03327	0.94065	0.99615	0.99860	0.99968	0.99974	Ν	0.719	0.806	0.042	0.737	0.363	0.715	0.539
		3CB	0.02997	0.90927	0.99306	0.99552	0.99832	0.99860	Ν	0.813	0.806	0.483	0.737	0.961	0.715	0.990
		3CC	0.00385	0.98288	0.99800	0.99979	0.99987	0.99994	Ν	0.737	0.806	0.084	0.737	0.521	0.715	0.705
		3CD	0.00200	0.53386	0.75246	0.86774	0.95063	0.96316	Ν	0.720	0.806	0.847	0.737	0.363	0.715	0.517

 Table 4.5 Bottleneck results based on heterozygosity excess, Mode shift and M-ratio.

D, discontinuous zone; C, continuous zone. Significant values ( $\alpha = 0.05$ ) given in bold for heterozygosity excess.

#### 4.5.7. Isolation by distance

Isolation by distance was observed among populations (r = 0.36, P < 0.001) and zones (r = 0.35, P = 0.008 and r = 0.36, P = 0.002 respectively for continuous and discontinuous zones) using  $F_{\text{ST}}/(1-F_{\text{ST}})$ . Tests were not significant when each transect was tested separately (r = -0.06, P = 0.533; r = 0.33, P = 0.139; and r = 0.06, P = 0.390; respectively for transects 1, 2 and 3). Similar results were obtained using  $D_{\text{S}}/(1-D_{\text{S}})$  (Table 4.6).

8r-					
Group	Number of	$F_{\rm ST}/(1 - 1)$	$F_{\rm ST}$ )	$D_{\rm S}/(1-L)$	D <sub>S</sub> )
Oroup	populations	r	Р	r	Р
All populations	23	0.36	0.001	0.36	0.001
Zone					
Discontinuous	11	0.35	0.008	0.36	0.005
Continuous	12	0.36	0.002	0.36	0.002
Transect					
1	8	-0.06	0.533	-0.05	0.532
2	7	0.33	0.139	0.26	0.184
3	8	0.06	0.390	0.02	0.476

 Table 4.6 Isolation by distance analysis results performed separately for different geographic groups.

## 4.5.8. Genetic structure

The neighbor-joining tree separated the populations of the transect 1 (except for 1-D-C, Rémigny) from the populations of the two other transects (Fig. A4.1). However, bootstrap support (< 25%) was poor overall. The strongest link was found between populations 2-D-A and 3-D-A with 79% bootstrap support. When the populations were divided into two groups (mature trees and saplings), similar groupings occurred for 12 populations (bootstrap support range: 20–99) among which 7 had bootstrap support  $\geq$  50% (Fig. A4.1).

The STRUCTURE output produced the highest value of  $\Delta K$  at K = 4 (Fig. A4.2). Many individuals membership were assigned in the four clusters indicating high levels of admixture (orange, blue, yellow and pink clusters; Fig. 4.1), however a structure could still be drawn. Trees from the northern populations, 1-D-A, 1-D-B, 2-D-A and 3-D-A, were predominantly assigned to the orange cluster. The blue cluster was better represented in trees from populations of transect 3. High assignment probabilities to the blue and yellow clusters were found in the populations of transects 1 and 2. The pink cluster separated trees from population 1-D-C from all the others.

AMOVA analysis showed no genetic structure of the populations at the level of zone (0%) and transect (1%, P = 0.000) (Table A4.9). Most of the genetic variation was found within populations for all grouping (within 97%, P = 0.000; among 2–3%, P = 0.000; Table A4.9).

## 4.5.9. Gene flow and population sizes

The  $\theta$  value showed that the largest sugar maple population was in the eastern area of Québec (transect 3; mean  $\theta = 3.68$ ) and the smallest was in the centre (transect 2: mean  $\theta = 1.22$ ; Table A4.10, Fig. 4.3). Northern populations also had a low mean scaled effective population size ( $\theta = 1.43$ ). Remarkable levels of gene flow were observed among populations from the three transects with slightly more scaled immigration rate (*M*) from the east (transect 3) than the other two transects (Table A4.10, Fig. 4.3). Gene flow was asymmetrical with higher migration rates from the south (transects 2 and 3) toward the northern populations, while low and equal migration rates were observed between the west (transect 1) and northern populations (Table A4.10, Fig. 4.3).



**Figure 4.3** Migration pattern of four groups of Québec sugar maple (*Acer saccharum*) populations, using MIGRATE-N (Beerli, 2006). The groups (red lines for the northern group and green lines for other groups) are defined using the results of STRUCTURE. Mean mutation-scaled population sizes ( $\theta = 4N_{e}\mu$ , where  $N_{e}$  = effective population size and  $\mu$  = mutation rate per generation per locus) are given for each group. Arrows represent direction of migration, and the thicknesses of the arrows are proportional to mean mutation-scaled immigration rate ( $M = m/\mu$ , where m = migration rate; small arrow, M = 6.6-8.5; medium arrow, M = 15.0-17.8; big arrow, M = 21-26). Locations of the 24 study sites (2-D-B was not included in genetics analysis) are given in red polygons for sites in discontinuous zone, and given in green polygons for sites in the continuous zone (see Graignic *et al.*, 2014 for more details). Sugar maple–yellow birch (*Betula alleghaniensis*) bioclimatic domains are in gray hatch and the boundary of all bioclimatic domains are given in thin brown lines (Saucier *et al.*, 2003).

## 4.6. Discussion

#### 4.6.1. Genetic diversity levels and differentiation in sugar maple populations

Using eighteen microsatellite markers we recently developed, we detected a high level of genetic diversity in sugar maple populations (see Table 4.1). No other data was available for natural populations of sugar maple with microsatellite markers. When we compared our results to other tree species such as *Acer mono* ( $H_E = 0.80$ , range: 0.70–0.85) and two *Quercus* species ( $H_E = 0.82$ , range: 0.77–0.86), sugar maple did not have the highest genetic diversity (Table 4.7). The level of genetic diversity reported for sugar maple populations with allozymes (see Table 4.7) was relatively high (mean  $H_E = 0.122$ ; range: 0.064–0.275) and quite similar to other *Acer* 

species, except for *Acer pseudoplatanus* ( $H_E = 0.280$ , range: 0.254–0.319; Table 4.7). Overall  $H_E$  was lower for sugar maple than for other north-eastern American tree species (Table 4.7).

Large variations in the inbreeding coefficient estimates was observed for sugar maple populations in Québec ( $F_{IS} = 0.138$ , range: -0.051–0.302). Similar results were reported with microsatellites for *Acer takesimense* ( $F_{IS} = 0.28$ , range: 0.08–0.47) (Takayama *et al.*, 2013) and for *Populus tremuloides* (see Table 4.7).

We found a very low level of genetic differentiation between sugar maple populations ( $F_{ST} = 0.016$ , range: 0.009–0.041). Values were similar to those of *Quercus* ( $F_{ST} = 0.01-0.02$ ; Table 4.7).

**Table 4.7** Comparison of genetic diversity using allozyme and microsatellite markers for old-growth, undisturbed or natural populations of tree species in northeastern North America, and *Acer* species around the world.

Species	A	$A_{\rm R}$	$H_0$	$H_{\rm E}$	F <sub>IS</sub>	$F_{\rm ST}/G_{\rm ST}$	Reference
Allozymes							
Angiospern	1						
Acer platanoides	1.88 (1.50-2.17)	—	0.134 (0.085-0.179)	0.133 (0.090-0.172)	-0.014 (-0.379-0.185	) —	Rusanen et al. (2000)*
Acer platanoides	2.0 (1.6–2.4)	—	0.126 (0.038-0.195)	0.132 (0.053-0.191)	0.066 (-0.085-0.285)	) 0.099	Rusanen et al. (2003)
Acer pseudoplatanus	2.78 (2.56-3.00)		0.293 (0.237-0.327)	0.280 (0.254-0.319)	-0.032 (-0.159–0.085	) 0.019 (0.003–0.035)	Belletti et al. (2007)
Acer macrophyllum	1.71 (1.5–2.2)	—	0.118 (0.102-0.160)	0.152 (0.102-0.189)	0.166 (-0.086-0.332)	) 0.054	Iddrisu and Ritland (2004)
Acer saccharum	1.95 (1.64–2.18)			0.110 (0.098-0.132)	—	0.033	Perry and Knowles (1989)
Acer saccharum	2.9		0.15	0.148	—	0.012	Foré et al. (1992a)
Acer saccharum	2.03			0.109	0.073	0.017	Young et al. (1993a)*
Acer saccharum	1.98 (1.78–2.41)			0.112 (0.088-0.138)	0.042 (-0.095-0.177)	) 0.033	Young et al. (1993b)
Acer saccharum	3.21 (2.50–3.75)	2.17 (1.66–2.98)	0.130 (0.072–0.294)	0.133 (0.064–0.275)	0.025 (-0.055-0.078)	) 0.060 (0.060–0.114)	Baucom et al. (2005)
Castanea dentata	1.69 (1.50–1.89)		0.184 (0.135-0.264)	0.151 (0.096–0.196)	-0.226	0.110	Huang et al. (1998)
Fagus grandifolia	2.9 (2.9–2.9)		0.387 (0.382-0.392)	0.395 (0.383-0.407)	0.024	0.063	Houston and Houston (1994)
Fagus grandifolia	3.0 (2.78–3.33)	—	0.163 (0.150-0.175)	0.165 (0.150–0.179)	0.009	0.030	Houston and Houston (2000)
Populus tremuloides	2.7 (2.1–2.9)		0.125 (0.101-0.160)	0.235 (0.207-0.270)	0.462 (0.295–0.568)	0.068	Hyun et al. (1987)
Populus tremuloides	2.6 (2.2–2.9)	_	0.217 (0.197–0.242)	0.220 (0.193–0.244)	0.017	0.003	Lund et al. (1992)
Quercus rubra	2.08 (1.8–2.3)	_	_	0.186 (0.145–0.245)	0.100	0.092	Sork et al. (1993)
Gymnospern	1						
Picea glauca	3.03 (2.17–3.83)	2.14 (1.86–2.37)	0.342 (0.221–0.414)	0.344 (0.199–0.412)	0.002 (-0.092-0.087)	) —	O'Connell et al. (2006)
Picea rubens	1.47 (1.25–1.64)	—	0.075 (0.059–0.092)	0.079 (0.061–0.104)	0.043 (-0.037-0.224)	) 0.007 (0.003-0.011)	Hawley and DeHayes (1994)
Pinus strobus	2.31 (2.24–2.37)	—	0.126 (0.125–0.126)	0.153 (0.149–0.157)		—	Buchert et al. (1997)*
Pinus strobus	2.35 (2.23–2.50)	—	0.215 (0.185–0.216)	0.195 (0.181–0.216)	-0.139 (-0.273-0.407	) 0.061	Rajora et al. (1998)
Thuja occidentalis	1.6 (1.5–1.8)	_	0.116 (0.102–0.133)	0.129 (0.113–0.141)	0.106	0.073	Lamy et al. (1999)
Microsatellites							
Angiospern	1				0.000		
Acer mono	12.63			0.802	-0.008	—	Kikuchi et al. $(2009)$
Acer mono	—	8.3/(/.38-9.65)		0.80 (0.70-0.85)	0.2/(0.20-0.32)	_	Takayama et al. $(2012)$
Acer okamotoanum	—	6.60 (6.11-/.41)	0 5 40 (0 5 42 0 5 5 2)	0.72(0.66-0.76)	0.18 (0.03–0.24)	_	Takayama et al. $(2012)$
Acer pseudoplatanus	_	— • (0 (0 70 5 05)	0.548 (0.543-0.553)	0.5/4 (0.5/3-0.5/4)	-	_	Pandey (2005)
Acer pseudosieboldianun	n	4.60 (3.79-5.25)	0.40 (0.32–0.46)	0.61 (0.53-0.68)	0.33 (0.21–0.43)	-	Takayama et al. (2013)
Acer saccharum	8.2 (6.6–9.0)	/.0 (5.8–/.6)	0.597 (0.496-0.716)	0.693 (0.63 / -0. / 15)	0.138 (-0.051-0.302	0.016 (0.009-0.041)	Our study
Acer skutchii	2.1 (1.5–2.5)	2 92 (2 50 4 22)		0.129(0.054-0.247)	0.1/4 (0.131 - 0.159)	0.075	Lara-Gomez et al. $(2005)$
Acer takesimense	12	3.82 (3.59–4.23)	0.38(0.30-0.47)	0.53(0.48-0.58)	0.28(0.08-0.47)		Lind and Calling $(2013)$
Quercus ellipsoidallis	13 145(12,15)	_	0.07(0.02-0.72)	0.79(0.77-0.81)	0.145(0.10-0.19)	0.01	Lind and Gailing (2013)
Quercus rubra	14.3(13-13)		0.75(0.70-0.75)	0.84(0.83-0.86)	0.12(0.07-0.17)	0.02	Lind and Galling (2013)
Populus tremuloides	8.83 (7.58–10.08)		0.465 (0.45–0.48)	0.0/(0.01-0./3)	0.30 (0.21–0.39)		Namroud et al. (2005)

 Table 4.7 (to continued)

Species	A	$A_{\rm R}$	Ho	$H_{\rm E}$	FIS	$F_{\rm ST}/G_{\rm ST}$	Reference
Populus tremuloides	—	5.99 (3.34-6.83)	—	0.758 (0.613-0.801)	0.019 (-0.12-0.19)	0.086	Callahan et al. (2013)
Populus tremuloides	7.44 (6.25-8.2)		0.556 (0.478-0.704)	0.725 (0.691-0.767)	0.201 (-0.054-0.325)	0.032	Wyman et al. (2003)
Gymnosperm	l						
Pinus strobus	9.43 (9.23–9.62)		0.522 (0.505-0.538)	0.607 (0.599-0.615)		—	Rajora et al. (2000)*
Pinus strobus	_	6.7	0.47	0.48	0.01	_	Marquardt and Epperson (2004)*
Thuja occidentalis	9.58 (7.83-11.17)	9.21 (7.66-10.68)	0.590 (0.505-0.640)	0.600 (0.519-0.662)	0.019 (-0.025-0.050)	) —	Pandey and Rajora (2012a)
Thuja occidentalis	7.3 (5.67–9.33)	6.8 (5.16-8.51)	0.601 (0.492-0.662)	0.611 (0.490-0.678)	0.013 (-0.063-0.105)	0.078	Pandey and Rajora (2012b)
Thuja occidentalis	7.8 (5.0–10.0)	5.9 (4.6-6.9)	0.734 (0.463-0.883)	0.773 (0.712-0.840)	0.145	0.065	Xu et al. (2012)

\* Study used logging or non-natural forest, we only reported populations from old-growth and natural forests. A, mean number of alleles per locus;  $A_{\rm R}$ , mean allelic richness;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm IS}$ , inbreeding coefficient,  $F_{\rm ST}$ , mean pairwise  $F_{\rm ST}$ ,  $G_{\rm ST}$ , mean pairwise  $G_{\rm ST}$ . Range values are given in parentheses.

#### *4.6.2. Genetic variation and the central–marginal hypothesis*

Most of the genetic variation was found within sugar maple populations, a frequent pattern for long-lived perennials, outcrossing and wind dispersed seeds plants (Nybom, 2004). Despite this fact, a genetic structure was observed and isolation by distance was detected.

The central-marginal hypothesis (Eckert *et al.*, 2008) was tested by delineating bioclimatic domains that reflected sugar maple abundance (see Graignic et al., 2014 for more details). As predicted by the central-marginal hypothesis, we found a lower allelic richness  $(A_R)$  and higher genetic differentiation  $(F_{ST})$  between populations in the discontinuous zone that represents the periphery of the sugar maple range in Québec (Table 4.2). The relationship between the proximity to the northern limit and genetic diversity was negative whereas it was positive with genetic differentiation (Table A4.6, 4.3, Fig. 4.2a, c). When these relations were tested for each transect separately, the pattern (discontinuous/continuous) remained significant only for the western transect (transect 1). The central-marginal hypothesis predicted that a lower abundance could lead to lower genetic diversity (Eckert et al., 2008). A lower level of sugar maple recruitment (lower seedlings density) in the discontinuous zone was only previously observed in transect 1 (Graignic *et al.*, 2014). The association between lower sexual reproductive capacity and lower genetic diversity in peripheral populations was also suggested for Thuja occidentalis (Xu et al., 2012; Paul et al., 2014) and Cirsium acaule (Jump et al., 2003) at their northern ranges. The number of rare and private alleles (f < 0.02 and in one population) was more frequent in the continuous zone than in the discontinuous zone (Table 4.4). These results are similar to those obtained for *Thuja occidentalis* at its north-eastern range with lower  $A_{\rm R}$ , higher  $F_{ST}$ , and a lower number of private alleles in peripheral populations than in core populations (Pandey & Rajora, 2012b).

No recent bottleneck was detected under the SMM while most populations were in bottleneck under the IAM (Table 4.5). The IAM is better at detecting subtle genetic bottlenecks, however is also known to occasionally identify heterozygosity excess in nonbottlenecked populations (Luikart & Cornuet, 1998). Luikart & Cornuet (1998) thought that the true model for most microsatellite loci was an intermediate between IAM and SMM. They recommended interpreting that a recent bottleneck had occurred when the IAM is significant and the SMM is close to significant  $(0.10 \ge P >$ 0.05). We found P-values above 0.86 for SMM when IAM was significant (Table 4.5). Therefore, we concluded that most of the sugar maple populations had not been through a recent bottleneck. The *M*-ratio analysis shown that three populations may have undergone an historical bottleneck when the mutation-scaled effective population size was estimated at  $\theta = 1$ , but no bottleneck was detected at  $\theta = 5$  or higher (Table 4.5). Among these populations, only 3-D-A were assigned to a distinct cluster (defined by STRUCTURE) with a mutation-scaled estimate population size close to  $\theta = 1$  (MIGRATE results), and had probably undergone a bottleneck in a distant past (Table A4.10 and Fig. 4.3). Xu et al. (2012) also reported the presence of bottlenecks in northern peripheral white cedar populations.

The results of other studies were not consistent with the central-marginal hypothesis (Rajora *et al.*, 1998; Gapare *et al.*, 2005; González-Martínez *et al.*, 2005; Hoban *et al.*, 2010; Jadwiszczak *et al.*, 2011). For example, Gapare *et al.* (2005) found higher inbreeding in peripheral (both marginal and continuous) populations of *Picea sitchensis* located at the northern and southern edges of the species range than core populations (both marginal and continuous). This pattern seems more consistent with the range–shift model that followed the LGM (Hampe & Petit, 2005). Hoban *et al.* (2010) tested both hypotheses and their results best supported the range shift model as the determinant of genetic structure for *Juglans cinerea*, rather than the ecological marginality in itself. In the present study, only the northern peripheral

populations (both discontinuous and continuous) were sampled in all sugar maple distribution and therefore we were not able to determine whether the range-shift or the central-marginal model was the best. The same problem was encountered in the works of González-Martínez *et al.* (2005) and Jadwiszczak *et al.* (2011) because only the eastern and south-western peripheral populations of *Pinus pinaster* and *Betula humilis* were sampled, respectively.

We found a genetic structure in our study area, it seems that core and peripheral populations were sampled. The most northern populations of each transect (1-D-A, 1-D-B, 2-D-A and 3-D-A), were assigned to the orange cluster (Fig. 4.1, 4.3). These populations were probably more representative of the northern peripheral zone than the discontinuous zone in itself.

## 4.6.3. Special case of 1-D-C, Rémigny

The population from Rémigny (1-D-C) was distinct compared to all the others. It had the lowest level of diversity (A,  $A_R$ ,  $H_E$  and  $F_{IS}$ ), the highest  $F_{ST}$  and a bottleneck was detected (M-ratio at  $\theta = 1$ ). Both neighbor-joining trees and structure analysis clearly disjointed this population from the others. These differences were probably due to the recent establishment of sugar maple at this site. A study conducted by Pilon (2013) at a sugar maple stand nearby (approximately 7 km away) dated the sugar maple's arrival few years to 80-yr ago, after a fire which occurred 160–220-yr ago. The presence of sugar maple at Rémigny was sporadic or scarce during the last millennia (Pilon, 2013). Probably the presence of this species was too recent to reach level of diversity because Lesser *et al.* (2013) has shown that, more than 230-yr are needed to reach allele saturation after *Pinus ponderosa* establishment.

### 4.6.4. Mature tree and sapling cohorts

In many coniferous tree species, the level of heterozygosity increases with the age of the trees (seeds vs. adults, seeds vs. seedlings, different age classes), possibly

due to higher fitness of heterozygous individuals (Bush & Smouse, 1992; Nijensohn *et al.*, 2005). Ballal (1994) reported that sugar maple embryos had lower heterozygosity and higher  $F_{IS}$  than seedlings (1-yr to  $\leq 1$  cm basal area) and mature trees ( $\geq 30$  cm dbh). Foré *et al.* (1992) found no difference between sugar maple embryos, 1-yr seedlings and three other cohorts (dbh  $\leq 2$  cm, 15–25 cm and  $\geq 40$  cm). We observed very small differences between the cohorts. First,  $F_{IS}$  was always higher for saplings than mature trees, however this difference was not significant. We also found a significant positive relationship between mature sugar maple basal area and  $H_0$  (Table A4.6, 4.3 and Fig. 4.2). This suggests that largest mature sugar maple trees were heterozygotic at a higher number of loci (Table A4.6, 4.3). However, no influence of mature sugar maple basal area on  $F_{IS}$  was detected and therefore did not support the hypothesis of higher heterozygosity in older cohort.

## 4.6.5. Genetic signature of migration routes

We found the largest population size in the eastern populations followed by the western and the central ones respectively (Table A4.8; Fig. 4.3), and a very different genetic structure between the eastern and the western transects (blue and yellow clusters in Fig. 4.1). That suggests two distinct migration routes, which separated a long time ago, or originated from different glacial refugia.

Earlier sugar maple migrations in Québec started approximately 9000-yr BP in south-eastern Québec (step 1 in Fig. 4; Lavoie & Richard, 2000) and near lake St-Jean (8500-yr BP; Richard & Grondin, 2009), and they could come from a north-westward or westward expansion (step 2 in Fig. 4.4). Sugar maple pollen was recorded 8,000-yr ago in southern Ontario, just south-west of Québec 6000-yr BP and south-west of Québec 5500-yr BP (steps 3, 4 and 5 in Fig. 4.4; Richard, 1980; Bennett, 1987). Therefore, the arrival of sugar maple in western Québec occurred around 3000-yr later than it did in the east and corroborating evidence of two distinct migration routes in Québec. Our hypothesis matches also Braun's hypothesis which

was based on pollen data analysis that suggests beech-maple associations moved northward via two routes following the ice retreat (Braun, 1950). The first movement was north-east, along the Appalachians and the second was directly northward from Louisiana to North Carolina. This second movement was delayed by a warm and dry post-glacial period in Ohio and Indiana. The northern lake states' beech-maple forests were derived from westward expansion of the first route and not from migrations directly from the south (Braun, 1950). Jackson *et al.* (2000) identified, from pollen and macrofossils data, two maple glacial refugia in the LGM (21 000-yr BP; Fig. 4.4). Interestingly, a genetic signature of these same refugia was identified for red maple (*Acer rubrum*) (McLachlan *et al.*, 2005) but no data was available for sugar maple across the US.

Southern populations in central Québec (transect 2) had a lower population size and were clustered with western and eastern populations (transect 1 and 3; Fig. 4.1). We hypothesized that these populations originated from an admixture of westward and eastward Québec migration (step 7 in Fig. 4.4). Similar routes for northern populations could also be drawn but with considerably more contribution from central-eastern populations (transects 2 and 3; step 7 in Fig. 4.4). We deduced that this is because very low contributions from the southwestern populations to the northernmost western populations were found (Table A4.10; Fig. 4.3). In addition, western population 3-D-A (Lac Patrick) was in bottleneck possibly because it represented a remnant of a larger northern sugar maple population. Some tree species were in decline in northern Québec 3000-yr BP because of the cool and dry climate (Carcaillet & Richard, 2000; Ali *et al.*, 2008). Therefore, population 3-D-A could be far older than other northern sugar maple populations and thus the withdrawal effect could have lasted longer.



**Figure 4.4** Map of sugar maple migration routes. Hatched zone, sugar maple modern distribution. Grey zones, isopoll map of maple pollen in Last Glacial Maximum; black circles, presence of maple macrofossils in Last Glacial Maximum (Jackson *et al.*, 2000). Green arrows, hypothesized beech-maple association migration deduced from pollen (Braun, 1950); black arrows, sugar maple migration deduced from our microsatellite genomic DNA study and from pollen diagrams; numbers, hypothesized chronological steps of sugar maple migration (see Discussion for more details). Red polygons, northern populations group deduced from ours STRUCTURE results; green polygons, our studied southern populations. Standardized two-letter state and province abbreviations were used.

#### 4.7. Conclusions

This is the first time that the genetic diversity and structure of sugar maple has been studied over a wide range at its northern range. Although there is evidence of some impact of marginality on its genetic structure, the question remains open of whether the "ecology marginality at the periphery" (central-marginal hypothesis; Eckert *et al.*, 2008) or "range shift following the last glacial maximum" (Hampe & Petit, 2005) models, better describe the pattern of variation. We found a high level of genetic diversity in Québec and a lower diversity in northern populations. However, no clear conclusion could be drawn to validate the hypothesis that northern populations were remnants of a larger population that migrated further north of the species range after the retreat of ice sheet. It is possible that it only occurred in the north-eastern part of the Québec range with population 3-D-A (Lac Patrick), which showed a bottleneck, and that other northern populations came from long distance dispersion (Cain *et al.*, 2000). Northern populations could be more susceptible to climate change than core populations, especially the north-western populations (1-D-A and 1-D-B) that seem more isolated from the south-western populations and that appear to be on the leading edge of a westward migration (Fig. 4.3). Nuclear and cytoplasmic sequences could be useful in the future because they can be more reliably ordered into an evolutionary sequence than through the use of microsatellites (Ellegren, 2004). Additional palynological data would also be essential to accurately establish the arrival of the northernmost populations and their migration routes.

## CHAPITRE V

## GENETIC CONSEQUENCES OF SELECTION CUTTING

# ON SUGAR MAPLE (ACER SACCHARUM MARSHALL)

Noémie Graignic, Francine Tremblay & Yves Bergeron

#### 5.1. Résumé

Les impacts de la déforestation et de la coupe sélective sur le niveau et la structure de la diversité génétique des arbres sont bien documentés. La coupe de jardinage consiste à prélever périodiquement des arbres choisis individuellement ou par petits groupes dans un peuplement mature à structure d'âge inéquienne. Le traitement crée de petites ouvertures du couvert forestier sur de grandes superficies et génère des structures forestières qui diffèrent souvent de celles des forêts anciennes. Dans cette étude, nos objectifs étaient d'évaluer si ce type de récolte avait un impact sur la diversité génétique de l'érable à sucre (Acer saccharum Marshall). Nous avons analysé la diversité génétique de semis, de jeunes arbres et d'arbres matures dans des peuplements coupés et des forêts anciennes au Québec, Canada. Nous avons observé un taux d'hétérozygotie observée ( $H_0$ ) plus élevée et un coefficient de consanguinité  $(F_{IS})$  plus bas pour les arbres matures comparativement aux cohortes plus jeunes, ceci dans les peuplements coupés et les vieilles forêts. Nous avons détecté la présence d'un goulot d'étranglement récent dans tous les peuplements jardinés. Toutefois, il ne s'est avéré significatif que pour le modèle du nombre infini d'allèles possibles (IAM). Les autres indices de diversité génétique (richesse allélique, hétérozygotie observée et attendue et allèles rares) étaient similaires entre les deux types de forêts. Nous en arrivons à la conclusion que la coupe de jardinage a peu ou pas d'effet sur la diversité génétique de l'érable à sucre. Nous n'avons détecté aucune présence d'érosion génétique au Québec après une récolte. Néanmoins, la présence d'un goulot d'étranglement dans les peuplements coupés indique que nous ne sommes pas en mesure de prédire avec précision l'effet cumulatif de plusieurs récoltes successives et c'est pour cette raison que nous recommandons un suivi à long terme.

#### 5.2. Abstract

Selection cutting is a treatment for forests with uneven-aged structures. It emulates tree-by-tree replacement that creates small openings, which differ from oldgrowth forests structure. Because previous studies of others type of forest fragmentation showed no consistent influence on sugar maple (Acer saccharum Marshall) genetic diversity; in this study, our objectives were to evaluate, using new genetic markers, whether this type of harvesting has a perceptible impact on genetic diversity of this species. Genetic diversity among seedlings, saplings and mature trees was compared between logged and old-growth forest stands in Québec, Canada. We found higher observed heterozygosity  $(H_0)$  and lower inbreeding coefficient  $(F_{IS})$  in mature trees than in younger regeneration cohorts in both forest types. We observed a recent bottleneck in all selection cuttings stands. However, this bottleneck was significant only for the infinite allele mutation (IAM) model and other genetic indices of diversity (allelic richness, observed and expected heterozygosity and rare alleles) were similar between forest types. We concluded that the effect of selection cutting on the genetic diversity of sugar maple was subtle, and no strong evidence of genetic erosion was detectable in Québec stands after one harvest. Nonetheless, we could not accurately predict the cumulative effect of multiple loggings and recommend longterm monitoring.

#### 5.3. Introduction

Forest ecosystems are exposed to natural (i.e. fire, windstorms, pests, diseases) and human (i.e. urbanization, logging, agriculture) disturbances. The ongoing effects of climate change are superimposed on those disturbances in boreal (Bergeron et al., 2010), temperate (Fischer et al., 2013), and tropical (Brodie et al., 2012) forests. In north-eastern North America, long-term logging has led to changes in forest composition and structure (Boucher et al., 2009). Following logging, subsequent reductions in tree population size (tree density and forest cover) may increase genetic drift and bottlenecks and, ultimately, decrease genetic diversity (Finkeldey & Ziehe, 2004). Lost of diversity may decrease potential adaptation of population to cope with global changes (Hamrick, 2004). Decreased in genetic diversity have been observed in white spruce (Picea glauca [Moench] Voss; Rajora, 1999) and eastern white pine (Pinus strobus L.; Buchert et al., 1997; Rajora et al., 2000). These studies have reported reductions in the mean number of alleles, low frequency alleles and rare alleles, together with level of heterozygosity. In contrast, other studies have shown no negative effects of logging on tree genetic diversity for white spruce (Fageria & Rajora, 2013), black spruce (Picea mariana [Miller] BSP; Perry & Bousquet, 2001), and black walnut (Juglans nigra L.; Robichaud et al., 2010). These results suggest that high intra-population genetic diversity, greater longevity and efficient longdistance pollen dispersal, which are generally observed in trees, could counterbalance the losses and attenuate genetic losses following harvesting (Hamrick, 2004).

Sugar maple (*Acer saccharum* Marshall) is a long-lived deciduous tree that forms uneven-aged stands. In addition to the syrup that it produces, this species has major economic value as saw timber in north-eastern North America (Majcen *et al.*, 1984; Godman *et al.*, 1990). It is insect- (bee) and wind-pollinated, and is shade-tolerant (Logan, 1965; Gabriel & Garrett, 1984). In Canada, its range extends from southern

Ontario and Québec, in the temperate deciduous forest, northwards into the boreal mixed-wood forest (Little, 1971; Saucier *et al.*, 2003).

The abundance of sugar maple has increase in North America over time following the colonization (Siccama, 1971; Nicholson et al., 1979). In Québec, forest clear-cutting and the deliberate setting of fires by settlers in the 20th century have favoured the expansion of sugar maple (Boucher *et al.*, 2006; Dupuis *et al.*, 2011). Since 1983, selection cutting, a type of partial cut, has become the most common harvesting treatment for sugar maple stands in Québec (Majcen, 1994). It consists of the removal of 25-35 % of the volume of trees having a d.b.h. (diameter at breast height,  $1.3m \ge 10$  cm (MRNFPQ, 2003). Cutting cycles occur at regular intervals of 15–25 years to sustain stand structure and maintain its quality. Stands that are harvested using single-tree selection have a smaller number of large and defective trees than do old-growth stands (Angers et al., 2005). Ten years after selection cutting in a sugar maple stand, the diameter growth of sugar maple, American beech (Fagus grandifolia Ehrhart) and yellow birch (Betula alleghaniensis Britton) generally increased in comparison to pre-harvested stands (Forget *et al.*, 2007). Fifteen years after an experimental selection cut in sugar maple stands, basal area, radial growth, and the development of sugar maple, American beech and yellow birch saplings were found to be greater compared to control plots (Majcen et al., 2005). However, a recent survey of hardwood forests that were treated with selection cuts in Québec showed that after 10 years, about 55 % of commercially harvested stands were less productive than experimental plots that had been treated with selection cutting (Guillemette et al., 2013). This response is possibly due to the low initial (preharvest) quality of sugar maple stands or to higher intensity harvesting of good quality stems (Guillemette et al., 2013).

Rainville (2007) noted that basic knowledge of hardwood tree species genetic diversity across Canada is lacking. Previous work on the genetic diversity of sugar

maple has focused on fragmentation effects that are incurred by agriculture and clearcutting. Foré *et al.* (1992b) compared canopy tree ( $\geq$  30 cm d.b.h.; pre-fragmentation cohort) and juvenile ( $\leq$  1cm basal diameter; post-fragmentation cohort) distributions in isolated forest patches in an agricultural landscape; the patches had been isolated for at least 50 years (1935-1988). Genetic differentiation between patches with respect to canopy trees was greater than for juveniles. Foré et al. (1992b) concluded that the level of gene flow between forest patches was higher in this open landscape. In examining a subsample of seeds taken from the same forest patches (only those with a minimum distance of 30–40 m between two patches), Ballal (1994) showed that embryos had a lower genetic diversity than juveniles and canopy trees within patches; however, both Foré et al. (1992a) and Baucom et al. (2005) found no differences between cohorts within patches. In comparing the genetic diversity of 1year-old seedlings between patches and continuous forest, Young et al. (1993a) and Young & Merriam (1994) found greater genetic diversity in the fragmented forest. Eighty years after one clear-cut, sugar maple seedlings had a lower percentage of polymorphic loci and allelic richness estimates compared to an old-growth forest (Baucom et al. 2005). These results show no consistent pattern of habitat fragmentation on sugar maple genetic diversity.

Selection cutting is a treatment that emulates tree-by-tree replacement in forests with uneven-aged structures. However, it creates small openings in large areas of canopy and often generates forest structures that differ markedly from old-growth forests. In this study, we evaluated whether this type of harvesting had an effect on sugar maple genetic diversity. Using new microsatellite, highly polymorphic markers (Graignic et al. 2013), we compared the genetic structure of mature and regenerating sugar maple cohorts in logged stands (selection cutting) to adjacent unlogged stands (old-growth). We hypothesized that: (1) the removal of around 30 % (by volume) of mature trees would reduce the level of diversity (mean number of alleles, rare alleles,

and the level of heterozygosity) of the mature sugar maple cohort compared to the old-growth mature sugar maple cohort; (2) sugar maple seedlings, which had established after logging, would have a lower level of genetic diversity due to reductions in the number of potential parent trees (mature trees); and (3) selection cutting may change levels of genetic structure among cohorts (adults–saplings–seedlings). The results are discussed with reference to the short-term consequences of logging on the genetic resources of sugar maple.

#### 5.4. Materials and methods

## 5.4.1. Study area and sampling

The study took place in the continuous portion of the northern range of sugar maple, i.e., southern Québec, Canada (Fig. 5.1). The study area was located between 45°46' N and 46°8' N, and 75°53' W and 74°59' W, at elevations ranging between 305 and 410 m (Table A5.1). This zone lies within the sugar maple–yellow birch (*A. saccharum–B. alleghaniensis*) and sugar maple–basswood (*A. saccharum–Tilia americana* L.) bioclimatic domains, where sugar maple is abundant (Saucier *et al.*, 2003).

The stands that had been subjected to selection cutting (SC) were harvested once, during winter 1990 (at the end of 1990 and early 1991). Our sampling was conducted in 2008. The SC was a commercial harvest that consisted of the removal of 25–35 % of tree basal area every 15–25 years (MRNFPQ, 2003). The old-growth (OG) stands, which had not been logged, were classified as Exceptional Forest Ecosystems (EFE) by the *ministère des Ressources naturelles du Québec* (Québec Ministry of Natural Resources; MRNQ, 2013). The logged stands (SC) were paired with adjacent unlogged stands (OG). The two sites (OG1 and SC1) were located in the sugar maple–basswood bioclimatic domain, while the remaining ones were in the sugar maple–yellow birch bioclimatic domain (Fig. 5.1). All 6 sites, except for OG2, were

similar to those described by Angers *et al.* (2005), and OG2 and OG3 are similar to 2-C-D and 2-C-C, which were described by Graignic *et al.* (2014), respectively. The respective distances between logged and unlogged sites were 17.3 km (OG1-SC1), 7.8 km (OG2-SC2), and 1.6 km (OG3-SC3) (Table A5.1).



**Figure 5.1** Map of the study area, which is situated at the northern continuous distributional limits of sugar maple (*Acer saccharum*) in Québec, showing the locations of the 6 study sites (circles, old-growth sites; triangles, selection cutting sites), sugar maple–yellow birch (*Betula alleghaniensis*) bioclimatic domain (slanted hatching), sugar maple–basswood (*Tilia americana*) bioclimatic domain (vertical hatching), and boundaries of all bioclimatic domain limits (thin lines) (Saucier *et al.*, 2003).

Tissue samples (usually leaves or bark) from 360 sugar maple individuals were collected in 2008 and 2009. Three cohorts were sampled per site: mature trees (M,  $\geq$  10 cm d.b.h., N = 20-22), saplings (Sa,  $1 \leq dbh < 10$  cm, N = 20), and seedlings (S1, dbh < 1 cm, N = 18-20). Seedling age was determined as described by Graignic *et al*.

(2014), with seedlings corresponding to seed that had been produced between 1986 and 2006. The mast of oldest seedlings originated from 1996 (Graignic *et al.* 2014). Most seedlings had germinated after logging, while saplings and mature trees were present prior to logging. We included an additional class, i.e. post-harvested seedlings (S2, N = 11-17), which had only seedlings that originated from seeds produced between 1991 and 2006. Samples were dried over silica gel and maintained at room temperature until they were needed for genetic analyses.

## 5.4.2. Molecular methods

DNA was extracted using Extract-N-Amp<sup>TM</sup> Plant PCR Kits (Sigma-Aldrich, Oakville, ON, Canada). All samples were genotyped for 18 variable microsatellite loci using PCR and genotyping protocols as previously reported (Graignic *et al.*, 2013). The following modifications to the protocol were applied: (1) five different multiplex PCR sets were used (see Table A5.2); (2) 0.1  $\mu$ M of each primer; and (3) 37 cycles in the PCR amplification procedure.

## 5.4.3. Marker genetic diversity

For each locus, the total number of alleles ( $A_T$ ), mean number of alleles per locus (A), mean observed ( $H_0$ ) and expected ( $H_E$ ) heterozygosity, and the inbreeding coefficient ( $F_{IS}$ ) were estimated using FSTAT 2.9.3.2 (Goudet, 2001). Departure from Hardy–Weinberg equilibrium (HWE) per locus in each population was tested, together with linkage equilibrium between all pairs of loci in each population, using exact tests that were conducted in GENEPOP 4.2.1 (Rousset, 2008). Markov chain parameters for HWE were 10 000 dememorizations, followed by 500 batches of 5000 iterations per batch. We corrected for multiple comparisons using a sequential Bonferroni adjustment of *P*-values to a predetermined experiment-wise error rate of 0.05 (Rice, 1989). Null allele frequencies were estimated using FREENA (10 000 replicates; Chapuis & Estoup, 2007). This program was selected because it uses the

algorithm of Dempster *et al.* (1977), which provided the most accurate estimate among the several algorithms that were tested by Chapuis & Estoup (2007). We performed a Mantel test (1000 permutations) between pairwise  $F_{ST}$  values with and without correction for null alleles that were calculated with FREENA. These tests were performed using the *mantel* function in the VEGAN library (Oksanen *et al.*, 2011) within the R statistical environment (version 2.13.1, R Development Core Team, 2011).

## 5.4.4. Genetic diversity and differentiation between cohorts and forest types

For each stand, the mean number of alleles per locus (A), mean allelic richness ( $A_R$ ), mean observed heterozygosity ( $H_O$ ), mean expected heterozygosity ( $H_E$ ), pairwise  $F_{ST}$ , mean pairwise  $F_{ST}$  and inbreeding coefficient ( $F_{IS}$ ) were estimated using FSTAT. Tests for heterozygote deficiency were performed using GENEPOP (Markov chain parameters: 10 000 dememorizations, followed by 500 batches of 5000 iterations per batch; Fisher's exact tests). Tests were calculated for each stand using data for pooled individuals (PI), mature trees (M), saplings (Sa), and seedlings (S1 and S2) separately. Allelic richness was estimated for each cohort and sampled sites using a rarefaction method (El Mousadik & Petit, 1996).

To compare  $A_{\rm R}$ ,  $H_{\rm O}$ , He and  $F_{\rm IS}$  between forest types (OG and SC), pairs (1, 2 and 3), and cohorts (M, Sa and S1, or M, Sa and S2), we performed a linear mixedmodel analysis (LMM, using the *lme* function in the NLME library of R; Pinheiro *et al.*, 2011). The fixed effects were forest type, pair and cohort and their two- and three-way interactions, and microsatellite marker was considered as a random effect. We used two data sets (M, Sa and S1; M, Sa and S2). Assumptions of normality and homoscedasticity were verified graphically. Models were simplified by stepwise backward elimination of non-significant fixed effects terms to produce the most parsimonious model. There was no significant difference between forest types for  $F_{\rm ST}$  and this index was subsequently not included in the analysis. To compare the same indices, we also performed the analysis using FSTAT (see Table A5.6).

#### 5.4.5. Partitioning of molecular variation

The structure of the genetic variation was determined using the hierarchical analysis of molecular variance (AMOVA) that is implemented in GENALEX 6.5b3 (Peakall & Smouse, 2006). Genetic differentiation among populations was estimated by the  $\Phi_{PT}$  statistic. We performed separate significance tests (9 999 permutations) between pairs (1, 2 and 3), between forest types (OG and SC) within pairs and within stands, and between pairs, between cohorts (M, Sa and S1) within pairs and within cohorts for OG and SC.

## 5.4.6. Allele frequencies

The effect of selection cutting on allele frequencies was tested using four classes: common ( $f \ge 0.75$ ), intermediate ( $0.75 > f \ge 0.25$ ), low ( $0.25 > f \ge 0.01$ ) and rare (f < 0.01). We also added two classes, low ( $0.25 > f \ge 0.05$ ) and rare (f < 0.05), using a level of 0.05, as suggested by Marshall and Brown (1975). Allele frequencies were estimated using FSTAT. The analyses were performed among forest types or cohorts (M, Sa, S1 and S2).

## 5.4.7. Bottlenecks

To test for recent reduction in effective population size following selection cutting, we used BOTTLENECK 1.2.02 (Cornuet & Luikart, 1996) for each population. Evidence of bottlenecks was tested using heterozygote excess and allele frequency mode shift tests. For the test of heterozygote excess, we used 3 different mutation models: infinite allele mutation (IAM), stepwise-mutation model (SMM), and two-phase mutational model (TPM). In TPM, we chose 70, 90, 95 and 99 % SMM, and 12 % variance of multi-step mutations was assumed (Piry *et al.*, 1999). Significance tests used one-tailed Wilcoxon signed-rank tests. Population bottlenecks cause a mode-

shift distortion of the typical L-shaped allele frequency distribution (Luikart & Cornuet, 1998).

#### 5.5. Results

## 5.5.1. Genetic variability of microsatellite markers

The total number of alleles per locus ranged from 3 to 28, while the mean number of alleles per locus ranged from 2.8 to 17.7 (Table A5.2). Five of the 18 loci (SM22, SM27, SM47, SM55 and SM56) failed to meet HWE in 3 populations (Table A5.3). These deviations from HWE were due to heterozygote deficiencies ( $F_{IS} \ge 0.296$ ; Table A5.2). Four other markers also showed heterozygote deficiencies ( $F_{IS} \ge 0.246$  for SM29, SM51, SM53 and SM60). These loci exhibited higher frequencies of null alleles ( $\ge 0.10$ ) (Table A5.4). Their genetic structures (pairwise  $F_{ST}$ ) were similar before and after correction for the presence of null alleles (r = 0.90, P = 0.003). All loci were considered independent because no significant linkage disequilibrium between pairs of loci within forest types was detected after Bonferroni correction. All loci (18) were used in further analyses.

#### 5.5.2. Genetic diversity and differentiation between cohorts and forest types

The mean number of alleles per locus (*A*) ranged from 9.2 (SC3) to 9.5 (SC1, OG2, SC2 and OG3) between stands, and from 5.9 (S2 of SC3) to 7.6 (M of OG3) for cohorts (Table 5.1). No significant difference was detected between cohorts and forest types in terms of allelic richness ( $A_R$ ) and expected heterozygosity ( $H_E$ ), using both LMM (in R) and FSTAT (Table A5.6, A5.7).

**Table 5.1** Genetic variability estimates of sugar maple (*Acer saccharum*) populations in Outaouais, Québec, for mature trees (M), saplings (Sa), seedlings (S1), pooled individuals (PI) and only seedlings that emerged after logging (S2), separately. To compare allelic richness between each cohort (M, Sa, S1 and S2), we calculated allelic richness using those cohorts in the same database ( $A_R^*$ ). Populations were old-growth (OG) forest or had received a single selection cutting (SC) in end of 1990–beginning of 1991.

Populations	Cohorts	Ν	Α	$A_{\rm R}$	$A_{R}^{*}$	$H_0$	$H_{\rm E}$	$F_{\rm IS}$	$F_{\rm ST}$
OG1	М	20	7.1	6.5	5.4	0.583	0.692	0.158***	0.003
	Sa	20	6.6	6.0	5.1	0.440	0.676	0.348***	0.000
	<b>S</b> 1	20	7.1	6.4	5.5	0.496	0.684	0.274***	0.004
	PI	60	9.4	9.0		0.507	0.686	0.260***	0.000
	S2	17	6.6	5.3	5.3	0.477	0.678	0.296***	0.004
SC1	М	20	7.2	6.6	5.5	0.583	0.706	0.174***	0.001
	Sa	20	6.9	6.2	5.3	0.480	0.697	0.312***	0.006
	<b>S</b> 1	20	6.9	6.3	5.3	0.521	0.690	0.245***	0.003
	PI	60	9.5	9.1		0.529	0.699	0.243***	0.001
	S2	15	6.3	5.4	5.4	0.505	0.687	0.265***	0.004
OG2	М	20	6.9	6.4	5.3	0.617	0.696	0.114***	0.001
	Sa	20	6.9	6.3	5.4	0.520	0.676	0.230***	0.003
	<b>S</b> 1	20	7.0	6.3	5.4	0.516	0.693	0.255***	0.002
	PI	60	9.5	9.0		0.552	0.688	0.198***	-0.001
	S2	17	6.7	5.3	5.3	0.510	0.678	0.248***	0.004
SC2	М	20	6.5	6.3	5.3	0.597	0.698	0.145***	0.005
	Sa	20	6.7	6.1	5.2	0.498	0.658	0.242***	0.003
	<b>S</b> 1	20	7.0	6.4	5.4	0.539	0.712	0.243***	0.007
	PI	60	9.5	9.1		0.543	0.694	0.218***	0.000
	S2	15	6.1	5.2	5.2	0.528	0.703	0.249***	0.010
OG3	М	22	7.6	6.7	5.5	0.629	0.695	0.095**	0.002
	Sa	20	6.7	6.1	5.2	0.544	0.688	0.209***	0.005
	<b>S</b> 1	18	6.8	6.3	5.4	0.548	0.687	0.202***	0.008
	PI	60	9.5	9.0		0.577	0.693	0.168***	0.002
	S2	11	6.0	5.6	5.6	0.555	0.699	0.206***	0.006
SC3	М	20	6.5	6.1	5.1	0.553	0.690	0.199***	0.001
	Sa	20	7.4	6.7	5.7	0.531	0.698	0.239***	0.003
	S1	20	6.8	6.1	5.2	0.452	0.683	0.337***	0.008
	PI	60	9.2	8.8		0.511	0.690	0.260***	0.001
	S2	15	5.9	5.1	5.1	0.460	0.673	0.317***	0.011
Means			9.4	9.0	5.5	0.536	0.692	0.225	0.000

N, number of individuals; A, mean number of alleles per locus;  $A_R$ , mean allelic richness;  $H_O$ , mean observed heterozygosity;  $H_E$ , mean expected heterozygosity;  $F_{IS}$ , inbreeding coefficient;  $F_{ST}$ , mean pairwise  $F_{ST}$ ; HD, P-values for heterozygote deficit; and HE, P-values for heterozygote excess. Means were determined using PI except for  $A_R^*$ . \*\*\*  $P \le 0.001$ ; \*\*  $0.001 < P \le 0.010$  for tests of heterozygote deficiency.

Observed heterozygosity ( $H_0$ ) and the inbreeding coefficient ( $F_{IS}$ ) ranged from 0.440 (saplings in OG1) to 0.629 (mature trees in OG3), and from 0.095 (mature trees in OG3) to 0.337 (seedlings in SC3), respectively (Table 5.1). All cohorts of all

populations had a significant heterozygote deficit (all P < 0.0054), suggesting a departure from random mating in both forest types (Table 5.1). We found a significant difference between cohorts for both forest types with higher mean  $H_0$  and lower mean  $F_{IS}$  for mature trees ( $H_0 = 0.562$ ;  $F_{IS} = 0.144$ ) compared to saplings ( $H_0 = 0.470$ ;  $F_{IS} = 0.269$ ), S1 ( $H_0 = 0.482$ ;  $F_{IS} = 0.260$ ), and S2 ( $H_0 = 0.472$ ;  $F_{IS} = 0.262$ ) (Table A5.7 and Fig. 5.2). There was also a significant forest type × pair interaction for  $H_0$  for S2 and a marginally significant difference for seedlings (Fig. 5.2). Similar results were obtained using FSTAT (Table A5.6).

Means populations  $F_{ST}$  were very low (-0.002–0.002; Table 5.1). Only 3 pairwise  $F_{ST}$  were significant (OG3-SC1, OG3-SC3, and SC1-SC3), but their values were very low (0.004, 0.004 and 0.002, respectively; Table A5.5).

### 5.5.3. Partitioning of molecular variation

AMOVA revealed that most of the variation (99–100 %, P = 0.015-0.176; Table 5.2) was attributable to genetic variation within populations (stands or cohorts). A slight effect of cohorts (1 %, P = 0.034; Table 5.2) was detected in the SC.

## 5.5.4. Allele frequencies

No difference for allele frequencies was detected between logged versus unlogged stands, and between cohorts (PI, M, Sa, S1 or S2) (Table A5.8; Table 5.3). We observed a lower number of alleles in the rare (f < 0.01) frequency class only for the S2 cohort. This reduction was around 10 % compared to other cohorts (Table 5.3). Reduction in the percentage of rare alleles could be due to lower sampling size of the S2 cohort; however, similar a number of alleles (total) was found in S2 and Sa (192 vs 198 - All), S2 and Sa (161 vs 165 - OG), and in S2 and M (162 vs 167 - SC) (Table 5.3).



**Figure 5.2** Predicted means (95% confidence intervals) of  $H_0$  for (a), (b), (d) and (e),  $F_{IS}$  for (c) and (f), cohorts (M: mature sugar maple, Sa: sugar maple sapling, and S1: sugar maple seedling and S2: post-harvest sugar maple seedling) for (a), (c), (d) and (f), forest types (OG: old-growth and SC: selection cutting) and pairs (1, 2, and 3) referred as sites names for (b) and (e), in Québec. (a), (b) and (c) used M, Sa and S1 data, and (d), (e) and (f) used M, Sa and S2 data.  $H_0$ , mean observed heterozygosity;  $F_{IS}$ , inbreeding coefficient. Means with the same letter do not differ at  $\alpha = 0.05$ , but differ (with an asterisk) at  $\alpha = 0.10$ .

Source of variation	d.f.	Sum of	Variance	Percentage	Phi (Φ)	P-values
		squares	components	of variance	statistics	
Between pairs	2	36.397	0.000	0	0.000	0.620
Between forest types within pairs	3	57.125	0.042	0	0.003	0.070
Within stands	354	5858.567	16.550	100	0.002	0.062
Total	359	5952.089	16.591	100		
OG						
Between pairs	2	34.050	0.000	0	0.000	0.476
Between cohorts within pairs	6	101.998	0.047	0	0.003	0.205
Within cohorts	171	2746.652	16.062	100	0.003	0.176
Total	179	2882.700	16.110	100		
SC						
Between pairs	2	39.261	0.006	0	0.000	0.416
Between cohorts within pairs	6	115.517	0.116	1	0.007	0.034
Within cohorts	171	2894.400	16.926	99	0.007	0.015
Total	179	3049.178	17.049	100		

Table 5.2 Results of analysis of molecular variance (AMOVA) showing the partitioning of genetic variance among pairs, forest types and cohorts.

d.f., degrees of freedom. Significant values at  $\alpha = 0.05$  are in bold type and at  $\alpha = 0.10$  in italics.

Populations	Cohorts	$A_{\mathrm{T}}$	С	Ι	L0.01	R0.01	L0.05	R0.05
OG	М	179	0 (0%)	22 (12%)	116 (65%)	41 (23%)	55 (31%)	102 (57%)
	Sa	165	1 (1%)	21 (13%)	103 (62%)	40 (24%)	57 (35%)	86 (52%)
	<b>S</b> 1	174	1 (1%)	20 (11%)	121 (70%)	32 (18%)	53 (30%)	100 (57%)
	PI	217	0 (0%)	21 (10%)	128 (59%)	68 (31%)	53 (24%)	143 (66%)
	S2	161	1 (1%)	20 (12%)	140 (87%)	0 (0%)	53 (33%)	87 (54%)
SC	М	167	0 (0%)	24 (14%)	110 (66%)	33 (20%)	53 (32%)	90 (54%)
	Sa	171	1 (1%)	18 (11%)	116 (68%)	36 (21%)	60 (35%)	92 (54%)
	<b>S</b> 1	175	0 (0%)	20 (11%)	118 (67%)	37 (21%)	57 (33%)	98 (56%)
	PI	215	1 (0%)	18 (8%)	127 (59%)	69 (32%)	61 (28%)	135 (63%)
	S2	162	0 (0%)	20 (12%)	142 (88%)	0 (0%)	55 (34%)	87 (54%)
All	М	202	0 (0%)	21 (10%)	126 (62%)	55 (27%)	56 (28%)	125 (62%)
	Sa	198	1 (1%)	20 (10%)	121 (61%)	56 (28%)	55 (28%)	122 (62%)
	S1	207	0 (0%)	21 (10%)	125 (60%)	61 (29%)	52 (25%)	134 (65%)
	PI	243	0 (0%)	19 (8%)	57 (23%)	167 (69%)	129 (53%)	95 (39%)
	S2	192	1 (1%)	19 (10%)	138 (72%)	34 (18%)	55 (29%)	117 (61%)

Table 5.3 Number of alleles per classes of frequency, per cohorts of the populations and grouping populations by forest types.

 $A_{\rm T}$ , total number of alleles; C, common  $f \ge 0.75$ ; I, intermediate  $0.75 > f \ge 0.25$ ; L0.01, low  $0.25 > f \ge 0.01$ ; R0.01, rare f < 0.01; L0.05, low  $0.25 > f \ge 0.05$ ; R0.05, rare f < 0.05; M, mature sugar maple; Sa, sugar maple sapling; S1, sugar maple seedling; PI, pooled individuals; S2, post-harvest sugar maple seedling.

#### 5.5.5. Bottlenecks

Recent bottlenecks were detected in the three SC stands while none were apparent in the OG stands using IAM (Table 5.4). However, no bottleneck was detected in all stands using SMM, TMM, and mode-shift models.

Heterozygosity excess												
Sites	IAM		TN	SMM	shift							
		70%	90%									
OG1	0.08368	0.97003	0.99961	0.99994	1.00000	1.00000	Normal					
SC1	0.04488	0.98075	0.99671	0.99883	0.99968	0.99974	Normal					
OG2	0.06487	0.98658	0.99968	0.99992	0.99999	0.99999	Normal					
SC2	0.04071	0.99552	0.99983	0.99995	0.99995	0.99997	Normal					
OG3	0.05935	0.97842	0.99480	0.99832	0.99903	0.99961	Normal					
SC3	0.00140	0.99800	0.99995	0.99999	1.00000	1.00000	Normal					

Table 5.4 Bottlenecks results based on heterozygosity excess and mode shift.

Significant values ( $\alpha = 0.05$ ) are in bold type.

#### 5.6. Discussion

#### 5.6.1. Genetic diversity in sugar maple populations

We observed a high level of genetic diversity in sugar maple. Previous studies (using allozymes) have also shown high levels of genetic diversity in this species (Table 5.5; A = 2.3, range: 1.10–5.00;  $H_E = 0.131$ , range: 0.015–0.275), which is typical of other *Acer* species and tree species in north-eastern North America (Table 5.5). Similar levels of diversity were reported for Norway maple (*Acer platanoides* L.), bigleaf maple (*A. macrophyllum* Pursh), American chestnut (*Castanea dentata* [Marsh.] Borkh.) and red oak (*Quercus rubra* L.), while lower levels have been observed in sycamore maple (*A. pseudoplatanus* L.) (Table 5.5). We previously found slightly lower values for *A* and  $A_R$  using the same 18 microsatellites (Table 5.5). In the present study, the number of individuals that were sampled per site was higher (N = 60 vs N = 40) and diversity was found principally within stands in both studies (Table 5.2; see Table A4.7 in Chapter IV). The values for *A*,  $A_R$  and  $H_E$  were
higher or similar to those reported for other *Acer* species and many other tree species in north-eastern North America (Table 5.5). However,  $A_R$  and  $H_E$  values were lower than those reported for mono maple (*A. mono* Maxim), northern pin oak (*Quercus ellipsoidallis* E.J.Hill) and red oak (Table 5.5).

We observed a positive and significant inbreeding coefficient ( $F_{IS} = 0.225$ , range: 0.168–0.260). The presence of heterozygote deficiency is common and has also been reported for Acer species and trembling aspen (Populus tremuloides Michaux) using microsatellite markers (Table 5.5). Heterozygote deficiency is often associated with inbreeding (high levels of consanguineous mating) or a Wahlund effect (mixing of differentiated gene pools). However, sugar maple is recognized for its low selfcompatibility, which is related to dichogamy (Gabriel, 1968), and its frequent postfertilization ovule abortion after self-pollination (Gabriel & Garrett, 1967). In the present study, sugar maple stands were localized compared to the species range and  $F_{\rm IS}$  values could vary widely between populations. We previously reported, in Chapter IV, large variation in  $F_{IS}$  (-0.051–0.302; Table 5.5) for sugar maple populations that were sampled over a wide range in Québec. Namroud et al. (2005) also detected highly significant heterozygote deficiency in two trembling aspen populations in Québec, while Callahan et al. (2013) reported significant variation of  $F_{\rm IS}$  (heterozygote excess to deficiency) in populations of this species across its natural geographic range (Table 5.5). In Korean maple (Acer takesimense), Takayama et al. (2013) observed high significant heterozygote deficiency in four populations, while others (3 populations) were at equilibrium. Hence, our  $F_{IS}$  values possibly reflected the level of genetic diversity for sugar maple in the study area.

 Table 5.5 Comparison of genetic diversity using allozyme and microsatellite markers for populations of tree species in northeastern North America, and Acer species around the world.

Species	A	A <sub>R</sub>	H <sub>0</sub>	H <sub>E</sub>	F <sub>IS</sub>	Reference
Allozymes						
Angiosperm	l					
Acer platanoides	1.92 (1.50-2.50)	—	0.129 (0.066-0.238)	0.128 (0.063-0.207)	-0.012 (-0.373-0.189)	Rusanen et al. (2000)
Acer platanoides	2.0 (1.6–2.4)	—	0.126 (0.038-0.195)	0.132 (0.053-0.191)	0.066 (-0.085–0.285)	Rusanen et al. (2003)
Acer pseudoplatanus	2.78 (2.56-3.00)	—	0.293 (0.237-0.327)	0.238 (0.254-0.319)	-0.032 (-0.159–0.085)	Belletti et al. (2007)
Acer macrophyllum	1.71 (1.5–2.2)		0.118 (0.102-0.160)	0.152 (0.102-0.189)	0.166 (-0.086–0.332)	Iddrisu & Ritland (2004)
Acer saccharum	1.95 (1.64–2.18)	—	_	0.110 (0.098-0.132)	-	Perry & Knowles (1989)
Acer saccharum	2.2 (1.1–2.8)	—	0.169 (0.08–0.28)	0.171 (0.08–0.29)	—	Foré et al. (1992b)
Acer saccharum	2.9	—	0.15	0.148	—	Foré et al. (1992a)
Acer saccharum	2.07 (2.03-2.10)	—		0.115 (0.109-0.121)	0.062 (0.050-0.073)	Young et al. (1993a)
Acer saccharum	1.98 (1.78–2.41)	—		0.112 (0.088-0.138)	0.042 (-0.095-0.177)	Young et al. (1993b)
Acer saccharum	1.83	—	0.136	0.148	0.077	Simon <i>et al.</i> (1995)
Acer saccharum	3.46 (2.00-5.00)	1.99 (1.14–2.98)	0.113 (0.021-0.294)	0.116 (0.015-0.275)	0.025 (-0.108-0.073)	Baucom et al. (2005)
Castanea dentata	1.69 (1.50–1.89)	—	0.184 (0.135-0.264)	0.151 (0.096-0.196)	-0.226	Huang et al. (1998)
Fagus grandifolia	2.9 (2.9–2.9)		0.387 (0.382-0.392)	0.395 (0.383-0.407)	0.024	Houston & Houston (1994)
Fagus grandifolia	3.0 (2.78–3.33)	—	0.163 (0.150-0.175)	0.165 (0.150-0.179)	—	Houston & Houston (2000)
Populus tremuloides	2.7 (2.1–2.9)		0.125 (0.101-0.160)	0.235 (0.207-0.270)	0.462 (0.295–0.568)	Hyun et al. (1987)
Populus tremuloides	2.6 (2.2–2.9)	—	0.217 (0.197-0.242)	0.220 (0.193-0.244)	0.017	Lund et al. (1992)
Quercus rubra	2.08 (1.8–2.3)	—		0.186 (0.145-0.245)	0.100	Sork et al. (1993)
Gymnosperm	l					
Picea glauca	3.03 (2.17-3.83)	2.14 (1.86–2.37)	0.342 (0.221-0.414)	0.344 (0.199-0.412)	0.002 (-0.092–0.087)	O'Connell et al. (2006)
Picea rubens	1.47 (1.25–1.64)	—	0.075 (0.059-0.092)	0.079 (0.061-0.104)	0.043 (-0.037-0.224)	Hawley & DeHayes (1994)
Pinus strobus	2.02 (1.69–2.37)	—	0.129 (0.121-0.143)	0.152 (0.146-0.157)	-	Buchert et al. (1997)
Pinus strobus	2.35 (2.23-2.50)	—	0.215 (0.185-0.216)	0.195 (0.181-0.216)	-0.090 (-0.200-0.053)	Rajora et al. (1998)
Thuja occidentalis	1.6 (1.5–1.8)	_	0.116 (0.102-0.133)	0.129 (0.113-0.141)	0.106	Lamy et al. (1999)
Microsatellites						
Angiosperm	l					
Acer mono	12.63	—		0.802	-0.008	Kikuchi et al., (2009)
Acer mono	—	8.37 (7.38–9.65)		0.80 (0.70–0.85)	0.27 (0.20-0.32)	Takayama <i>et al.</i> , (2012)
Acer okamotoanum	_	6.60 (6.11–7.41)	—	0.72 (0.66–0.76)	0.18 (0.03-0.24)	Takayama et al., (2012)
Acer pseudoplatanus	_		0.548 (0.543-0.553)	0.574 (0.573-0.574)		Pandey (2005)
Acer pseudosieboldianum	· —	4.60 (3.79–5.25)	0.40 (0.32–0.46)	0.61 (0.53–0.68)	0.33 (0.21–0.43)	Takayama et al. (2013)

 Table 5.5 (to continued)

Species	A	$A_{\rm R}$	$H_0$	$H_{\rm E}$	F <sub>IS</sub>	Reference
Acer saccharum	9.4 (9.2–9.5)	9.0 (8.8–9.1)	0.536 (0.507-0.577)	0.692 (0.686-0.699)	0.225 (0.168-0.260)	Our study
Acer saccharum	8.2 (6.6–9.0)	7.0 (5.8–7.6)	0.597 (0.496-0.716)	0.693 (0.637-0.715)	0.138 (-0.051-0.302)	Chapter IV of this thesis
Acer skutchii	2.1 (1.5-2.5)		—	0.129 (0.054-0.247)	—	Lara-Gomez et al. (2005)
Acer takesimense		3.82 (3.59-4.23)	0.38 (0.30-0.47)	0.53 (0.48-0.58)	0.28 (0.08-0.47)	Takayama et al. (2013)
Quercus ellipsoidallis	13	—	0.67 (0.62-0.72)	0.79 (0.77–0.81)	0.145 (0.10-0.19)	Lind & Gailing (2013)
Quercus rubra	14.5 (13–15)	_	0.73 (0.70-0.75)	0.84 (0.83-0.86)	0.12 (0.07-0.17)	Lind & Gailing (2013)
Populus tremuloides	8.83 (7.58-10.08)	—	0.465 (0.45-0.48)	0.67 (0.61-0.73)	0.30 (0.21-0.39)	Namroud et al. (2005)
Populus tremuloides		5.99 (3.34-6.83)		0.758 (0.613-0.801)	0.019 (-0.12-0.19)	Callahan et al. (2013)
Populus tremuloides	7.44 (6.25-8.2()	—	0.556 (0.478-0.704)	0.725 (0.691-0.767)	0.201 (-0.054–0.325)	Wyman <i>et al.</i> (2003)
Gymnosperm						
Pinus strobus	8.21 (6.85–9.62)		0.516 (0.485-0.538)	0.597 (0.585-0.615)	—	Rajora et al. (2000)
Pinus strobus	7.7	7.0 (6.7–7.3)	0.465 (0.46-0.47)	0.485 (0.48-0.49)	0.03 (0.01-0.05)	Marquardt & Epperson (2004)
Thuja occidentalis	9.58 (7.83-11.17)	9.21 (7.66-10.68)	0.590 (0.505-0.640)	0.600 (0.519-0.662)	0.019 (-0.025-0.050)	Pandey & Rajora (2012a)
Thuja occidentalis	7.8 (5.0–10.0)	5.9 (4.6-6.9)	0.734 (0.463-0.883)	0.773 (0.712-0.840)	0.145	Xu et al. (2012)

A, mean number of alleles per locus;  $A_{\rm R}$ , mean allelic richness;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm IS}$ , inbreeding coefficient. Range values were in parenthesis. "—" data not available.

# 5.6.2. Cohort differences

A lower observed heterozygosity  $(H_0)$  and higher  $F_{IS}$  was observed in the younger cohorts (Sa, S1 and S2) compared to mature sugar maple trees (Fig. 5.2). Contrary to our expectations, the level of genetic structure between cohorts was similar among forest types. The level of heterozygosity increased with age (here, in trees with d.b.h.  $\geq$  10 cm) in the SC and the OG stands, a result that is similar to other studies that have compared cohorts of different ages. This pattern has been reported in various coniferous tree species (Bush & Smouse, 1992; Nijensohn et al., 2005). Wang et al. (2010) further demonstrated that the number of alleles (A) and  $H_E$  were higher in mature trees (> 300-years-old) than in seedlings (2- to 3-years-old) in an intact old-growth forest of dragon spruce (Picea asperata Mast.). Ballal (1994) reported lower heterozygosity and higher  $F_{IS}$  in sugar maple embryos than in seedlings and mature trees ( $\geq$  30 cm d.b.h.). However, this pattern is not always observed. For example, Foré et al. (1992a) found no difference between embryos and older cohorts (1-year, d.b.h.  $\leq 2$  cm, 15–25 cm and  $\geq 40$  cm). Similarly, we found no difference between saplings (1–10 cm d.b.h.) and mature trees ( $\geq$  10 cm d.b.h.) in sugar maple populations in Québec (see Chapter IV).

One explanation for the presence of higher level of heterozygosity in mature trees is selection against homozygotes that occurred during the self-thinning process. This may be explained either by higher/lower fitness of heterozygote/homozygote individuals, respectively (Charlesworth & Willis, 2009). Sugar maple typically forms uneven-aged stands and the mortality rate at the seedling stage is very high, particularly for younger seedlings (Graignic *et al.*, 2014). In theory, we could expect a decrease in the low-frequency allele class with age because during selection, rare deleterious alleles would be eliminated (Charlesworth & Willis, 2009). In fact, the percentage of rare alleles was similar among mature sugar maple, saplings and A5.1, and slightly lower in A5.2 (Table 5.3).

Another possibility is that mature sugar maple trees originated from overlapping generations, given that sugar maple can live from 300 to 400 years (Godman *et al.*, 1990). Seedlings (S2 cohort) originated from a maximum of four seed masts (Graignic *et al.*, 2014). The d.b.h. of mature sugar maple trees that were sampled ranged between 10 to 82 cm, which corresponds to ages between 35 and 285 years (Majcen *et al.*, 1984; Graignic *et al.*, 2014). Those generations could have been influenced by different random selection processes at their establishment and throughout their entire lifespans.

### 5.6.3. Selection cutting influences and implications

We detected a significant deviation from mutation-drift equilibrium under IAM in the three harvested stands, but not in the three old-growth stands (Table 5.4). However, IAM is very sensitive to violations of assumptions about mutation models and may not be the most appropriate model for microsatellite markers (Luikart & Cornuet, 1998). In addition, bottlenecks were not verified by the other models (Table 5.4).

Recent bottlenecks, which resulted from logging, are typically accompanied by a reduction in the mean number of alleles, the number of low frequency rare alleles, and allelic richness (Pautasso, 2009). In some cases, a reduction in the level of heterozygosity is observed (Rajora, 1999) because allelic diversity is reduced more rapidly than heterozygosity under bottlenecks (Nei *et al.*, 1975; Spencer *et al.*, 2000). We found no difference in  $A_{\rm R}$ ,  $H_{\rm E}$ ,  $H_{\rm O}$ , and  $F_{\rm IS}$  between OG and SC stands (Fig. 5.2). A very low level of differentiation between cohorts in SC stands was detected, while there was no differentiation in OG stands (Table 5.2). Thus, it appears that SC had a low negative effect on the genetic diversity of the remaining mature sugar maple trees.

The negative impact of SC on sugar maple stands was very weak and could be transient because we found (1) a high level of genetic diversity in sugar maple stands, (2) low genetic differentiation between stands ( $F_{ST} \leq 0.004$ ) (Table 5.1 and Table A5.5), and (3) genetic diversity resided mostly within-population, as revealed by AMOVA (Table 5.2). Low  $F_{ST}$  values between stands that were separated by 80 km indicated high levels of gene flow. Long-distance effective wind-dispersal of pollen is reported for trees, e.g. up to 100 km for Scots pine (*Pinus sylvestris* L., Kremer *et al.*, 2012). Sugar maple is a wind-pollinated species (Gabriel & Garrett, 1984) and its seeds are dispersed to a maximum distance of 100 m (Johnson, 1988). Therefore, pollen gene flow is very important in sugar maple populations. Low levels of genetic differentiation have been reported for sugar maple stands in an adjacent region ( $F_{ST} \leq 0.017$ ) and over a broader regional scale ( $F_{ST} \leq 0.049$ ; Ontario to Nova Scotia, Canada; Young *et al.*, 1993a; Young *et al.*, 1993b).

### 5.7. Conclusions

In conclusion, our results indicated that the level of gene flow in OG and SC stands is sufficient to maintain the level of genetic diversity for future generations. However, we also found a small negative influence of SC on mature sugar maple diversity, which may create a genetic bottleneck. Harvesting on our sites had occurred 18 years before sampling and the next SC is planned in 15–25 years. A second harvesting in the same stands is thus possibly very soon. It could over-imposed several very weak lost of genetic diversity and possibly lead to erosion of maternal genetic diversity. After one SC harvest, we would suggest that this type of treatment, whereby forest stands are effectively managed as uneven-aged structures, was not detrimental in terms of its effects on genetic diversity. However, we could not accurately predict the effects of multiple logging. Therefore, we would recommend the monitoring of genetic diversity in sugar maple cohorts after multiple selection cutting.

# 5.8. Acknowledgements

We thank Elizabeth Turcotte who provided field assistance. We are grateful to William F. J. Parsons for careful language editing. This work was funded by an NSERC (Natural Sciences and Engineering Research Council of Canada) strategic grant (STPGP 336871) to Francine Tremblay.

#### CHAPITRE VI

#### CONCLUSION GÉNÉRALE

### 6.1. Effets des différents types de fragmentation

L'objectif de cette thèse était d'étudier la dynamique de régénération et la diversité génétique des populations nordiques, fragmentées d'érable à sucre. Il s'articule autour de l'hypothèse que ces populations pourraient être à l'origine de l'expansion vers le nord de l'érable à sucre dans un contexte de changements climatiques.

Nous avons montré que la reproduction sexuée était importante (Chapitre II), la diversité génétique était élevée (Chapitre IV) et la différenciation génétique était faible (Chapitre IV), chez les populations d'érable à sucre réparties sur l'ensemble du Québec. Le climat joue un rôle sur la reproduction de cette espèce, car des températures estivales chaudes et sèches coïncidaient avec de bonnes années semencières l'année suivante. Il joue aussi un rôle sur la survie des plantules d'érable à sucre, car un mois de juillet chaud et humide était lié à l'augmentation du nombre de plantules. Les changements climatiques anticipés devraient donc contribuer à l'augmentation du nombre de plantules d'érable à sucre au nord (Chapitre II). Tous ces éléments valident l'hypothèse selon laquelle les populations nordiques fragmentées d'érable à sucre seront en mesure de se régénérer, survivre et sont suffisamment diversifiées au niveau génétique pour être à l'origine de l'expansion vers le nord de cette espèce.

Par ailleurs, nous avons montré l'existence de différences, au niveau écologique et génétique, entre la zone de répartition discontinue fragmentée et la zone continue (Chapitre II et IV). Les différences observées permettent d'appuyer l'hypothèse d'un patron de variation génétique des populations centrales vers les populations marginales (Eckert *et al.*, 2008). Toutefois, comme nous n'avons échantillonné que des populations marginales situées au nord (et non est, ouest ou sud), nous ne sommes pas en mesure de privilégier cette hypothèse comparativement à celle d'un patron de variation génétique qui serait le résultat du déplacement de l'espèce vers le nord de son aire de répartition suite au dernier maximum glaciaire (Hampe & Petit, 2005).

Il faut toutefois noté que les différences observées n'étaient significatives que pour le transect ouest (région d'Abitibi-Témiscamingue). En effet, nous avons vu que l'érable à sucre se reproduisait moins bien au nord (zone discontinue) par rapport au sud (zone continue) en Abitibi-Témiscamingue (Chapitre II). Cette différence se situait tant au niveau du nombre de samares pleines (contenant une graine) pour la période se situant entre 2008–2011, qu'au niveau du nombre de plantules dénombrées en 2008–2009 (années semencières en 1996, 2002, 2006). Au niveau génétique (Chapitre IV), nous avons trouvé aussi une différence entre les populations du sud et du nord du gradient ouest avec une plus grande diversité génétique (richesse allélique,  $A_R$ ) et une plus faible différenciation génétique ( $F_{ST}$ ) au sud ( $A_R = 7.3$ ;  $F_{ST} = 0.004$ ) comparativement au nord ( $A_R = 6.6$ ;  $F_{ST} = 0.030$ ).

Les analyses génétiques ont mis en évidence que, pour l'érable à sucre, (1) les populations les plus nordiques (1-D-A, 1-D-B, 2-D-A et 3-D-A) étaient déterminantes pour révéler la différence entre la zone discontinue et continue, (2) la migration de l'érable à sucre au Québec aurait commencé à l'est, puis à l'ouest et (3) les populations nordiques de l'ouest sont probablement issues d'une migration d'est en ouest et non du sud vers le nord.

La coupe forestière quant à elle, n'a pas eu d'impact majeur au niveau génétique (Chapitre V). Toutefois, les populations qui ont subi des coupes de jardinage en 1990–1991, semblent toutes avoir subi un goulot d'étranglement (bottleneck) récent. Ce goulot d'étranglement nous semble intrigant, des investigations à plus longtemps terme (observations après plus de générations d'érable à sucre) nous paraissent indispensables.

### 6.2. Implications pour l'exploitation forestière et la conservation

### 6.2.1. Implications de nos travaux de recherches pour l'aménagement forestier

Nous n'avons observé aucune perte de diversité dans les érablières 18 ans après la coupe de jardinage, comparativement aux forêts anciennes. Les résultats ont révélé la présence d'un goulot d'étranglement, mais ils doivent être interprétés avec prudence puisque les seuils de significativité variaient suivant le modèle utilisé (effet significatif sous IAM et non significatif pour SMM; Luikart & Cornuet, 1998). C'est la raison pour laquelle l'exploitation de l'érable à sucre par la coupe de jardinage, telle que pratiquée actuellement, semble appropriée du point de vue de la préservation de la diversité génétique. Toutefois suivant le principe de précaution, des mesures de diversité et structure génétique devraient être faites après les coupes pratiquées sur la même zone, ceci afin de valider le peu d'effet à long terme de ce type de coupe sur la diversité génétique de l'érable à sucre.

L'exploitation par coupe de jardinage des érablières les plus nordiques serait à proscrire puisqu'elles sont moins diversifiées au niveau génétique et plus différenciées. Les populations au nord-ouest du Québec, en particulier, semblent ainsi encore plus à risques. Certaines populations nordiques, sont déjà protégées par le gouvernement du Québec, car elles sont des écosystèmes forestiers exceptionnels (EFE).

#### 6.2.2. Intérêts pour les compagnies forestières

Les problèmes environnementaux et de gestion des ressources représentent un enjeu important et suscitent un intérêt de plus en plus important de la part des consommateurs. L'achat par le consommateur de bois issu d'une bonne gestion de la ressource est possible grâce à des certifications environnementales. La plus connue en Amérique du Nord est la certification FSC (Forest Stewardship Council), créée en 1993, elle est internationale et a connu un essor ces dernières années (https://ca.fsc.org). L'industrie forestière, consciente de cet intérêt, cherche à obtenir cette certification pour le bois qu'elle exploite. Au Québec, lors de l'entrée en vigueur du nouveau régime forestier, le 1<sup>er</sup> avril 2013, près de 51 % de la forêt du domaine de l'État avait été certifiée FSC par les compagnies forestières (https://www.mrn.gouv.qc.ca/forets/amenagement/amenagement-certification.jsp). Le 27 mars 2013, le MRNQ et le Conseil de l'industrie forestière du Québec ont convenu que les sociétés forestières demeureraient titulaires de ces certificats.

Deux des principes de la certification FSC sont en lien avec nos travaux. Le principe 6 du FSC souligne l'intérêt de maintenir et conserver les valeurs environnementales. Ces valeurs environnementales comprennent l'aspect de la fonction des écosystèmes incluant les processus évolutifs comme les flux génétiques. L'évaluation du maintien des valeurs environnementales pourraient donc passer par une évaluation avant et après coupes de la diversité et de la structure génétique de la zone coupée comme nous le suggérons pour les coupes de jardinage. La norme FSC exige de l'organisation que celle-ci identifie et protège des superficies représentatives des écosystèmes natifs (nommées « aires-échantillons »). Dans ce cadre, nous proposons que le gouvernement et les compagnies s'entendent pour couvrir l'aire de répartition d'une espèce avec des aires protégées de forêts anciennes comme c'est déjà le cas pour les EFE et en leur absence, des massifs forestiers les moins perturbés possibles. La certification FSC est internationale, une concertation serait donc

souhaitable entre les provinces canadiennes et entre le Canada et les USA afin d'améliorer la couverture de protection sur l'ensemble de l'aire de répartition d'une espèce. Le principe 9 de la certification FSC demande que l'organisation préserve et/ou augmente le nombre de massifs à Hautes Valeurs de Conservation (HVC) dans l'unité de gestion en appliquant le principe de précaution. Dans ce cadre, la protection des populations nordiques d'érable à sucre et les forêts anciennes plus au sud est fondamentale.

### 6.3. Perspectives de recherche

# 6.3.1. Élargir la zone d'étude & mariage à trois (écologie–génétique–palynologie)

Les données génétiques (Chapitre IV) et palynologiques tendent à démontrer que la migration de l'érable à sucre aurait débuté par l'est du Québec (Montréal–Québec–Saguenay–Lac-St-Jean) et une deuxième voie de migration serait arrivée par l'ouest du Québec (Témiscamingue). Les données palynologiques ont montré qu'une voie migratoire de l'érable à sucre en Amérique du Nord aurait suivi les Appalaches (Braun, 1950) pour atteindre le sud-est du Québec vers 9 000 ans AA (Lavoie & Richard, 2000), alors que la présence de l'érable à sucre est détectée 4 000 ans AA à la pointe de la Gaspésie (Richard & Grondin, 2009). Il serait intéressant d'échantillonner des populations en Gaspésie et de couvrir l'extrême sud-est du Québec afin de vérifier si une signature génétique confirme une migration plus tardive en Gaspésie. Faire une étude, en parallèle, approfondie au niveau écologique permettrait de mieux comprendre pour quelle raison la migration de l'érable à sucre? D'autres espèces étaient déjà établies et rentraient en compétition avec l'érable à sucre ?

Nous n'avons pas pu départager l'hypothèse de « centre-périphérie » (Eckert *et al.*, 2008) vs la « migration post-glaciaire » (Hampe & Petit, 2005), en relation avec la variation génétique retrouvée entre les populations d'érable à sucre nordiques

marginales et les populations de la zone continue (Chapitre IV). Cette question a un intérêt majeur est reste en suspens pour de nombreuses espèces (Guo, 2012). Afin de répondre à cette question, les recherches futures devront couvrir les niveaux écologiques (capacité des populations périphériques fragmentées à se reproduire) et génétiques (populations périphériques différenciées génétiquement) vers le sud, l'est et l'ouest de l'aire de répartition de l'érable à sucre.

Nous n'avons pas pu valider l'hypothèse du retrait de l'érable à sucre ou de « saut-de-puce » à l'origine des populations marginales nordiques d'érable à sucre. Afin d'arriver à répondre à cette question, une approche intéressante consisterait à prélever des carottes de sédiments lacustres dans les zones situées à proximité des populations très nordiques afin d'établir si l'abondance de l'espèce et la taille des populations ont varié dans le temps. L'analyse du pollen et des macrorestes nous indiquerait si, par le passé, l'érable à sucre était plus abondant dans la région par rapport à maintenant et à quel moment remonte son arrivée. Si une quantité suffisamente d'échantillon pouvait être récoltée, une analyse génétique sur l'ADN ancien d'érable à sucre pourrait aussi apporter des réponses à cette question.

### 6.3.2. Quels marqueurs utiliser ? Vers la révolution du séquençage à haut débit

De nouvelles études génétiques pourront êtres faites en utilisant nos marqueurs microsatellites nucléaires (Chapitre III et IV). Les marqueurs microsatellites ont un taux d'évolution rapide, une grande diversité dans les populations, et sont distribués abondamment et aléatoirement sur le génome (Ellegren, 2004). Ils font partie des marqueurs nucléaires d'un locus avec le plus de variabilité (Sunnucks, 2000). Moins de variabilité intra-spécifique est retrouvée pour les organelles que pour l'ADN du noyau (ADNn) (Wolfe *et al.*, 1987). Malgré cela, il nous a été difficile d'obtenir des marqueurs microsatellites nucléaires polymorphes pour cette espèce (Chapitre II). Dans ce cas, les marqueurs microsatellites nucléaires nucléaires nucléaires continuent à êtres utilisés pour

les études intra-spécifiques à large échelle pour plusieurs espèces comme, par exemple, pour le peuplier faux-tremble (Callahan *et al.*, 2013).

Cependant, pour avoir une signature de la migration postglaciaire plus pertinente, il est recommandé d'utiliser d'autres types de marqueurs nucléaires et des marqueurs génétiques provenant d'organelles (chloroplaste et mitonchondrie) pouvant être mieux ordonnés dans l'évolution que des marqueurs microsatellites (Ellegren, 2004). Chez certaines espèces, une concordance des structures de l'ADN du chloroplaste (ADNcp), de la mitonchondrie (ADNmt) et du noyau a été observée, tandis que chez d'autres espèces il y a divergence (Gamache et al., 2003; Fontaine et al., 2004; Petit et al., 2005; Meng et al., 2007; Bai et al., 2010). L'ADN des organelles a un intérêt tout particulier puisqu'il est hérité d'un seul parent (contrairement à l'ADNn) et il est donc possible de séparer une dispersion par les graines et par le pollen (Ennos, 1994; Newton et al., 1999). Le chloroplaste est d'origine maternelle pour les angiospermes et paternelle pour les gymnospermes. La mitochondrie est généralement d'origine maternelle sauf pour quelques gymnospermes. Chez l'érable à sucre, l'ADNcp et l'ADNmt sont donc d'origine maternelle et reflètent la dispersion par les graines. Les amorces cytoplasmiques utilisées avec succès pour l'érable rouge et l'érable argenté (Mclachlan et al., 2005; Saeki et al., 2011), pour des pruniers, appartenant à la même classe que l'érable à sucre (Shaw & Small, 2005) et des amorces universelles (Taberlet et al., 1991; Demesure et al., 1995) pourront être testées pour amplifier l'ADN des organelles de l'érable à sucre.

Toutefois, comme ce type de marqueurs est moins variable, le séquençage à haut débit représenterait une alternative intéressante (Emerson *et al.*, 2010; Peñalba *et al.*, 2014). Cette technique a beaucoup évolué ces dernières années et elle est devenue beaucoup moins coûteuse. Nous l'avons utilisé dans le développement de nouveaux marqueurs microsatellites pour l'érable à sucre et découvrir d'autres types de marqueur est possible (Chapitre II) (Davey *et al.*, 2011). Nous proposons d'utiliser

cette technique pour séquencer tout l'ADN chloroplastique de nos 23 populations d'érable à sucre (Mariac *et al.*, 2014).

#### 6.3.3. Le séquençage à haut débit au service de la détection de gènes adaptatifs

La technique du séquençage à haut débit est très performante et facilite la détection des gènes adaptatifs (Stapley *et al.*, 2010). Par exemple, elle permet de détecter un très grand nombre de marqueurs polymorphiques pour un seul nucléotide (SNPs, single nucleotide polymorphisms; Emerson *et al.*, 2010). Ce type de marqueur est parfois situé au niveau d'un gène (loci non neutre), ce qui le rend utile pour tester une signature d'adaptation locale. Ce type de marqueur est intéressant pour l'érable à sucre, car cette espèce montre des variations clinales, signe d'adaptation aux conditions locales (Kriebel, 1957). L'érable à sucre étant un organisme non-modèle, très peu de séquences ont été obtenues jusqu'à maintenant et aucun SNP développé, ce qui en fait un candidat intéressant pour la découverte de SNPs par séquençage à haut débit (Seeb *et al.*, 2011). L'identification de SNPs associé au *fitness* de l'espèce, offrirait un bon indicateur du niveau de variation adaptative au sein des populations (Van Tienderen *et al.*, 2002; Namroud *et al.*, 2008; Keller *et al.*, 2012).

Le séquençage de nouvelle génération peut aussi être utilisé pour cribler des gènes candidats de traits phénologiques en utilisant la différence d'expression des gènes (ARN), le séquençage aléatoire global du transcrit (whole-transcriptome shotgun sequencing; Ueno *et al.*, 2013).

Les recherches qui associent les gènes et les traits sont importantes dans un contexte de changements climatiques, car il sera possible d'identifier des populations menacées (Buckley & Kingsolver, 2012). Il serait recommandé d'utiliser ces nouvelles techniques sur des populations couvrant l'ensemble de l'aire de répartition de l'érable à sucre, ou tout au moins sur les populations marginales situées au nord et au sud (Alberto *et al.*, 2013). Les populations au sud de l'aire de répartition sont-elles

adaptées à la sécheresse et celles au nord au froid (Namroud *et al.*, 2008; Eckert *et al.*, 2009; Holliday *et al.*, 2010; Keller *et al.*, 2012) ? Les nouvelles techniques nous permettraient d'avoir une vue d'ensemble rapide et peu coûteuse sur les variations génétiques de ces populations. Pour répondre à ces questions, il est important d'observer les traits de ces populations, et il est nécessaire de faire des nouveaux tests de provenances localisés en dehors de l'aire de répartition de l'espèce d'intérêt (Alberto *et al.*, 2013).

# ANNEXE A2

### SUPPORTING INFORMATION FOR CHAPTER II



Figure A2.1 Diameter at breast height (1.30 m) structure of mature sugar maple trees in all 24 sites (n = 1133) in Québec.

Study ID	Study name	Protection type	Zone	Latitude	Longitude	Elevation (m)
		Transect 1				
1-D-A	Lac Labelle	EFE	D	48° 10′ 41.42″ N	79° 27′ 21.54″ W	379
1-D-B	Lac Okiwakamik	EFE	D	47° 51′ 46.18″ N	77° 44′ 6.85″ W	349
1-D-C	Rémigny		D	47° 40′ 52.23″ N	79° 2′ 18.82″ W	332
1-D-D	Lac de la Tour		D	47° 21′ 30.12″ N	78° 40′ 48.27″ W	351
1-C-A	Lac St Amand		С	47° 13′ 2.04″ N	79° 8′ 38.83″ W	318
1-C-B	Kipawa		С	46° 45′ 0.40″ N	78° 54′ 20.16″ W	345
1-C-C	Lac Six Milles		С	46° 26′ 43.69″ N	78° 0′ 44.32″ W	348
1-C-D	Lac Percival	EFE	С	46° 20′ 57.3″ N	78° 19′ 44.36″ W	330
		Transect 2				
2-D-A	Lac Pénobscot	EFE	D	47° 35′ 55.79″ N	74° 52′ 43.39″ W	493
2-D-B	Lac aux Cèdres		D	47° 20′ 47.83″ N	75° 45′ 15.95″ W	441
2-D-C	Réservoir Mitchinamécus		D	47° 12′ 52.70″ N	75° 11′ 59.32″ W	439
2-D-D	Lac des Polonais		D	47° 2′ 22.49″ N	75° 25′ 49.44″ W	343
2-C-A	Montagne du Diable	EFE	С	46° 42′ 41.36″ N	75° 34′ 7.03″ W	401
2-С-В	Lac Ecuyer		С	46° 30′ 46.30″ N	74° 36′ 21.02″ W	455
2-C-C	Lac Marie-Lefranc	EFE	С	46° 7′ 5.34″ N	75° 0′ 36.47″ W	410
2-C-D	Lac de l'Ecluse	EFE	С	45° 51′ 35.64″ N	75° 23′ 55.75″ W	388
		Transect 3				
3-D-A	Lac Patrick		D	48° 59′ 8.66″ N	71° 52′ 20.89″ W	268
3-D-B	Fjord du Saguenay	Réserve écologique G.Oscar-Villeneuve; EFE	D	48° 27′ 46.76″ N	70° 29′ 54.38″ W	362
3-D-C	Baie Eternité	Parc du Saguenay; EFE	D	48° 16′ 39.68″ N	70° 21′ 51.55″ W	157
3-D-D	Lac Edouard	Réserve écologique Judith-de-Brésole; EFE	D	47° 36′ 18.18″ N	72° 20′ 9.56″ W	450
3-C-A	Lac Paul		С	47° 18′ 39.89″ N	72° 37′ 57.65″ W	338
3-С-В	Lac Dickey		С	47° 4′ 29.06″ N	72° 51′ 12.02″ W	374
3-C-C	Lac Grandbois	EFE	С	47° 1′ 13.98″ N	72° 7′ 9.16″ W	342
3-C-D	Lac Larose		С	46° 36′ 34.34″ N	73° 2′ 27.67″ W	330

**Table A2.1** Summary of sample site coordinates, protection types, and site and stand characteristics. EFE, 'écosystèmes forestiers exceptionnels'; D, discontinuous; C, continuous; 1AY, undifferentiated till 50–100 cm thick; 1A, undifferentiated till; 1AM, undifferentiated till 25–50 cm thick; R1A, undifferentiated till < 50 cm thick. Data age class ecotone: VIN, old uneven-aged, more than three age classes and > 80-years-old; 70, 61–80 years old; 30, 21–40 years old.

Study ID	Study name	Aspect	Slope inclination	Deposit type	Drainage	data age class SIFORT	Age* (years)	d.b.h*. (cm)
				Transect 1				
1-D-A	Lac Labelle	south	18%	1AY	moderate	VIN	123	33.8
1-D-B	Lac Okiwakamik	north	2%	1A	good	30	148	34.7
1-D-C	Rémigny	south	12%	1AM	moderate	VIN	162	45.6
1-D-D	Lac de la Tour	north	18%	1AY	moderate	VIN	182	56.2
1-C-A	Lac St Amand	east	14%	1AM	moderate	VIN	120	49.1
1-C-B	Kipawa	east	16%	1AY	good	VIN	156	47.5
1-C-C	Lac Six Milles	north	7%	1A	moderate	VIN	114	32.5
1-C-D	Lac Percival	north	23%	1AM	good	VIN	147	54.5
				Transect 2				
2-D-A	Lac Pénobscot	south	18%	1AM	good	VIN	153	38.6
2-D-B	Lac aux Cèdres	north	14%	1AM	good	VIN	140	46.2
2-D-C	Réservoir Mitchinamécus	west	19%	1AM	moderate	VIN	167	50.4
2-D-D	Lac des Polonais	west	19%	1AM	moderate	VIN	173	40.8
2-C-A	Montagne du Diable	north	32%	1AY	good	VIN	161	49.7
2-С-В	Lac Ecuyer	south	12%	1AY	good	VIN	136	43.5
2-C-C	Lac Marie-Lefranc	south	5%	1AY	moderate	VIN	207	61.3
2-C-D	Lac de l'Ecluse	south	9%	1AM	good	VIN	155	51.5
				Transect 3				
3-D-A	Lac Patrick	south	9%	1AY	good	70	96	30.4
3-D-B	Fjord du Saguenay	south	47%	R1A	moderate	VIN	132	46.0
3-D-C	Baie Eternité	south	23%	1AY	moderate	VIN	222	55.9
3-D-D	Lac Edouard	total	0%	1AY	moderate	VIN	71	33.9
3-C-A	Lac Paul	north	11%	1AY	moderate	VIN	110	50.3
3-С-В	Lac Dickey	east	9%	1AM	good	VIN	139	44.5
3-C-C	Lac Grandbois	west	51%	1AM	good	VIN	99	42.0
3-C-D	Lac Larose	east	18%	1A	good	VIN	185	59.2

 Table A2.1 (to continued)

\*Age or d.b.h., mean estimated of five largest trees

Transect		1		2		3
Zone	D	С	D	С	D	С
Mean annual temperature (°C)	1.6	4.1	2.2	3.3	1.4	3.1
Growing degree days above 5 °C	1405	1702	1445	1575	1341	1555
Mean temperature of coldest month (January, °C)	-16.9	-13.2	-14.9	-14.2	-17.0	-14.7
Mean temperature of warmest month (July, °C)	17.3	19.2	16.9	18.3	16.8	18.3
Total annual precipitation (mm)	927	929	1139	1062	1079	1029
Snow fall	28%	25%	31%	24%	32%	22%

Table A2.2 Climatic data for the study area in the D (discontinuous) and C (continuous) zones along the three transects located in Québec, Canada.

			Negative e	xponential	Power fu	nction
Transect	Zone	Site	ΔAIC <sub>c</sub>	ω	ΔAIC <sub>c</sub>	ω
1	discontinuous	1-D-A'	0.0	1.00	20.6	0.00
		1-D-B'	27.9	0.00	0.0	1.00
		1-D-C'	49.6	0.00	0.0	1.00
		1-D-D'	48.8	0.00	0.0	1.00
	continuous	1-C-A	2.7	0.21	0.0	0.79
		1-C-B	17.0	0.00	0.0	1.00
		1-C-C	16.2	0.00	0.0	1.00
		1-C-D	1.1	0.37	0.0	0.63
2	discontinuous	2-D-A'	46.1	0.00	0.0	1.00
		2-D-B'	35.0	0.00	0.0	1.00
		2-D-C'	6.5	0.04	0.0	0.96
		2-D-D'	14.2	0.00	0.0	1.00
	continuous	2-C-A	43.6	0.00	0.0	1.00
		2-С-В	66.8	0.00	0.0	1.00
		2-C-C	31.0	0.00	0.0	1.00
		2-C-D	41.5	0.00	0.0	1.00
3	discontinuous	3-D-A'	9.7	0.01	0.0	0.99
		3-D-B'	24.2	0.00	0.0	1.00
		3-D-C'	102.7	0.00	0.0	1.00
		3-D-D'	23.2	0.00	0.0	1.00
	continuous	3-C-A	0.0	0.99	13.4	0.00
		3-С-В	0.0	0.81	2.9	0.19
		3-C-C	0.0	0.63	1.1	0.37
		3-C-D	0.0	0.98	8.1	0.02

**Table A2.3** Difference in Akaike information criterion ( $\Delta AIC_c$ ) and weights ( $\omega$ ) of age structure models (negative exponential and power function) for sugar maple seedling 5-year age-class frequency distributions in all sites (n = 2880) located in Québec.

Response variable	Explanatory variables	Num. d.f.	Den. d.f.	F	Р
Total mature tree density	zone	1	18	8.9691	0.0078
	transect	2	18	1.4972	0.2503
	$zone \times transect$	2	18	0.9381	0.4097
Mature sugar maple density	zone	1	18	0.1714	0.6837
	transect	2	18	3.9380	0.0381
	zone × transect	2	18	1.2546	0.3090
Total sapling density	zone	1	6	0.3446	0.5786
	transect	2	99	5.8395	0.0040
	$zone \times transect$	2	99	1.4791	0.2328
Sugar maple sapling density	zone	1	6	0.4230	0.5395
	transect	2	99	4.9508	0.0089
	zone × transect	2	99	3.1813	0.0458
Total mature tree basal area	zone	1	18	0.0440	0.8363
	transect	2	18	1.1911	0.3267
	zone × transect	2	18	0.3520	0.7080
Mature sugar maple basal area	zone	1	18	2.3250	0.1447
	transect	2	18	1.3391	0.2870
	zone × transect	2	18	0.9355	0.4106
Total sapling basal area	zone	1	6	0.0026	0.9609
	transect	2	99	2.9039	0.0595
	zone × transect	2	99	0.2113	0.8099
Sugar maple sapling basal area	zone	1	6	0.3875	0.5565
	transect	2	99	1.1968	0.3067
	zone × transect	2	99	0.2708	0.7633

**Table A2.4** Results of two-way ANOVA with interaction for mature trees (n = 24) and linear mixed-effects models for saplings (n = 111) in Québec, prior to model simplification. Num. d.f., numerator degrees of freedom; den. d.f., denominator degrees of freedom.



**Figure A2.2** Sugar maple age structures of all 24 sites (n = 2880) in Québec. For each population, the linearized form of the power function is included, together with its coefficient of determination ( $r^2$ ).  $y_i = r - M \times x_i$ , where  $y_i$  is ln(stems m<sup>-2</sup> + 1), r is intercept of the curve [ln(R), where R is initial recruitment], M is slope of the curve (mortality rate), and  $x_i$  is ln(age).



**Figure A2.3** Predicted means ( $\pm$  95% confidence intervals) of sugar maple seedling mortality rate (*n* = 24) in Québec. Means with the same letter do not differ at  $\alpha$  = 0.05, but differ (with an asterisk) at  $\alpha$  = 0.10. C, continuous zone; D, discontinuous zone.

**Table A2.5** Difference in Akaike information criterion ( $\Delta AIC_c$ ), weights ( $\omega$ ) and number of estimated parameters (*K*) for model comparisons of the relative importance of stand characteristics and climate conditions as predictors of sugar maple seedling density (n = 353) in Québec. For brevity, only models with a cumulative  $\omega \le 0.95$  were shown.

Model	K	$\Delta AIC_{c}$	ω
m_BA + m_ers_BA + s_BA + s_ers_BA + July_T + July_P	9	0.00	0.79
$m_d + m_{ers_d} + s_{ers_d} + July_T + July_P$	8	5.66	0.05
$m_BA + m_ers_BA + s_BA + s_ers_BA + Oct_T + Oct_P$	9	6.40	0.03
$m_BA + m_ers_BA + s_BA + s_ers_BA + May_T + May_P$	9	6.65	0.03
m_BA + m_ers_BA + s_BA + s_ers_BA + April_T + April_P	9	7.19	0.02
$m_BA + m_ers_BA + s_BA + s_ers_BA$	7	7.53	0.02

Explanatory variable abbreviations: m\_BA, mature tree basal area; m\_ers\_BA, mature sugar maple basal area; s\_BA, sapling basal area; s\_ers\_BA, sugar maple sapling basal area; m\_d, mature tree density; m\_ers\_d, mature sugar maple basal area; s\_ers\_d, sugar maple sapling basal area; July\_T, July mean temperature; July\_P, July mean precipitation; Oct\_T, October mean temperature; Oct\_P, October mean precipitation; May\_T, May mean temperature; May\_P, May mean precipitation; April\_T, April mean temperature; April P, April mean precipitation.

# ANNEXE A4

#### SUPPORTING INFORMATION FOR CHAPTER IV

**Table A4.1** Genetic variability estimates of microsatellite markers used in the study of sugar maple (*Acer saccharum* Marsh.) in Québec.

Locus	GenBank	М	Size range	$A_{\mathrm{T}}$	A	$H_0$	$H_{\rm E}$	$F_{\rm IS}$	References
	access no.		(bp)						
SM11	KC731552	3	178-200	13	6.3	0.575	0.629	0.083	(Graignic et al., 2013)
SM14	KC751436	4	70-120	22	14.3	0.773	0.888	0.130	(Graignic et al., 2013)
SM21A	KC751437	4	173-243	31	14.8	0.780	0.866	0.096	(Graignic et al., 2013)
SM22	KC751438	2	293-325	19	12.6	0.658	0.890	0.248	(Graignic et al., 2013)
SM27	KC751440	4	242-260	9	6.4	0.548	0.722	0.227	(Graignic et al., 2013)
SM29	KC751441	4	272-307	12	6.4	0.598	0.724	0.164	(Graignic et al., 2013)
SM34	KC751442	3	118-171	26	13.7	0.819	0.854	0.041	(Graignic et al., 2013)
SM36	KC751443	5	146–196	21	11.3	0.795	0.844	0.057	(Graignic et al., 2013)
SM37	KC751444	2	174-200	14	7.9	0.633	0.669	0.061	(Graignic et al., 2013)
SM42	KC751445	1	90-135	20	7.9	0.726	0.789	0.080	(Graignic et al., 2013)
SM47	KC751446	2	201-225	11	6.2	0.428	0.634	0.328	(Graignic et al., 2013)
SM51	KC751447	1	269–290	9	4.5	0.395	0.482	0.172	(Graignic et al., 2013)
SM53	KC751448	5	287-310	10	3.1	0.435	0.520	0.156	(Graignic et al., 2013)
SM55	KC751449	2	246-276	14	9.1	0.567	0.678	0.166	(Graignic et al., 2013)
SM56	KC751450	3	287–299	6	4.7	0.433	0.618	0.278	(Graignic et al., 2013)
SM60	KC751452	3	231-243	4	2.8	0.393	0.457	0.137	(Graignic et al., 2013)
Aop943	EF531298	1	143-158	6	4.6	0.529	0.527	-0.005	(Segarra-Moragues <i>et al.</i> , 2008)
									(Graignic et al., 2013)
Am116	AB303350	1	230-273	21	10.3	0.659	0.682	0.031	(Kikuchi and Shibata 2008)
									(Graignic et al., 2013)

M, multiplexing arrangement;  $A_{\rm T}$ , total number of alleles; A, mean number of alleles;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm IS}$ , inbreeding coefficient.

	IDA	1DB	1DC	IDD	1 CA	1CB	ICC	1CD	2DA	2DB	2DC	2DD	2CA	2CB	2CC	2CD	3DA	3DB	3DC	3DD	3CA	3CB	3CC	3CD
SM11	0.0084	0.4624	0.5799	0.0734	0.0534	0.0360	0.3199	0.0337	0.2573	Х	0.0423	0.0453	0.0395	0.0793	0.2054	0.5682	0.0083	0.9193	0.3279	0.5603	0.0012	0.3101	0.2052	0.5629
SM14	0.1373	0.2148	0.0197	0.0768	0.0162	0.1430	0.0700	0.0014	0.2848	Х	0.0018	0.6783	0.2159	0.2097	0.1402	0.3979	0.0040	0.0028	0.2548	0.5538	0.2092	0.2191	0.2174	0.3110
SM21A	0.0501	0.0786	0.3479	0.4064	0.0142	0.4691	0.7796	0.0002	0.0143	Х	0.0373	0.3263	0.0331	0.7113	0.7279	0.0666	0.0342	0.5222	0.9107	0.5043	0.0000	0.1551	0.0316	0.0098
SM22*	0.0000	0.0173	0.2494	0.0024	0.0591	0.0362	0.0001	0.0000	0.0240	Х	0.0010	0.0000	0.0003	0.0000	0.0002	0.0000	0.0000	0.2187	0.1322	0.5501	0.0062	0.0426	0.0000	0.5493
SM27*	0.0041	0.0061	0.1534	0.0186	0.0552	0.7882	0.7882	0.0000	0.0001	Х	0.0000	0.0007	0.0193	0.0001	0.2009	0.0003	0.0000	0.1258	0.9890	0.7288	0.6601	0.6601	0.0000	0.7657
SM29	0.3152	0.1248	0.5151	0.1278	0.4147	0.8772	0.2453	0.0099	0.0099	Х	0.0011	0.0000	0.1115	0.0002	0.0209	0.0078	0.0001	0.2102	0.5353	0.0892	0.1776	0.3961	0.0000	0.5572
SM34	0.3955	0.4077	0.5409	0.5499	0.1266	0.6818	0.6818	0.4333	0.0772	Х	0.4617	0.0957	0.7941	0.1372	0.1517	0.3600	0.5645	0.4699	0.0625	0.2237	0.0851	0.0851	0.4614	0.2222
SM36	0.6068	0.0593	0.2524	0.3244	0.1953	0.2661	0.7693	0.0374	0.8950	Х	0.8950	0.4894	0.8393	0.7280	0.2362	0.6907	0.0096	0.1099	0.1302	0.7423	0.1329	0.0420	0.0420	0.9558
SM37	0.3314	0.1111	0.0000	0.9321	0.8434	0.5688	0.1369	0.9524	0.0240	Х	0.3770	0.1183	0.7735	0.2469	0.5934	0.0214	0.0024	0.5528	0.4821	0.5336	0.4371	0.4371	0.1365	0.9537
SM42	0.0116	0.1733	0.1762	0.0175	0.5883	0.3816	0.3063	0.5406	0.3180	Х	0.1948	0.6008	0.2435	0.9011	0.1066	0.0187	0.6354	0.3495	0.2544	0.0680	0.0680	0.3038	0.1323	0.1042
SM47*	0.0001	0.0001	0.0034	0.0000	0.0000	0.0148	0.0109	0.0109	0.0436	Х	0.0344	0.0049	0.0005	0.0000	0.0005	0.0139	0.0000	0.8770	0.0748	0.0723	0.0001	0.0002	0.0000	0.0129
SM51	0.4826	0.0358	1.0000	0.4594	0.2546	0.7119	0.9345	0.0168	0.8109	Х	0.0084	0.1212	0.0879	0.0000	0.4727	0.0964	0.0000	1.0000	0.1210	0.9500	0.0009	1.0000	0.0013	0.6977
SM53	0.0404	0.1050	0.4928	0.0198	0.2152	0.5037	0.0127	0.3701	0.5096	Х	0.0001	0.0563	0.5266	0.0079	0.0002	0.1633	0.0793	0.4904	0.7101	0.5299	0.0407	0.1953	0.0050	0.3397
SM55	0.3359	0.3359	1.0000	0.0649	0.7161	0.2880	0.4801	0.7785	0.4721	Х	0.0001	0.0024	0.1680	0.0000	0.0046	0.0013	0.1252	0.5268	0.0038	0.2160	0.1743	0.0520	0.0017	0.1076
SM56*	0.0048	0.1476	0.2541	0.2859	0.1478	0.0094	0.9618	0.0232	0.1542	Х	0.0000	0.0091	0.0056	0.0000	0.0014	0.0001	0.0000	0.2096	0.4728	0.4354	0.0001	0.2286	0.0000	0.5389
SM60	0.5242	0.2293	0.6131	0.6162	0.4016	0.3282	1.0000	0.0186	0.0058	Х	0.0018	0.0591	0.6880	0.0431	0.0199	0.6393	0.0005	0.8042	0.0092	0.1839	0.2218	0.1332	0.0012	0.4719
Aop943	0.1550	0.1506	0.2770	0.2268	0.1954	0.9360	0.2570	0.4430	0.7489	Х	0.4999	0.0912	0.1071	0.7181	0.7937	0.3839	0.8410	1.0000	1.0000	0.3578	0.0059	0.2053	0.5332	0.9946
Am116	0.4107	0.1192	0.8810	0.1784	0.1821	0.2278	0.9833	0.1946	0.8047	Х	0.1805	0.0236	0.4223	0.1911	0.6536	0.8772	0.6798	0.0297	0.4957	0.2049	0.1392	0.7788	0.4427	0.5740

**Table A4.2** Summary of *P*-values for the Hardy–Weinberg equilibrium using GENEPOP (Marchov chain parameters: 10,000 dememorizations, followed by 500 batches of 5,000 iterations per batch).

Significant P-values after Bonferroni correction are in bold type. \*indicates locus at which most populations show signs of deviance from HWE.

	1DA	1DB	1DC	IDD	1CA	1CB	1CC	1CD	2DA	2DB	2DC	2DD	2CA	2CB	2CC	2CD	3DA	3DB	3DC	3DD	3CA	3C-B	3CC	3CD
SM11	0.102	0.000	0.013	0.000	0.062	0.020	0.000	0.079	0.021	Х	0.101	0.058	0.054	0.066	0.031	0.038	0.052	0.000	0.038	0.000	0.101	0.036	0.000	0.003
SM14	0.037	0.017	0.069	0.034	0.064	0.000	0.051	0.092	0.025	Х	0.099	0.024	0.031	0.045	0.043	0.030	0.123	0.112	0.045	0.036	0.080	0.063	0.069	0.059
SM21A	0.093	0.051	0.029	0.022	0.071	0.000	0.013	0.098	0.000	Х	0.097	0.067	0.063	0.000	0.000	0.094	0.107	0.000	0.000	0.000	0.144	0.000	0.048	0.009
SM22*	0.165	0.080	0.000	0.131	0.039	0.014	0.149	0.205	0.105	Х	0.138	0.286	0.094	0.242	0.146	0.161	0.212	0.006	0.063	0.005	0.07	0.075	0.254	0.015
SM27*	0.156	0.147	0.026	0.110	0.046	0.041	0.023	0.213	0.172	Х	0.230	0.156	0.111	0.204	0.084	0.142	0.192	0.000	0.000	0.000	0.094	0.031	0.232	0.000
SM29	0.029	0.039	0.000	0.057	0.017	0.000	0.035	0.032	0.127	Х	0.177	0.212	0.059	0.178	0.049	0.139	0.202	0.000	0.000	0.066	0.099	0.041	0.149	0.000
SM34	0.003	0.000	0.000	0.000	0.039	0.000	0.000	0.065	0.000	Х	0.000	0.034	0.000	0.018	0.000	0.000	0.026	0.017	0.031	0.015	0.085	0.044	0.000	0.000
SM36	0.074	0.034	0.005	0.064	0.056	0.000	0.000	0.094	0.000	Х	0.000	0.000	0.000	0.022	0.027	0.000	0.056	0.008	0.000	0.000	0.064	0.058	0.065	0.000
SM37	0.077	0.078	0.000	0.000	0.009	0.000	0.000	0.000	0.022	Х	0.000	0.076	0.000	0.046	0.000	0.071	0.127	0.005	0.034	0.000	0.000	0.036	0.096	0.000
SM42	0.109	0.011	0.037	0.056	0.009	0.030	0.000	0.061	0.021	Х	0.018	0.000	0.000	0.000	0.070	0.048	0.034	0.000	0.002	0.088	0.000	0.025	0.009	0.019
SM47*	0.157	0.176	0.000	0.218	0.178	0.085	0.156	0.093	0.042	Х	0.113	0.160	0.143	0.179	0.095	0.137	0.173	0.009	0.096	0.085	0.161	0.162	0.231	0.090
SM51	0.034	0.081	0.000	0.000	0.035	0.000	0.000	0.104	0.014	Х	0.131	0.107	0.064	0.237	0.059	0.105	0.206	0.000	0.000	0.000	0.165	0.000	0.160	0.000
SM53	0.084	0.058	0.041	0.000	0.000	0.000	0.021	0.046	0.038	Х	0.161	0.112	0.035	0.172	0.187	0.076	0.105	0.040	0.000	0.000	0.113	0.079	0.167	0.000
SM55	0.019	0.054	0.000	0.070	0.000	0.010	0.010	0.031	0.036	Х	0.146	0.155	0.063	0.214	0.074	0.054	0.084	0.000	0.116	0.069	0.012	0.076	0.129	0.057
SM56*	0.125	0.083	0.000	0.080	0.082	0.116	0.010	0.135	0.081	Х	0.212	0.152	0.150	0.226	0.155	0.145	0.313	0.000	0.000	0.000	0.191	0.076	0.257	0.000
SM60	0.037	0.075	0.000	0.012	0.047	0.000	0.000	0.125	0.132	Х	0.130	0.106	0.000	0.112	0.127	0.033	0.179	0.000	0.000	0.076	0.030	0.068	0.158	0.000
Aop943	0.020	0.000	0.000	0.000	0.049	0.000	0.000	0.064	0.000	Х	0.058	0.000	0.099	0.002	0.000	0.000	0.000	0.000	0.000	0.011	0.000	0.000	0.000	0.000
Am116	0.051	0.080	0.000	0.046	0.025	0.000	0.000	0.000	0.000	Х	0.034	0.000	0.000	0.033	0.000	0.000	0.000	0.035	0.000	0.000	0.048	0.000	0.002	0.000
High (	$\geq 10^{\circ}$	%) fr	equer	ncy of	fnull	allele	es are	in bo	ld ty	pe.	*indi	cates	locu	s at w	hich	most	popu	latior	ns sho	w sig	gns of	null	allele	e.

Table A4.3 Summary of null allele frequencies for each pair of loci and populations using FREENA.

**Table A4.4** Pairwise population  $F_{ST}$  of (a) all individuals pooled, (b) mature trees and (c) saplings of sugar maple (*Acer saccharum*) for all 23 sites in Québec below the diagonal. *P*-values are found above the diagonal. Significant values after adjusted nominal level ( $\alpha = 0.05$ ) for multiple comparisons are given in bold.

(a)	1DA	1DB	1DC	1DD	1CA	1CB	1CC	1CD	2DA	2DC	2DD	2CA	2CB	2CC	2CD	3DA	3DB	3DC	3DD	3CA	3CB	3CC	3CD
1DA		0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0057	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1DB	0.0163		0.0002	0.0002	0.0002	0.0002	0.0002	0.0172	0.0002	0.0002	0.2727	0.0002	0.0002	0.0002	0.0103	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0012	0.0002
1DC	0.0457	0.0376		0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1DD	0.0210	0.0149	0.0546		0.0012	0.0002	0.0002	0.0002	0.0002	0.0002	0.0008	0.0002	0.0006	0.0002	0.0004	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1CA	0.0122	0.0072	0.0356	0.0060	_	0.0002	0.0002	0.0858	0.0002	0.0004	0.1654	0.0002	0.0002	0.0002	0.0006	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0295	0.0002
1CB	0.0198	0.0112	0.0335	0.0136	0.0044	_	0.0010	0.0061	0.0002	0.0004	0.0016	0.0002	0.0002	0.0002	0.0055	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0008	0.0002
1CC	0.0209	0.0131	0.0408	0.0147	0.0047	0.0079		0.0544	0.0002	0.0002	0.0638	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1CD	0.0140	0.0033	0.0303	0.0130	0.0012	0.0000	0.0029	-	0.0002	0.0008	0.1132	0.0053	0.0101	0.0332	0.0103	0.0002	0.0002	0.0002	0.0002	0.0002	0.0004	0.0190	0.0002
2DA	0.0271	0.0168	0.0505	0.0197	0.0139	0.019	0.0211	0.0167		0.0002	0.0022	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
2DC	0.0144	0.0150	0.0344	0.0128	0.0127	0.0108	0.0175	0.0101	0.0214	_	0.0708	0.0008	0.0091	0.0002	0.0334	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0008	0.0002
2DD	0.0174	0.0079	0.0329	0.0112	0.0051	0.0093	0.0132	0.0073	0.0103	0.0089	_	0.1192	0.6575	0.1441	0.6166	0.0018	0.0028	0.0002	0.0006	0.0077	0.0935	0.3455	0.1191
2CA	0.0141	0.0137	0.0362	0.0146	0.0096	0.0094	0.0147	0.0078	0.0236	0.0051	0.0038		0.0065	0.0595	0.0014	0.0002	0.0002	0.0002	0.0002	0.0002	0.0004	0.0063	0.0004
2CB	0.0208	0.0113	0.0427	0.0082	0.0086	0.0136	0.0242	0.0120	0.0148	0.0039	0.0009	0.0075	_	0.0059	0.0293	0.0002	0.0002	0.0002	0.0002	0.0002	0.0129	0.0316	0.0004
2CC	0.0179	0.0109	0.0455	0.0119	0.0078	0.0089	0.0168	0.0061	0.0161	0.0112	0.0051	0.0022	0.0054		0.1777	0.0002	0.0002	0.0002	0.0002	0.0002	0.0020	0.2374	0.0002
2CD	0.0138	0.0098	0.0327	0.0103	0.0047	0.0021	0.0130	0.0050	0.0156	0.0045	0.0052	0.0077	0.0060	-0.0003	_	0.0006	0.0002	0.0002	0.0002	0.0002	0.0067	0.1186	0.0006
3DA	0.029	0.0202	0.0519	0.0316	0.0211	0.0207	0.0267	0.0150	0.0107	0.0244	0.0223	0.0307	0.018	0.0150	0.0097	_	0.0002	0.0002	0.0002	0.0002	0.0002	0.0004	0.0002
3DB	0.0191	0.0251	0.0574	0.0181	0.0158	0.0188	0.0278	0.0186	0.0179	0.0170	0.0116	0.0115	0.0125	0.0075	0.0120	0.0205	_	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
3DC	0.0380	0.0332	0.0613	0.0320	0.0223	0.0217	0.0379	0.0265	0.0334	0.0316	0.0258	0.0163	0.0198	0.0121	0.0230	0.0335	0.0185	_	0.0002	0.0002	0.0002	0.0002	0.0002
3DD	0.0266	0.029	0.038	0.0310	0.0172	0.0143	0.0225	0.0184	0.0254	0.0131	0.0138	0.0144	0.0197	0.0133	0.0074	0.0187	0.0196	0.0238		0.0002	0.0002	0.0002	0.0002
3CA	0.0298	0.0205	0.0578	0.0146	0.0153	0.0213	0.0197	0.0140	0.0227	0.0200	0.0093	0.0125	0.0072	0.0118	0.0189	0.0242	0.0112	0.0224	0.0297	_	0.0002	0.0002	0.0002
3CB	0.0175	0.0133	0.0448	0.0112	0.0095	0.0125	0.0166	0.0110	0.0097	0.0089	0.0075	0.0076	0.0055	0.0034	0.0018	0.0150	0.0100	0.0180	0.0126	0.0130	_	0.2328	0.0002
3CC	0.0193	0.0098	0.0415	0.0164	0.0041	0.0079	0.0167	0.0086	0.0157	0.0174	0.0037	0.0073	0.0086	0.0007	0.0033	0.0193	0.0154	0.0195	0.0170	0.0158	0.0042		0.0055
3CD	0.0165	0.0163	0.0397	0.0180	0.0062	0.0085	0.0135	0.0084	0.0192	0.0108	0.0141	0.0079	0.0112	0.005	0.0045	0.0137	0.0118	0.0181	0.0096	0.0173	0.0068	0.0090	—

Table A4.4	(to continued)

(b)	1DA	1DB	1DC	1DD	1CA	1CB	1CC	1CD	2DA	2DC	2DD	2CA	2CB	2CC	2CD	3DA	3DB	3DC	3DD	3CA	3CB	3CC	3CD
1DA		0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0057	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1DB	0.0163	_	0.0002	0.0002	0.0002	0.0002	0.0002	0.0172	0.0002	0.0002	0.2727	0.0002	0.0002	0.0002	0.0103	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0012	0.0002
1DC	0.0457	0.0376	_	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1DD	0.0210	0.0149	0.0546	_	0.0012	0.0002	0.0002	0.0002	0.0002	0.0002	0.0008	0.0002	0.0006	0.0002	0.0004	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1CA	0.0122	0.0072	0.0356	0.0060		0.0002	0.0002	0.0858	0.0002	0.0004	0.1654	0.0002	0.0002	0.0002	0.0006	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0295	0.0002
1CB	0.0198	0.0112	0.0335	0.0136	0.0044		0.0010	0.0061	0.0002	0.0004	0.0016	0.0002	0.0002	0.0002	0.0055	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0008	0.0002
1CC	0.0209	0.0131	0.0408	0.0147	0.0047	0.0079		0.0544	0.0002	0.0002	0.0638	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1CD	0.0140	0.0033	0.0303	0.0130	0.0012	0.0000	0.0029	_	0.0002	0.0008	0.1132	0.0053	0.0101	0.0332	0.0103	0.0002	0.0002	0.0002	0.0002	0.0002	0.0004	0.0190	0.0002
2DA	0.0271	0.0168	0.0505	0.0197	0.0139	0.019	0.0211	0.0167	_	0.0002	0.0022	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
2DC	0.0144	0.0150	0.0344	0.0128	0.0127	0.0108	0.0175	0.0101	0.0214	—	0.0708	0.0008	0.0091	0.0002	0.0334	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0008	0.0002
2DD	0.0174	0.0079	0.0329	0.0112	0.0051	0.0093	0.0132	0.0073	0.0103	0.0089	_	0.1192	0.6575	0.1441	0.6166	0.0018	0.0028	0.0002	0.0006	0.0077	0.0935	0.3455	0.1191
2CA	0.0141	0.0137	0.0362	0.0146	0.0096	0.0094	0.0147	0.0078	0.0236	0.0051	0.0038	_	0.0065	0.0595	0.0014	0.0002	0.0002	0.0002	0.0002	0.0002	0.0004	0.0063	0.0004
2CB	0.0208	0.0113	0.0427	0.0082	0.0086	0.0136	0.0242	0.0120	0.0148	0.0039	0.0009	0.0075		0.0059	0.0293	0.0002	0.0002	0.0002	0.0002	0.0002	0.0129	0.0316	0.0004
2CC	0.0179	0.0109	0.0455	0.0119	0.0078	0.0089	0.0168	0.0061	0.0161	0.0112	0.0051	0.0022	0.0054		0.1777	0.0002	0.0002	0.0002	0.0002	0.0002	0.0020	0.2374	0.0002
2CD	0.0138	0.0098	0.0327	0.0103	0.0047	0.0021	0.0130	0.0050	0.0156	0.0045	0.0052	0.0077	0.0060	-0.0003	_	0.0006	0.0002	0.0002	0.0002	0.0002	0.0067	0.1186	0.0006
3DA	0.029	0.0202	0.0519	0.0316	0.0211	0.0207	0.0267	0.0150	0.0107	0.0244	0.0223	0.0307	0.018	0.0150	0.0097	_	0.0002	0.0002	0.0002	0.0002	0.0002	0.0004	0.0002
3DB	0.0191	0.0251	0.0574	0.0181	0.0158	0.0188	0.0278	0.0186	0.0179	0.0170	0.0116	0.0115	0.0125	0.0075	0.0120	0.0205	_	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
3DC	0.0380	0.0332	0.0613	0.0320	0.0223	0.0217	0.0379	0.0265	0.0334	0.0316	0.0258	0.0163	0.0198	0.0121	0.0230	0.0335	0.0185	_	0.0002	0.0002	0.0002	0.0002	0.0002
3DD	0.0266	0.029	0.038	0.0310	0.0172	0.0143	0.0225	0.0184	0.0254	0.0131	0.0138	0.0144	0.0197	0.0133	0.0074	0.0187	0.0196	0.0238	_	0.0002	0.0002	0.0002	0.0002
3CA	0.0298	0.0205	0.0578	0.0146	0.0153	0.0213	0.0197	0.0140	0.0227	0.0200	0.0093	0.0125	0.0072	0.0118	0.0189	0.0242	0.0112	0.0224	0.0297		0.0002	0.0002	0.0002
3CB	0.0175	0.0133	0.0448	0.0112	0.0095	0.0125	0.0166	0.0110	0.0097	0.0089	0.0075	0.0076	0.0055	0.0034	0.0018	0.0150	0.0100	0.0180	0.0126	0.0130	—	0.2328	0.0002
3CC	0.0193	0.0098	0.0415	0.0164	0.0041	0.0079	0.0167	0.0086	0.0157	0.0174	0.0037	0.0073	0.0086	0.0007	0.0033	0.0193	0.0154	0.0195	0.0170	0.0158	0.0042	_	0.0055
3CD	0.0165	0.0163	0.0397	0.0180	0.0062	0.0085	0.0135	0.0084	0.0192	0.0108	0.0141	0.0079	0.0112	0.005	0.0045	0.0137	0.0118	0.0181	0.0096	0.0173	0.0068	0.0090	—

Table	A4.4	(to	continued).

	1DA	1DB	1DC	1DD	1CA	1CB	1CC	1CD	2DA	2DC	2DD	2CA	2CB	2CC	2CD	3DA	3DB	3DC	3DD	3CA	3CB	3CC	3CD
1DA	_	0.0038	0.0002	0.0002	0.1348	0.0014	0.0002	0.0376	0.0002	0.0004	0.0629	0.0047	0.0144	0.0014	0.0032	0.0004	0.0004	0.0002	0.0002	0.0004	0.0004	0.0002	0.0002
1DB	0.0188	—	0.0002	0.0480	0.1887	0.0360	0.2668	0.6065	0.0008	0.0018	0.4737	0.0002	0.1935	0.1170	0.2332	0.0004	0.0002	0.0002	0.0002	0.0061	0.0002	0.0215	0.0008
1DC	0.0438	0.0297	—	0.0002	0.0002	0.0002	0.0002	0.0004	0.0002	0.0002	0.0006	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002	0.0002
1DD	0.0238	0.0061	0.0529	—	0.0063	0.0004	0.0613	0.0190	0.0002	0.0024	0.0125	0.0004	0.0295	0.0004	0.0061	0.0010	0.0002	0.0002	0.0002	0.0034	0.0002	0.0004	0.0002
1CA	0.0018	0.0049	0.0270	0.0109	_	0.0381	0.0077	0.8992	0.0004	0.0051	0.2298	0.0682	0.0340	0.0658	0.3595	0.0016	0.0006	0.0008	0.0002	0.0044	0.0217	0.1710	0.0075
1CB	0.0149	0.0062	0.0345	0.0198	0.0031	_	0.1872	0.9055	0.0012	0.0014	0.1316	0.0253	0.0059	0.1233	0.1490	0.0010	0.0002	0.0002	0.0002	0.0210	0.0010	0.1283	0.0069
1CC	0.0217	0.0119	0.0392	0.0095	0.0081	0.0073	—	0.4842	0.0012	0.0050	0.0743	0.0002	0.0040	0.0097	0.0354	0.0059	0.0002	0.0002	0.0002	0.0030	0.0012	0.0028	0.0014
1CD	0.0162	0.0018	0.0242	0.0101	-0.0030	-0.0014	0.0010	_	0.0047	0.0166	0.8344	0.2150	0.3719	0.7710	0.5812	0.0370	0.0049	0.0030	0.0004	0.2372	0.2937	0.3504	0.0461
2DA	0.0233	0.0132	0.0440	0.0245	0.0067	0.0132	0.0147	0.0133	_	0.0002	0.0407	0.0002	0.0457	0.0002	0.0012	0.0223	0.0002	0.0004	0.0002	0.0002	0.0002	0.0075	0.0002
2DC	0.0176	0.0267	0.0345	0.0239	0.0154	0.0220	0.0169	0.0225	0.0255	_	0.0617	0.0006	0.0251	0.0223	0.0712	0.0008	0.0002	0.0002	0.0002	0.0004	0.0060	0.0065	0.0006
2DD	0.0190	0.0038	0.0269	0.0097	0.0041	0.0058	0.0119	-0.0030	0.0054	0.0219	_	0.0551	0.5314	0.2913	0.9476	0.1168	0.0026	0.0004	0.0573	0.0453	0.4700	0.1012	0.3822
2CA	0.0092	0.0225	0.0330	0.0223	0.0033	0.0056	0.0164	0.0052	0.0159	0.0109	0.0019		0.0779	0.4449	0.0646	0.0002	0.0004	0.0002	0.0004	0.0636	0.0032	0.0217	0.0320
2CB	0.0246	0.0105	0.0333	0.0121	0.0061	0.0126	0.0193	0.0086	0.0045	0.0107	0.0002	0.0076	_	0.4210	0.3441	0.0310	0.0036	0.0111	0.0002	0.0192	0.0431	0.0073	0.0057
2CC	0.0236	0.0111	0.0471	0.0157	0.0117	0.0126	0.0151	0.0024	0.0184	0.0215	0.0013	0.0019	0.0069	_	0.3609	0.0085	0.0304	0.0154	0.0002	0.1636	0.2259	0.2127	0.1658
2CD	0.0159	0.0122	0.0242	0.0198	-0.0003	0.0098	0.0159	0.0010	0.0192	0.0145	-0.0021	0.0073	0.0099	0.0132	_	0.0178	0.0014	0.0002	0.0449	0.0004	0.5028	0.3763	0.0385
3DA	0.0223	0.0223	0.0391	0.0313	0.0159	0.0217	0.0185	0.0137	0.0141	0.0239	0.0066	0.0235	0.0086	0.008	0.0116	_	0.0018	0.0006	0.0002	0.0018	0.0004	0.0200	0.0030
3DB	0.0153	0.0249	0.0566	0.0241	0.0151	0.0218	0.0330	0.0160	0.0188	0.0299	0.0107	0.0089	0.0151	0.0104	0.0169	0.0162	_	0.0002	0.0002	0.0285	0.0010	0.0002	0.0113
3DC	0.0371	0.0447	0.0688	0.0413	0.0227	0.0175	0.0322	0.0267	0.0327	0.0413	0.0257	0.0117	0.0268	0.0136	0.0342	0.0284	0.0224	—	0.0002	0.0079	0.0002	0.0002	0.0002
3DD	0.0253	0.0342	0.0391	0.0334	0.0121	0.0178	0.0243	0.0182	0.0277	0.0238	0.0093	0.0105	0.0266	0.0218	0.0040	0.0218	0.0225	0.0259	—	0.0002	0.0002	0.0006	0.0002
3CA	0.027	0.0192	0.0590	0.0181	0.0161	0.0139	0.0189	0.0102	0.0200	0.0313	0.0081	0.0057	0.0106	0.0069	0.0220	0.0174	0.0083	0.0135	0.0376	_	0.0872	0.0312	0.0105
3CB	0.0200	0.0185	0.0497	0.0145	0.0090	0.0161	0.0242	0.0092	0.0123	0.0230	0.0081	0.0112	0.0146	0.0050	-0.0004	0.0155	0.0064	0.0248	0.0171	0.0092	—	0.5057	0.0014
3CC	0.0249	0.0080	0.0450	0.0238	0.0058	0.0028	0.0252	0.0065	0.0181	0.0277	0.0054	0.0051	0.0129	0.0025	0.0052	0.0162	0.0165	0.0273	0.0251	0.0093	0.0017	—	0.0061
3CD	0.0086	0.0199	0.0384	0.0221	0.0018	0.0044	0.0089	0.0086	0.0174	0.019	0.0067	0.0016	0.0138	0.0055	0.0038	0.0102	0.0114	0.0104	0.0095	0.0134	0.0111	0.0155	—

Table A4.5 Comparison of mean genetic variability estimates between cohorts and cohorts in zones, transects and zones in each transect o	f
sugar maple (Acer saccharum) in Québec.	

Genetic	Cohor	ts		Zones						Transe	ects							
Indices	М	Sa	P-value	D			С			1			2			3		
				М	Sa	P-value	М	Sa	P-value	М	Sa	P-value	М	Sa	P-value	М	Sa	P-value
$A_{\rm R}$	5.378	5.275	0.2620	5.262	5.200	0.6480	5.484	5.343	0.2320	5.251	5.271	0.8840	5.482	5.382	0.4810	5.414	5.185	0.1340
$H_{\rm O}$	0.610	0.591	0.3670	0.603	0.598	0.8740	0.616	0.585	0.2500	0.620	0.612	0.8020	0.583	0.551	0.423	0.622	0.601	0.1340
$H_{\rm E}$	0.693	0.694	0.8990	0.689	0.689	0.9790	0.697	0.698	0.8560	0.687	0.692	0.5440	0.693	0.697	0.6260	0.699	0.692	0.5450
$F_{\rm IS}$	0.121	0.148	0.3690	0.126	0.132	0.8780	0.116	0.163	0.2460	0.097	0.097	0.6830	0.158	0.210	0.3890	0.111	0.131	0.7050
$F_{\rm ST}$	0.015	0.017	0.6650	0.023	0.027	0.6390	0.009	0.009	0.9730	0.021	0.021	0.5470	0.007	0.011	0.6120	0.013	0.016	0.6880
									Transects	s & Zor	nes							

	D									С								
	1			2			3			1			2			3		
	М	Sa	P-value	Μ	Sa	P-value	М	Sa	P-value	Μ	Sa	P-value	М	Sa	P-value	М	Sa	P-value
$A_{\rm R}$	4.951	5.078	0.5040	5.591	5.457	0.5360	5.326	5.130	0.2970	5.552	5.464	0.6810	5.400	5.327	0.6900	5.501	5.240	0.1950
$H_0$	0.599	0.596	0.9700	0.568	0.565	0.9670	0.627	0.620	0.8750	0.641	0.627	0.7750	0.593	0.541	0.3120	0.616	0.582	0.4900
$H_{\rm E}$	0.676	0.682	0.6350	0.703	0.703	0.9970	0.692	0.685	0.6360	0.696	0.703	0.6140	0.686	0.693	0.5420	0.707	0.698	0.4860
$F_{\rm IS}$	0.114	0.126	0.8730	0.193	0.197	0.9610	0.094	0.096	0.9810	0.079	0.109	0.6770	0.135	0.219	0.2560	0.129	0.166	0.6150
$F_{\rm ST}$	0.034	0.029	0.6610	0.013	0.018	0.7160	0.019	0.022	0.7590	0.007	0.003	0.6900	0.003	0.008	0.6430	0.006	0.010	0.7690

 $A_{\rm R}$ , mean allelic richness;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm ST}$ , mean pairwise  $F_{\rm ST}$ ;  $F_{\rm IS}$ , inbreeding coefficient; D, discontinuous zone; C, continuous zone; M, mature trees; Sa, saplings. Significant values ( $\alpha = 0.05$ ) given in bold.

**Table A4.6** Difference in Akaike information criterion ( $\Delta AIC_C$ ), weights ( $\omega$ ) and number of estimated parameters (*K*) for model comparisons of the relative importance of stand characteristics and distance from northern limit as predictors of  $A_R$ ,  $H_O$ ,  $H_E$  and  $F_{IS}$  (n = 414), and  $F_{ST}$  (n = 23) in Québec. For brevity, only models with a  $\Delta AIC_C \le 4.0$  and the next model were shown.

Explained variables	Model	Κ	$\Delta AIC_{C}$	ω
$A_{\mathrm{R}}$	m_ers_PS + D_north	5	0.00	0.19
	$s_ers_PS + D_north$	5	0.27	0.16
	$m_{ers}PS + s_{ers}PS + D_{north}$	6	0.39	0.15
	D_north	4	0.70	0.13
	$m_{ers_d} + D_{north}$	5	1.24	0.10
	$m_{ers}BA + D_{north}$	5	1.67	0.08
	s_ers_BA +D_north	5	2.68	0.05
	s_ers_d + D_north	5	2.68	0.05
	$m_{ers_d + s_{ers_d + D_north}}$	6	2.99	0.04
	$m_ers_BA + s_ers_BA + D_north$	6	3.71	0.03
	m_ers_BA	4	7.37	1.00
$H_0$	$m_{ers}BA + s_{ers}BA$	5	0.00	0.30
	$m_ers_BA + s_ers_BA + D_north$	6	1.75	0.12
	m_ers_BA	4	1.86	0.12
	$m_{ers}PS + D_{north}$	5	2.50	0.09
	$m\_ers\_PS + s\_ers\_PS + D\_north$	6	2.99	0.07
	m_ers_d	4	3.19	0.06
	$m_{ers}BA + D_{north}$	5	3.90	0.04
	$m_{ers_d} + D_{north}$	5	4.51	0.03
$H_{\rm E}$	m_ers_d	4	0.00	0.09
	$m_{ers}PS + D_{north}$	5	0.07	0.09
	m_ers_BA	4	0.09	0.09
	D_north	4	0.35	0.08
	s_ers_PS	4	0.40	0.07
	m_ers_PS	4	0.52	0.07
	s_ers_PS + D_north	5	0.70	0.06
	s_ers_BA	4	0.79	0.06
	$m_ers_PS + s_ers_PS + D_north$	6	1.05	0.05
	s_ers_d	4	1.08	0.05
	$m_{ers}BA + s_{ers}BA$	5	1.29	0.05
	$m_ers_d + D_north$	5	1.41	0.04
	m_ers_BA	5	1.95	0.03
	$m_ers_d + s_ers_d$	5	1.96	0.03
	s_ers_BA + D_north	5	2.12	0.03
	$m_{ers}_{PS} + s_{ers}_{PS}$	5	2.22	0.03
	$s_ers_d + D_north$	5	2.38	0.03

Explained variables	Model	K	$\Delta AIC_{C}$	ω
	$m_ers_BA + s_ers_BA + D_north$	6	3.29	0.02
	$m_ers_d + s_ers_d + D_north$	6	3.46	0.02
F <sub>IS</sub>	m_ers_BA + s_ers_BA	5	0.00	0.22
	m_ers_BA	4	1.23	0.12
	$m_{ers}BA + s_{ers}BA + D_{north}$	6	1.49	0.10
	m_ers_d	4	2.29	0.07
	s_ers_BA	4	2.31	0.07
	m_ers_PS	4	2.70	0.06
	m_ers_PS + D_north	5	3.16	0.04
	m_ers_BA + D_north	5	3.16	0.04
	s_ers_PS	4	3.27	0.04
	D_north	4	3.53	0.04
	s_ers_d	4	3.65	0.03
	$m_{ers_d + s_{ers_d}}$	5	4.02	0.03
F <sub>ST</sub>	D_north	3	0.00	0.22
	$m_{ers_d} + D_{north}$	4	0.41	0.18
	$m_{ers}BA + D_{north}$	4	1.71	0.09
	s_ers_BA + D_north	4	1.98	0.08
	m_ers_BA	3	1.99	0.08
	s_ers_PS + D_north	4	2.83	0.05
	s_ers_d + D_north	4	2.93	0.05
	$m_{ers}PS + D_{north}$	4	2.96	0.05
	$m_ers_d + s_ers_d + D_north$	5	3.68	0.03
	m_ers_d	3	3.83	0.03
	m_ers_PS	3	4.26	0.03

 Table A4.6 (to continued)

Explanatory variable abbreviations: m ers\_BA, mature sugar maple basal area (m<sup>2</sup> ha<sup>-1</sup>); s\_ers\_BA, sugar maple sapling basal area (m<sup>2</sup> ha<sup>-1</sup>); m\_ers\_d, mature sugar maple density (stems ha<sup>-1</sup>); s\_ers\_d, sugar maple sapling density (stems ha<sup>-1</sup>); m\_ers\_PS, mature sugar maple population size (stems); s\_ers\_PS, sugar maple sapling population size (stems); D\_north, distance of each site to the northern limit (km).

Transects	Zones	Sites (	CW5	CL5	CP5	RW5	RL5	RP5	CW2	CL2	CP2	RW2	RL2	RP2
1	Discontinuous	1DA	53	1	0	41	5	1	73	2	1	21	4	0
		1DB	57	0	0	39	5	0	77	3	0	18	2	0
		1DC	54	0	0	42	4	0	68	0	0	28	4	0
		1DD	55	1	0	37	8	1	78	3	1	13	6	0
	Continuous	1CA	52	0	0	39	9	1	71	4	0	20	4	1
		1CB	58	1	0	36	5	0	81	2	0	14	3	0
		1CC	54	1	1	41	5	2	77	2	1	17	3	2
		1CD	50	1	0	43	6	0	66	3	0	27	3	0
2	Discontinuous	2DA	60	0	0	38	3	0	79	1	0	18	2	0
		2DC	56	0	0	37	7	1	76	2	1	16	5	0
		2DD	49	1	0	43	8	1	75	3	0	16	6	1
	Continuous	2CA	51	0	0	41	7	1	78	3	0	15	5	1
		2CB	54	0	0	38	8	2	70	4	0	22	4	2
		2CC	52	0	0	43	6	0	74	2	0	20	4	0
		2CD	51	0	0	42	7	1	75	1	0	18	6	1
3	Discontinuous	3DA	56	0	0	36	8	1	78	4	0	13	4	1
		3DB	63	1	0	33	4	0	84	3	0	12	1	0
		3DC	58	2	1	34	6	0	74	4	1	18	5	0
		3DD	68	1	0	27	4	1	82	1	0	13	4	1
	Continuous	3CA	56	0	0	37	7	1	72	1	0	21	6	1
		3CB	50	1	0	44	5	2	76	2	1	18	4	1
		3CC	55	0	0	40	5	0	78	1	0	17	3	0
		3CD	59	0	0	32	9	2	77	0	0	14	9	2

**Table A4.7** Raw data (%) of allele classification based on the frequency and geographic distribution of all individuals from 23 Québec sugar maple (*Acer saccharum*) sites.

CW5, common (allele frequency  $(f) \ge 0.05$ ) and widespread; CL5, common and localized; CP5, common and private; RW5, rare (f < 0.05) and widespread; RL5, rare and localized; RP5, rare and private; CW2, common  $(f \ge 0.02)$  and widespread; CL2, common and localized; CP2, common and private; RW2, rare (f < 0.02) and widespread; RL2, rare and localized; RP2, rare and private.

Transects	Zone	Sites	Matu	re trees	S				Sapli	ngs				
			CW5	CL5 C	CP5	RW5	RL5	RP5	CW5	CL5	CP5	RW5	RL5	RP5
1	Discontinuous	1DA	68	1	0	27	4	2	66	4	0	25	6	1
		1DB	72	2	0	19	6	1	69	2	0	26	3	1
		1DC	71	0	0	24	5	0	67	2	0	28	3	1
		1DD	63	2	0	27	7	1	73	2	1	20	5	0
	Continuous	1CA	65	3	0	26	6	2	62	5	1	29	4	0
		1CB	76	1	0	18	6	0	69	2	0	23	7	0
		1CC	70	3	1	22	5	1	73	0	0	21	6	2
		1CD	64	2	0	26	8	1	59	2	0	32	7	1
2	Discontinuous	2DA	69	1	0	25	6	0	76	4	1	18	2	0
		2DC	68	1	1	22	9	1	73	4	1	16	6	1
		2DD	68	1	0	25	6	1	61	4	0	28	8	1
	Continuous	2CA	62	2	0	29	7	2	68	2	0	23	7	2
		2CB	67	2	2	26	5	1	63	3	0	27	7	2
		2CC	50	1	0	40	9	1	68	1	0	26	6	2
		2CD	68	1	0	26	6	2	71	1	0	18	10	2
3	Discontinuous	3DA	70	6	0	19	6	1	73	2	0	20	5	2
		3DB	77	1	0	20	2	0	73	4	0	18	5	0
		3DC	70	3	1	22	6	1	73	3	1	20	5	1
		3DD	79	1	0	17	3	0	78	1	0	20	2	1
	Continuous	3CA	67	2	0	20	10	2	69	1	0	24	6	1
		3CB	69	2	0	22	7	1	68	2	1	26	5	1
		3CC	71	0	0	21	7	0	70	2	0	24	3	0
		3CD	75	1	0	18	6	2	70	2	0	21	7	2
		mean	68.6	1.7	0.2	23.5	6.3	0.8	69.1	2.3	0.2	23.1	5.4	0.9
		(SD)	5.8	1.3	0.4	4.9	1.8	0.6	4.7	1.4	0.4	4.1	1.9	0.7

 Table A4.8 Raw data (%) of allele classification based on the frequency and geographic distribution of each cohort for all 23 sugar maple (*Acer saccharum*) sites in Québec.

CW5, common (allele frequency  $(f) \ge 0.05$ ) and widespread; CL5, common and localized; CP5, common and private; RW5, rare (f < 0.05) and widespread; RL5, rare and localized; RP5, and rare and private; SD, standard deviation.



**Figure A4.1** Neighbour-joining tree based on 18 microsatellite loci (a) for the 23 sugar maple (*Acer saccharum*) sites in Québec and (b) separated into two cohorts. Genetic distances ( $D_S$ ) were based on Nei's (1972) genetic distance. Numbers given in nodes indicate the support value (%) of the respective group based on 1000 bootstrapped trees. Results are shown only for branches with  $\geq 25\%$  support. Abbreviations given after site name: m, mature sugar maple and s: sugar maple sapling.


**Figure A4.2** Detection of the number of clusters,  $\Delta K$  plot showing greatest support at K = 4, using STRUCTURE for 23 sugar maple (*Acer saccharum*) sites according to Evanno *et al.* (2005).

Table A4.9 Results of analysis of molecular variance (AMOVA) showing the partitionin	ng of
genetic variance among transects and zones.	

Source of variation	d.f.	Sum of	Variance	Percentage	Phi $(\Phi)$	P-values
		squares	components	of variance	statistics	
Original grouping: transect						
Between transect	2	132.851	0.108	1	0.007	0.000
Between populations within transect	20	672.927	0.454	3	0.028	0.000
Within populations	890	13906.340	15.625	97	0.035	0.000
Total	912	14712.118	16.187	100		
Original grouping: zone						
Between zone	1	29.170	0.000	0	-0.001	1.000
Between populations within zone	21	776.608	0.538	3	0.033	0.000
Within populations	890	13906.340	15.625	97	0.032	0.000
Total	912	14712.118	16.163	100		
Original grouping: zone in transect 1						
Between zone	1	42.443	0.057	0	0.004	0.012
Between populations within zone	6	201.280	0.465	3	0.029	0.000
Within populations	303	4693.708	15.491	97	0.033	0.000
Total	310	4937.431	16.013	100		
Original grouping: zone in transect 2						
Between zone	1	34.716	0.034	0	0.002	0.092
Between populations within zone	5	150.366	0.335	2	0.020	0.000
Within populations	275	4558.382	16.576	98	0.022	0.000
Total	281	4743.465	16.945	100		
Original grouping: zone in transect 3						
Between zone	1	30.866	0.000	0	-0.002	0.960
Between populations within zone	6	213.256	0.516	3	0.033	0.000
Within populations	312	4654.250	14.917	97	0.032	0.000
Total	319	4898.372	15.433	100		
STRUCTURE grouping A						
Between groups	4	249.711	0.187	1	0.012	0.000
Between populations within group	18	556.067	0.381	2	0.024	0.000
Within populations	890	13906.340	15.625	96	0.035	0.000
Total	912	14712.118	16.193	100		
STRUCTURE grouping B						
Between groups	3	182.386	0.124	1	0.008	0.000
Between populations within group	19	623.392	0.433	3	0.027	0.000
Within populations	890	13906.340	15.625	97	0.034	0.000
Total	912	14712.118	16.182	100		

d.f., degree of freedom. Significant values at  $\alpha = 0.05$  given in bold and at  $\alpha = 0.10$  given in italics.

**Table A4.10** Mean mutation-scaled population size ( $\theta = 4N_e\mu$ , where  $N_e$  = effective population size and  $\mu$  = mutation rate per generation per locus) and mean mutation-scaled immigration rate ( $M = m/\mu$ , where m = migration rate) between pairs of sugar maple (*Acer saccharum*) populations. Both parameters are estimated using MIGRATE-N (Beerli, 2006).

	- · · · · · · · · · · · · · · · · · · ·											
	$\theta$	NP	Transect 1	Transect 2	Transect 3							
NP→i	1.43		7.8	8.5	7.6							
	(0.57–2.27)		(6.7 - 8.8)	(6.7–9.9)	(5.8–9.8)							
Transect 1→i	2.77	6.6	_	15.5	15.6							
	(1.7–3.80)	(4.6 - 7.5)		(14.6 - 16.4)	(14.5 - 16.9)							
Transect $2 \rightarrow i$	1.22	16.7	17.8	_	15.0							
	(0.03–2.37)	(15.5 - 17.9)	(16.7 - 18.8)		(13.9–16.1)							
Transect 3→i	3.68	25.7	20.8	16.9	_							
	(2.33 - 5.30)	(24.8 - 26.7)	(19.1 - 22.5)	(15.6 - 18.8)								

95% confidence intervals are shown in parentheses. Source groups are presented in the first column and sink groups are presented in the first row. NP, northern populations (1-D-A, 1-D-B, 2-D-A, 3-D-A). Here Transect 1 représents 1-D-C, 1-D-D, 1-C-A, 1-C-B, 1-C-C, 1-C-D; Transect 2, 2-D-C, 2-D-D, 2-C-A, 2-C-B, 2-C-C, 2-C-D; and Transect 3,3-D-B, 3-D-C, 3-D-D, 3-C-A, 3-C-B, 3-C-C, 3-C-D.

## ANNEXE A5

## SUPPORTING INFORMATION FOR CHAPTER V

Table A5.1 Summary of sample site coordinates and protection types.

Study	Study	Protection	Latitude	Longitude	Elevation	Distance to
ID	name	type			(m)	paired OG
OG1	Lac Tucker	EFE	45° 56′ 36.6″ N	75° 47′ 53.5″ W	386	
SC1	Lac St Charles		45° 47′ 37.3″ N	75° 52′ 3.7″ W	305	17.3
OG2	Lac de l'Ecluse	EFE	45° 51′ 35.6″ N	75° 23′ 55.8″ W	388	
SC2	Lac Faucon		45° 49′ 15.5″ N	75° 19′ 18.8″ W	288	7.8
OG3	Lac Marie-Lefranc	EFE	46° 70′ 5.34″ N	75° 00′ 36.5″ W	410	
SC3	Lac Marie-Lefranc		46° 50′ 52.8″ N	75° 00′ 40.3″ W	386	1.6

OG, old-growth stand; SC, selection cutting stand; EFE, *écosystèmes forestiers exceptionnels*. For more details on stand characteristics (except for OG2), see Angers *et al.* (2005), and for OG2 and OG3 (same as 2-C-D and 2-C-C, respectively), see Graignic *et al.* (2014).

Locus	GenBank	М	Size range	$A_{\mathrm{T}}$	A	$H_0$	$H_{\rm E}$	$F_{\rm IS}$	References
	access no.		(bp)						
SM11	KC731552	3	178–199	12	8.2	0.586	0.629	0.070	Graignic et al. (2013)
SM14	KC751436	4	70-120	21	16.2	0.742	0.894	0.170	Graignic et al. (2013)
SM21A	KC751437	4	179–237	28	17.7	0.755	0.867	0.129	Graignic et al. (2013)
SM22	KC751438	2	293-323	17	13.7	0.526	0.896	0.414	Graignic et al. (2013)
SM27	KC751440	4	242-260	9	7.0	0.403	0.673	0.405	Graignic et al. (2013)
SM29	KC751441	4	278-307	10	7.5	0.527	0.724	0.274	Graignic et al. (2013)
SM34	KC751442	3	118–167	23	15.8	0.772	0.843	0.084	Graignic et al. (2013)
SM36	KC751443	5	146–182	18	13.3	0.714	0.828	0.137	Graignic et al. (2013)
SM37	KC751444	2	174–196	12	9.0	0.564	0.667	0.153	Graignic et al. (2013)
SM42	KC751445	1	90–133	17	9.5	0.678	0.804	0.157	Graignic et al. (2013)
SM47	KC751446	2	201-225	12	8.2	0.390	0.667	0.415	Graignic et al. (2013)
SM51	KC751447	1	269–290	7	5.5	0.344	0.464	0.258	Graignic et al. (2013)
SM53	KC751448	5	287-310	7	3.0	0.352	0.517	0.317	Graignic et al. (2013)
SM55	KC751449	2	248-276	15	10.3	0.480	0.686	0.296	Graignic et al. (2013)
SM56	KC751450	3	287–299	6	5.0	0.364	0.611	0.400	Graignic et al. (2013)
SM60	KC751452	3	231-237	3	2.8	0.305	0.410	0.246	Graignic et al. (2013)
Aop943	EF531298	1	143-160	8	5.7	0.565	0.567	0.002	Segarra-Moragues et al. (2008)
									Graignic et al. (2013)
Am116	AB303350	1	230-267	18	11.5	0.589	0.701	0.160	Kikuchi & Shibata (2008)
									Graignic et al. (2013)

**Table A5.2** Genetic variability estimates of microsatellite markers used in the Québec study of sugar maple (*Acer saccharum*).

M, multiplexing arrangement;  $A_{\rm T}$ , total number of alleles; A, mean number of alleles per locus;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm IS}$ , inbreeding coefficient.

**Table A5.3** Summary of *P*-values for Hardy–Weinberg equilibrium using GENEPOP (Markov chain parameters: 10,000 dememorizations, followed by 500 batches of 5,000 iterations per batch).

	OG1	SC1	OG2	SC2	OG3	SC3
SM11	0.1903	0.1722	0.5947	0.0212	0.1487	0.0110
SM14	0.0011	0.0026	0.0102	0.0027	0.0035	0.0000
SM21A	0.0384	0.1047	0.0140	0.0066	0.8285	0.0370
SM22*	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
SM27*	0.0002	0.0000	0.0000	0.0017	0.0028	0.0000
SM29	0.0004	0.0024	0.0072	0.0052	0.0735	0.0077
SM34	0.0163	0.1669	0.1153	0.3337	0.3417	0.2668
SM36	0.1236	0.0788	0.5338	0.1049	0.2210	0.0000
SM37	0.7559	0.0632	0.0316	0.0175	0.3030	0.0013
SM42	0.2747	0.0067	0.0000	0.3748	0.0674	0.0007
SM47*	0.0000	0.0000	0.0000	0.0000	0.0000	0.0001
SM51	0.0328	0.1688	0.0963	0.0050	0.2176	0.0583
SM53	0.0000	0.5993	0.0048	0.4135	0.0008	0.0007
SM55*	0.0000	0.0541	0.0027	0.0000	0.0023	0.0001
SM56*	0.0039	0.0000	0.0000	0.0001	0.0000	0.2673
SM60	0.5995	0.0000	0.0519	0.0959	0.0025	0.2962
Aop943	0.0979	0.8544	0.0638	0.7918	0.7999	0.0339
Am116	0.0000	0.0189	0.5458	0.0000	0.5295	0.1456

*P*-values significant after Bonferroni correction are in bold type. \*indicates locus at which most populations show signs of deviance from HWE.

	OG1	SC1	OG2	SC2	OG3	SC3
SM11	0.068	0.043	0.022	0.076	0.036	0.005
SM14	0.061	0.088	0.058	0.066	0.067	0.110
SM21A	0.066	0.073	0.086	0.080	0.000	0.041
SM22*	0.248	0.212	0.182	0.185	0.151	0.165
SM27*	0.141	0.197	0.186	0.111	0.127	0.220
SM29*	0.142	0.121	0.112	0.128	0.071	0.116
SM34	0.092	0.010	0.005	0.022	0.000	0.036
SM36	0.047	0.039	0.013	0.032	0.040	0.147
SM37	0.000	0.058	0.074	0.037	0.000	0.132
SM42	0.031	0.034	0.070	0.035	0.063	0.104
SM47*	0.158	0.145	0.202	0.203	0.104	0.164
SM51	0.075	0.082	0.092	0.106	0.065	0.096
SM53*	0.184	0.023	0.125	0.038	0.139	0.128
SM55*	0.168	0.108	0.034	0.153	0.105	0.138
SM56*	0.133	0.185	0.185	0.170	0.205	0.067
SM60	0.013	0.150	0.065	0.075	0.134	0.045
Aop943	0.061	0.006	0.000	0.000	0.000	0.061
Am116	0.155	0.087	0.000	0.164	0.000	0.083

Table A5.4 Summary of null allele frequencies for each pair of loci and populations using FREENA.

High ( $\geq 10$  %) frequencies of null alleles are in bold type. \*indicates locus at which most populations show signs of null alleles.

Quebe									
	OG1	SC1	OG2	SC2	OG3	SC3			
Poole	d individua	ıls							
OGI		0.58000	0.57667	0.22333	0.45667	0.11000			
SC1	-0.0001		0.89333	0.57667	0.00333	0.00333			
OG2	-0.0014	-0.0016		0.31667	0.10000	0.04333			
SC2	0.0015	-0.0005	-0.0010		0.03000	0.02333			
OG3	0.0014	0.0039	-0.0002	0.0008		0.00333			
SC3	-0.0020	0.0021	0.0000	0.0009	0.0044				
Mature trees									
OG1		0.07667	0.01667	0.02667	0.03333	0.56000			
SC1	0.0041		0.27333	0.14667	0.05333	0.16333			
OG2	0.0078	0.0005		0.01667	0.51333	0.05000			
SC2	0.0077	-0.0003	0.0061		0.02000	0.14000			
OG3	0.0040	0.0022	-0.0056	0.0069	_	0.02000			
SC3	-00053	-0.0009	-0.0002	0.0084	0.0031				
Saplin	igs								
OG1	_	0.06000	0.16667	0.28333	0.49333	0.18667			
SC1	0.0065		0.21333	0.14667	0.18667	0.17333			
OG2	-0.0022	0.0059		0.28000	0.32667	0.50000			
SC2	-0.0044	0.0113	-0.0013		0.08333	0.04000			
OG3	-0.0007	0.0071	0.0132	0.0051		0.27000			
SC3	-0.0019	0.0059	-0.0006	0.0065	0.0073				
Seedli	ngs								
OG1	_	0.19667	0.21667	0.20000	0.17000	0.07667			
SC1	0.0019		0.89333	0.75000	0.01667	0.19333			
OG2	-0.0035	-0.0053		0.03333	0.06667	0.37333			
SC2	0.0155	0.0034	0.0118		0.15000	0.18333			
OG3	0.0051	0.0063	0.0051	0.0079		0.01000			
SC3	0.0071	0.0135	0.0013	0.0063	0.0209				
Post-h	arvest see	dlings							
OG1		0.29667	0.39667	0.16333	0.32667	0.11667			
SC1	0.0037		0.81667	0.53667	0.24000	0.06667			
OG2	-0.0044	-0.0051		0.03000	0.30667	0.29333			
SC2	0.0179	0.0026	0.0200		0.12333	0.05333			
OG3	0.0021	0.0018	0.0049	0.0074		0.02667			
SC3	0.0031	0.0228	0.0074	0.0138	0.0218				

**Table A5.5** Pairwise population  $F_{ST}$  (below the diagonal) of pooled individuals, mature trees, saplings, seedlings and post-harvest seedlings of sugar maple (*Acer saccharum*) for the 6 sites, in Québec.

SC30.00310.02280.00740.01380.0218—P-values are above the diagonal. Significant values after adjusted nominal level ( $\alpha = 0.05$ ) for multiple comparisons are in bold type.

**Table A5.6** Comparison of mean genetic variability estimates ( $A_R$ ,  $H_O$ ,  $H_E$ ,  $F_{ST}$  and  $F_{IS}$ ) between old-growth (OG) and selection cutting stands (SC) of sugar maple (*Acer saccharum*) in Québec for mature trees (M), saplings (Sa), seedlings (S1), post-harvest seedlings (S2) and pooled individuals (PI) separately, between cohorts (M, Sa and S1, and M, Sa and S2) for PI, OG and SC separately, and between OG and SC for each pair of stands. Analyses were performed using FSTAT and tested for significance using 1,000 permutations.

Genetic Indices		OG	SC	P-value		М	Sa	S1	P-value	М	Sa	S2	P-value	
$A_{\rm R}$	PI	9.011	8.983	0.9880	PI	6.288	6.227	6.299	0.8370	5.369	5.318	5.323	0.8700	
$H_{\rm O}$		0.546	0.528	0.5030		0.594	0.503	0.512	0.0040	0.594	0.503	0.504	0.0010	
$H_{ m E}$		0.689	0.694	0.1960		0.696	0.682	0.691	0.1050	0.696	0.682	0.686	0.1270	
$F_{\rm IS}$		0.208	0.240	0.4040		0.146	0.262	0.260	0.0030	0.146	0.262	0.266	0.0010	
$F_{\rm ST}$		-0.000	0.001	0.7340		0.002	0.004	0.006	0.5360	0.002	0.004	0.008	0.2890	
$A_{\rm R}$	М	6.557	6.310	0.2900	OG	6.401	6.131	6.352	0.2260	5.432	5.236	5.438	0.3150	
$H_{\rm O}$		0.610	0.577	0.1780		0.610	0.502	0.519	0.0150	0.610	0.502	0.509	0.0160	
$H_{ m E}$		0.694	0.698	0.6100		0.694	0.679	0.688	0.3500	0.694	0.679	0.683	0.3840	
$F_{\rm IS}$		0.121	0.173	0.1780		0.121	0.261	0.245	0.0190	0.121	0.261	0.255	0.0180	
$F_{\rm ST}$		0.002	0.002	0.9260		0.002	0.004	0.002	0.9550	0.002	0.004	0.000	0.8570	
$A_{\mathrm{R}}$	Sa	6.131	6.323	0.5330	SC	6.175	6.323	6.245	0.7090	5.305	5.399	5.208	0.4570	
$H_{\rm O}$		0.502	0.503	1.0000		0.577	0.503	0.504	0.1750	0.577	0.503	0.498	0.1570	
$H_{ m E}$		0.679	0.684	0.8140		0.698	0.684	0.695	0.3890	0.698	0.684	0.688	0.4260	
$F_{\rm IS}$		0.261	0.264	0.8320		0.173	0.264	0.275	0.1780	0.173	0.264	0.277	0.1920	
$F_{\rm ST}$		0.004	0.008	0.3640	_	0.002	0.008	0.008	0.5810	0.002	0.008	0.013	0.1840	
$A_{\mathrm{R}}$	<b>S</b> 1	6.352	6.245	0.4170										
$H_{\rm O}$		0.519	0.504	0.7200										
$H_{ m E}$		0.688	0.695	0.7080										
$F_{\rm IS}$		0.245	0.275	0.6200										
$F_{\rm ST}$		0.002	0.008	0.4370		OG1	SC1	P-value	OG2	SC2	P-value	OG3	SC3	P-value
A <sub>R</sub>	S2	5.438	5.208	0.2460	PI	6.250	6.317	0.7170	6.281	6.193	0.6300	6.344	6.233	0.5120
$H_{\rm O}$		0.509	0.498	0.7100		0.507	0.529	0.6480	0.569	0.543	0.5960	0.577	0.511	0.1240
$H_{ m E}$		0.683	0.688	0.7170		0.684	0.698	0.1730	0.686	0.689	0.8120	0.690	0.690	0.9840
$F_{\rm IS}$		0.255	0.277	0.5100		0.259	0.242	0.7930	0.171	0.212	0.5400	0.164	0.260	0.1000
$F_{\rm ST}$		0.000	0.013	0.1470		0.003	0.002	0.8250	0.002	0.011	0.2000	0.006	0.000	0.3020

 $A_{\rm R}$ , mean allelic richness;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm ST}$ , mean pairwise  $F_{\rm ST}$ ;  $F_{\rm IS}$ , inbreeding coefficient. Significant values ( $\alpha = 0.05$ ) are in bold type.

Data used	Response	Explanatory variables	Num.	Den.	F	Р
	variable		d.f.	d.f.		
M, Sa and S1	$A_{\rm R}$	forest type	1	289	0.177	0.6739
		pair	2	289	0.073	0.9294
		cohort	2	289	0.164	0.8492
		forest type × pair	2	289	0.262	0.7694
		forest type × cohort	2	289	1.252	0.2874
		pair × cohort	4	289	0.717	0.5812
		forest type × pair × cohort	4	289	1.461	0.2141
	H <sub>0</sub>	forest type	1	289	1.447	0.2299
		pair	2	289	2.155	0.1178
		cohort	2	289	19.962	<0.001
		forest type × pair	2	289	3.590	0.0288
		forest type × cohort	2	289	0.550	0.5778
		pair × cohort	4	289	1.416	0.2288
		forest type $\times$ pair $\times$ cohort	4	289	0.671	0.6128
	$H_{\rm E}$	forest type	1	289	0.430	0.5127
		pair	2	289	0.021	0.9792
		cohort	2	289	1.228	0.2945
		forest type × pair	2	289	0.342	0.7109
		forest type × cohort	2	289	0.023	0.9770
		pair × cohort	4	289	1.100	0.3570
		forest type $\times$ pair $\times$ cohort	4	289	0.443	0.7775
	$F_{\rm IS}$	forest type	1	289	1.890	0.1703
		pair	2	289	1.365	0.2570
		cohort	2	289	17.347	<0.0001
		forest type × pair	2	289	1.879	0.1545
		forest type $\times$ cohort	2	289	0.660	0.5177
		pair × cohort	4	289	1.239	0.2945
		forest type $\times$ pair $\times$ cohort	4	289	0.361	0.8366
M, Sa and S2	$A_{\rm R}$	forest type	1	289	0.516	0.4730
		pair	2	289	0.374	0.6882
		cohort	2	289	0.129	0.8787
		forest type × pair	2	289	0.876	0.4176
		forest type $\times$ cohort	2	289	1.718	0.1813
		pair × cohort	4	289	0.471	0.7568
		forest type $\times$ pair $\times$ cohort	4	289	1.498	0.2028

**Table A5.7** Results of linear mixed-effects models for genetic variability estimates (n = 432) in Québec, prior to model simplification.

variable d.f. d.f.		
M, Sa and S2 $H_0$ forest type 1 289	1.385	0.2402
pair 2 289	2.778	0.0638
cohort 2 289	19.528	<0.0001
forest type $\times$ pair 2 289	3.314	0.0378
forest type $\times$ cohort 2 289	0.506	0.6036
pair × cohort 4 289	0.901	0.4636
forest type $\times$ pair $\times$ cohort 4 289	0.567	0.6868
$H_{\rm E}$ forest type 1 289	0.173	0.6781
pair 2 289	0.177	0.8379
cohort 2 289	1.108	0.3317
forest type $\times$ pair 2 289	0.621	0.5381
forest type $\times$ cohort 2 289	0.004	0.9963
pair $\times$ cohort 4 289	0.622	0.6474
forest type $\times$ pair $\times$ cohort 4 289	0.685	0.6032
$F_{\rm IS}$ forest type 1 289	1.790	0.1819
pair 2 289	1.802	0.1668
cohort 2 289	16.291	<0.0001
forest type $\times$ pair 2 289	1.379	0.2536
forest type $\times$ cohort 2 289	0.610	0.5440
pair × cohort 4 289	0.791	0.5320
forest type $\times$ pair $\times$ cohort 4 289	0.195	0.9409

 Table A5.7 (to continued)

Num. d.f., numerator degrees of freedom; den. d.f., denominator degrees of freedom; M, mature sugar maple; Sa, sugar maple sapling; S1, sugar maple seedling; S2, post-harvest sugar maple seedling.  $A_{\rm R}$ , mean allelic richness;  $H_{\rm O}$ , mean observed heterozygosity;  $H_{\rm E}$ , mean expected heterozygosity;  $F_{\rm ST}$ , mean pairwise  $F_{\rm ST}$ ;  $F_{\rm IS}$ , inbreeding coefficient. Significant values at  $\alpha = 0.05$  are in bold type and at  $\alpha = 0.10$  in italics.

Populations	Cohorts	$A_{\mathrm{T}}$	С	Ι	L0.01	R0.01	L0.05	R0.05
OG1	М	128	2 (2%)	20 (16%)	106 (83%)	0 (0%)	64 (50%)	42 (33%)
	Sa	119	1 (1%)	26 (22%)	92 (77%)	0 (0%)	54 (45%)	38 (32%)
	S1	128	1 (1%)	22 (17%)	105 (82%)	0 (0%)	67 (52%)	38 (30%)
	PI	169	1 (1%)	18 (11%)	122 (72%)	28 (17%)	58 (34%)	92 (54%)
	S2	118	2 (2%)	20 (17%)	96 (81%)	0 (0%)	62 (53%)	34 (29%)
SC1	М	129	0 (0%)	24 (19%)	105 (81%)	0 (0%)	66 (51%)	39 (30%)
	Sa	124	1 (1%)	23 (19%)	100 (81%)	0 (0%)	59 (48%)	41 (33%)
	S1	124	1 (1%)	20 (16%)	103 (83%)	0 (0%)	64 (52%)	39 (31%)
	PI	171	0 (0%)	22 (13%)	116 (68%)	33 (19%)	58 (34%)	91 (53%)
	S2	114	2 (2%)	23 (20%)	89 (78%)	0 (0%)	48 (42%)	41 (36%)
OG2	М	125	0 (0%)	22 (18%)	103 (82%)	0 (0%)	64 (51%)	39 (31%)
	Sa	125	1 (1%)	23 (18%)	101 (81%)	0 (0%)	66 (53%)	35 (28%)
	S1	126	0 (0%)	22 (17%)	104 (83%)	0 (0%)	62 (49%)	42 (33%)
	PI	171	0 (0%)	23 (13%)	114 (67%)	34 (20%)	55 (32%)	93 (54%)
	S2	120	2 (2%)	21 (18%)	97 (81%)	0 (0%)	53 (44%)	44 (37%)
SC2	М	117	1 (1%)	25 (21%)	91 (78%)	0 (0%)	63 (54%)	28 (24%)
	Sa	120	2 (2%)	20 (17%)	98 (82%)	0 (0%)	58 (48%)	40 (33%)
	S1	126	0 (0%)	23 (18%)	103 (82%)	0 (0%)	64 (51%)	39 (31%)
	PI	171	0 (0%)	22 (13%)	113 (66%)	36 (21%)	51 (30%)	98 (57%)
	S2	109	0 (0%)	27 (25%)	82 (75%)	0 (0%)	47 (43%)	35 (32%)
OG3	М	137	1 (1%)	21 (15%)	115 (84%)	0 (0%)	48 (35%)	67 (49%)
	Sa	121	1 (1%)	24 (20%)	96 (79%)	0 (0%)	58 (48%)	38 (31%)
	S1	123	1 (1%)	21 (17%)	101 (82%)	0 (0%)	60 (49%)	41 (33%)
	PI	171	0 (0%)	22 (13%)	113 (66%)	36 (21%)	52 (30%)	97 (57%)
	S2	108	1 (1%)	19 (18%)	88 (81%)	0 (0%)	58 (54%)	30 (28%)
SC3	М	117	1 (1%)	21 (18%)	95 (81%)	0 (0%)	58 (50%)	37 (32%)
	Sa	134	2 (1%)	18 (13%)	114 (85%)	0 (0%)	73 (54%)	41 (31%)
	S1	122	1 (1%)	24 (20%)	97 (80%)	0 (0%)	57 (47%)	40 (33%)
	PI	166	1 (1%)	18 (11%)	115 (69%)	32 (19%)	59 (36%)	88 (53%)
	S2	106	2 (2%)	21 (20%)	83 (78%)	0 (0%)	52 (49%)	31 (29%)

**Table A5.8** Number of alleles per classes of frequency, per cohorts of the populations and grouping populations by pairs and forest types.

 $A_{\rm T}$ , total number of alleles; C, common  $f \ge 0.75$ ; I, intermediate  $0.75 > f \ge 0.25$ ; L0.01, low  $0.25 > f \ge 0.01$ ; R0.01, rare f < 0.01; L0.05, low  $0.25 > f \ge 0.05$ ; R0.05, rare f < 0.05; M, mature sugar maple; Sa, sugar maple sapling; S1, sugar maple seedling; PI, pooled individuals; S2, post-harvest sugar maple seedling.

## BIBLIOGRAPHIE

Abrams, M. (1992) Fire and the development of oak forests. *Bioscience*, 42, 346–353.

- Alberto, F.J., Aitken, S.N., Alía, R., González-Martínez, S.C., Hänninen, H., Kremer, A., Lefèvre, F., Lenormand, T., Yeaman, S., Whetten, R. & Savolainen, O. (2013) Potential for evolutionary responses to climate change – evidence from tree populations. *Global Change Biology*, **19**, 1645–1661.
- Ali, A.A., Asselin, H., Larouche, A.C., Bergeron, Y., Carcaillet, C. & Richard, P.J.H. (2008) Changes in fire regime explain the Holocene rise and fall of *Abies balsamea* in the coniferous forests of western Québec, Canada. *The Holocene*, 18, 693–703.
- Angers, V.A., Messier, C., Beaudet, M. & Leduc, A. (2005) Comparing composition and structure in old-growth and harvested (selection and diameter-limit cuts) northern hardwood stands in Quebec. *Forest Ecology and Management*, **217**, 275–293.
- APG II (the Angiosperm Phylogeny Group II) (2003) An update of the angiosperm phylogeny group classification for the orders and families of flowering plants: APG II. *Botanical Journal of the Linnean Society*, **141**, 399–436.
- Avise, J.C., Arnold, J., Ball, R.M., Bermingham, E., Lamb, T., Neigel, J.E., Reeb, C.A. & Saunders, N.C. (1987) Intraspecific phylogeography: the mitochondrial DNA bridge between population genetics and systematics. *Annual Review of Ecology and Systematics*, 18, 489–522.
- Bai, W.-N., Liao, W.-J. & Zhang, D.-Y. (2010) Nuclear and chloroplast DNA phylogeography reveal two refuge areas with asymmetrical gene flow in a temperate walnut tree from East Asia. *New Phytologist*, **188**, 892–901.
- Ballal, S.R. (1994) Apparent gene flow and genetic structure of *Acer saccharum* subpopulations in forest fragments. *Canadian journal of botany*, **72**, 1311–1315.
- Bates, D., Maechler, M. & Bolker, B. (2011) *lme4: linear mixed-effects models using S4 classes*. R package version 0.999375-38. Available at: <u>http://cran.r-project.org/web/packages/lme4/index.html</u>.
- Baucom, R.S., Estill, J.C. & Cruzan, M.B. (2005) The effect of deforestation on the genetic diversity and structure in *Acer saccharum* (Marsh): evidence for the loss

and restructuring of genetic variation in a natural system. *Conservation Genetics*, **6**, 39–50.

- Beerli, P. (2006) Comparison of bayesian and maximum-likelihood inference of population genetic parameters. *Bioinformatics*, 22, 341–345.
- Beerli, P. & Felsenstein, J. (2001) Maximum likelihood estimation of a migration matrix and effective population sizes in n subpopulations by using a coalescent approach. *Proceedings of the National Academy of Sciences*, **98**, 4563–4568.
- Belletti, P., Monteleone, I. & Ferrazzini, D. (2007) Genetic variability at allozyme markers in sycamore (*Acer pseudoplatanus*) populations from northwestern Italy. *Canadian Journal of Forest Research*, **37**, 395–403.
- Bennett, K.D. (1987) Holocene history of forest trees in southern Ontario. *Canadian Journal of Botany*, **65**, 1792–1801.
- Bergeron, Y., Cyr, D., Girardin, M.P. & Carcaillet, C. (2010) Will climate change drive 21st century burn rates in Canadian boreal forest outside of its natural variability: collating global climate model experiments with sedimentary charcoal data. *International Journal of Wildland Fire*, **19**, 1127–1139.
- Bernier, B., Pare, D. & Brazeauu, M. (1989) Natural stresses, nutrient imbalances and forest decline in southeastern Québec. *Water, Air, & Soil Pollution*, **48**, 239–250.
- Bouchard, A. & Domon, G. (1997) The transformations of the natural landscapes of the Haut-Saint-Laurent (Québec) and their implications on future resource management. *Landscape and Urban Planning*, **37**, 99–107.
- Boucher, Y., Arseneault, D. & Sirois, L. (2006) Logging-induced change (1930–2002) of a preindustrial landscape at the northern range limit of northern hardwoods, eastern Canada. *Canadian Journal of Forest Research*, **36**, 505–517.
- Boucher, Y., Arseneault, D., Sirois, L. & Blais, L. (2009) Logging pattern and landscape changes over the last century at the boreal and deciduous forest transition in Eastern Canada. *Landscape Ecology*, **24**, 171–184.
- Bowyer, J.L., Shmulsky, R. & Haygreen, J.G. (2005) *Le bois et ses usages*. Centre collégial de développement de matériel didactique (CCDMD), Montréal, Québec, Canada. 528pp.
- Braun, E.L. (1950) Deciduous forests of eastern North America. Afner press, New York, New York, USA. 596pp.

- Brodie, J., Post, E. & Laurance, W.F. (2012) Climate change and tropical biodiversity: a new focus. *Trends in Ecology & Evolution*, **27**, 145–150.
- Buchert, G.P., Rajora, O.P., Hood, J.V. & Dancik, B.P. (1997) Effects of harvesting on genetic diversity in old-growth eastern white pine in Ontario, Canada. *Conservation Biology*, **11**, 747–758.
- Buckley, L.B. & Kingsolver, J.G. (2012) Functional and phylogenetic approaches to forecasting species' responses to climate change. *Annual Review of Ecology, Evolution, and Systematics*, **43**, 205–226.
- Burnham, K.P. & Anderson, D.R. (2004) Multimodel inference: understanding AIC and BIC in model selection. *Sociological Methods and Research*, **33**, 261–304.
- Bush, R. & Smouse, P. (1992) Evidence for the adaptive significance of allozymes in forest trees. *New Forests*, **6**, 179–196.
- Caccianiga, M. & Payette, S. (2006) Recent advance of white spruce (*Picea glauca*) in the coastal tundra of the eastern shore of Hudson Bay (Québec, Canada). *Journal of Biogeography*, **33**, 2120–2135.
- Cain, M.L., Milligan, B.G. & Strand, A.E. (2000) Long-distance seed dispersal in plant populations. *American Journal of Botany*, **87**, 1217–1227.
- Callahan, C.M., Rowe, C.A., Ryel, R.J., Shaw, J.D., Madritch, M.D. & Mock, K.E. (2013) Continental-scale assessment of genetic diversity and population structure in quaking aspen (*Populus tremuloides*). *Journal of Biogeography*, **40**, 1780– 1791.
- Carcaillet, C. & Richard, P.J.H. (2000) Holocene changes in seasonal precipitation highlighted by fire incidence in eastern Canada. *Climate Dynamics*, **16**, 549–559.
- Carl, C.M.J. (1983) Stratification of sugar maple seeds. *Tree Planters' Notes*, **34**, 25–27.
- Caspersen, J.P. & Saprunoff, M. (2005) Seedling recruitment in a northern temperate forest: the relative importance of supply and establishment limitation. *Canadian Journal of Forest Research*, **35**, 978–989.
- Chapuis, M.-P. & Estoup, A. (2007) Microsatellite null alleles and estimation of population differentiation. *Molecular Biology and Evolution*, **24**, 621–631.

- Charlesworth, D. & Willis, J.H. (2009) The genetics of inbreeding depression. *Nature Reviews Genetics*, **10**, 783–796.
- Christensen, J.H., Hewitson, B., Busuioc, A., Chen, A., Gao, X., Held, I., Jones, R., Kolli, R.K., Kwon, W-T., Laprise, R., Maga a Rueda, V., Mearns, L., Menéndez, C.G., R is n en, J., Rinke, A., Sarr, A. Whetton, P. (2007) Regional climate projections. *Climate change 2007: the physical science basis. Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change* (ed. by S. Solomon, D. Qin, M. Manning, Z. Chen, M. Marquis, K.B. Averyt, M. Tignor and H.L. Miller), pp. 847–940. Cambridge University Press, Cambridge, UK.
- Chuine, I. & Beaubien, E. G. (2001) Phenology is a major determinant of tree species range. *Ecology Letters*, **4**, 500–510.
- Cleavitt, N.L., Fahey, T.J. & Battles, J.J. (2011) Regeneration ecology of sugar maple (*Acer saccharum*): seedling survival in relation to nutrition, site factors, and damage by insects and pathogens. *Canadian Journal of Forest Research*, **41**, 235–244.
- Cornuet, J.M. & Luikart, G. (1996) Description and power analysis of two tests for detecting recent population bottlenecks from allele frequency data. *Genetics*, **144**, 2001–2014.
- Darlington, C.D. & Wylie, A.P. (1955) Chromosome atlas of flowering plants. George Allen and Unwin LTD, London, UK.
- Davey, J.W., Hohenlohe, P.A., Etter, P.D., Boone, J.Q., Catchen, J.M. & Blaxter, M.L. (2011) Genome-wide genetic marker discovery and genotyping using nextgeneration sequencing. *Nature Reviews Genetics*, **12**, 499–510.
- Demesure, B., Sodzi, N. & Petit, R.J. (1995) A set of universal primers for amplification of polymorphic non-coding regions of mitochondrial and chloroplast DNA in plants. *Molecular Ecology*, **4**, 129–131.
- Dempster, A.P., Laird, N.M. & Rubin, D.B. (1977) Maximum likelihood from incomplete data via the EM algorithm. *Journal of the Royal Statistical Society*. *Series B (Methodological)*, **39**, 1–38.
- Diochon, A., Rigg, L.S., Goldblum, D. & Polans, N. (2003) Regeneration dynamics and genetic variability of sugar maple (*Acer saccharum* Marsh.) seedlings at the species northern growth limit, Lake Superior provincial park, Canada. *Physical Geography*, 24, 399–413.

- Duchesne, L., Ouimet, R., Moore, J.-D. & Paquin, R. (2005) Changes in structure and composition of maple-beech stands following sugar maple decline in Québec, Canada. *Forest Ecology and Management*, **208**, 223–236.
- Dupuis, S., Arseneault, D. & Sirois, L. (2011) Change from pre-settlement to presentday forest composition reconstructed from early land survey records in eastern Québec, Canada. *Journal of Vegetation Science*, 22, 564–575.
- Earl, D. & Vonholdt, B. (2012) Structure harvester: a website and program for visualizing structure output and implementing the evanno method. *Conservation Genetics Resources*, **4**, 359–361.
- Eckert, A.J., Bower, A.D., Wegrzyn, J.L., Pande, B., Jermstad, K.D., Krutovsky, K.V., St. Clair, J.B. & Neale, D.B. (2009) Association genetics of coastal douglas fir (*Pseudotsuga menziesii* var. *Menziesii*, Pinaceae). I. Cold-hardiness related traits. *Genetics*, **182**, 1289–1302.
- Eckert, C.G., Samis, K.E. & Lougheed, S.C. (2008) Genetic variation across species' geographical ranges: the central-marginal hypothesis and beyond. *Molecular Ecology*, **17**, 1170–1188.
- El Mousadik, A. & Petit, R.J. (1996) Chloroplast DNA phylogeography of the argan tree of Morocco. *Molecular Ecology*, **5**, 547–555.
- Ellegren, H. (2004) Microsatellites: simple sequences with complex evolution. *Nature Reviews Genetics*, **5**, 435–445.
- Emerson, K.J., Merz, C.R., Catchen, J.M., Hohenlohe, P.A., Cresko, W.A., Bradshaw, W.E. & Holzapfel, C.M. (2010) Resolving postglacial phylogeography using high-throughput sequencing. *Proceedings of the National Academy of Sciences*, **107**, 16196–16200.
- Ennos, R.A. (1994) Estimating the relative rates of pollen and seed migration among plant populations. *Heredity*, **72**, 250–259.
- Environment Canada (2009) *Canadian climate normals or averages 1971–2000*. Available at: http://climate.weatheroffice.gc.ca/climate\_normals/index\_e.html (accessed 4 February 2010).
- Evanno, G., Regnaut, S. & Goudet, J. (2005) Detecting the number of clusters of individuals using the software structure: a simulation study. *Molecular Ecology*, 14, 2611–2620.

- Fageria, M.S. & Rajora, O.P. (2013) Effects of harvesting of increasing intensities on genetic diversity and population structure of white spruce. *Evolutionary Applications*, **6**, 778–794.
- Farrar, J.L. (1996) Les arbres du canada, Les Editions Fides.
- Finkeldey, R. & Ziehe, M. (2004) Genetic implications of silvicultural regimes. *Forest Ecology and Management*, **197**, 231–244.
- Fischer, A., Marshall, P. & Camp, A. (2013) Disturbances in deciduous temperate forest ecosystems of the northern Hemisphere: Their effects on both recent and future forest development. *Biodiversity and Conservation*, **22**, 1863–1893.
- Fontaine, C., Lovett, P.N., Sanou, H., Maley, J. & Bouvet, J-M (2004) Genetic diversity of the shea tree (*Vitellaria paradoxa* C.F. Gaertn), detected by RAPD and chloroplast microsatellite markers. *Heredity*, **93**, 639–648.
- Foré, S.A., Hickey, R.J., Guttman, S.I. & Vankat, J.L. (1992a) Temporal differences in genetic diversity and structure of sugar maple in an old-growth forest. *Canadian Journal of Forest Research*, **22**, 1504–1509.
- Foré, S.A., Hickey, R.J., Vankat, J.L., Guttman, S.I. & Schaefer, R.L. (1992b) Genetic structure after forest fragmentation: a landscape ecology perspective on *Acer saccharum. Canadian Journal of Botany*, **70**, 1659–1668.
- Forget, E., Nolet, P., Doyon, F.D.R., Delagrange, S. & Jardon, Y. (2007) Ten-year response of northern hardwood stands to commercial selection cutting in southern Quebec, Canada. *Forest Ecology and Management*, 242, 764–775.
- Gabriel, W.J. (1968) Dichogamy in Acer saccharum. Botanical Gazette, **129**, 334–338.
- Gabriel, W.J. & Garrett, P.W. (1967) Reproductive behaviour in sugar maple: selfcompatibility, cross-compatibility, agamospermy, and agamocarpy. *Silvae Genetica*, **16**, 165–168.
- Gabriel, W.J. & Garrett, P.W. (1984) Pollen vectors in sugar maple (*Acer saccharum*). Canadian Journal of Botany, **62**, 2889–2890.
- Gamache, I., Jaramillo-Correa, J.P., Payette, S. & Bousquet, J. (2003) Diverging patterns of mitochondrial and nuclear DNA diversity in subarctic black spruce: imprint of a founder effect associated with postglacial colonization. *Molecular Ecology*, **12**, 891–901.

- Gapare, W.J., Aitken, S.N. & Ritland, C.E. (2005) Genetic diversity of core and peripheral sitka spruce (*Picea sitchensis* (Bong.) Carr) populations: implications for conservation of widespread species. *Biological Conservation*, 123, 113–123.
- Garrett, P.W. & Graber, R.E. (1995) Sugar maple seed production in northern New Hampshire. Research Paper NE-697. USDA Forest Service, Northeastern Forest Experiment Station, Radnor, PA.
- Garza, J.C. & Williamson, E.G. (2001) Detection of reduction in population size using data from microsatellite loci. *Molecular Ecology*, **10**, 305–318.
- Gaston, K.J. (2009) Geographic range limits: achieving synthesis. *Proceedings of the Royal Society B: Biological Sciences*, **276**, 1395–1406.
- Geburek, T. (1993) Are genes randomly distributed over space in mature populations of sugar maple (*Acer saccharum* Marsh.)? *Annals of Botany*, **71**, 217–222.
- Geburek, T.H. & Knowles, P. (1992) Ecological-genetic investigations in environmentally stressed mature sugar maple (*Acer saccharum* Marsh.) populations. *Water, Air, & Soil Pollution*, **62**, 261–268.
- Godman, R.M., Yawney, H.W. & Tubbs, C.H. (1990) Acer saccharum Marsh. Sugar maple. Silvics of North America, Vol. 2, Hardwoods (tech. coords. R.M. Burns and B.H. Honkala), pp. 78–91. Agricultural Handbook 654. USDA Forest Service, Washington, DC.
- Goldblum, D. & Rigg, L.S. (2002) Age structure and regeneration dynamics of sugar maple at the deciduous/boreal forest ecotone, Ontario, Canada. *Physical Geography*, **23**, 115–129.
- Goldblum, D. & Rigg, L.S. (2005) Tree growth response to climate change at the deciduous-boreal forest ecotone, Ontario, Canada. *Canadian Journal of Forest Research*, **35**, 2709–2718.
- Goldblum, D. & Rigg, L.S. (2009) Simulating climate change (temperature and soil moisture) in a mixed-deciduous forest, Ontario, Canada. World Academy of Science, Engineering and Technology, 57, 103–107.
- González-Martínez, S.C., Gil, L. & Alía, R. (2005) Genetic diversity estimates of pinus pinaster in the iberian peninsula: A comparison of allozymes and quantitative traits. *Forest Systems*, **14**, 3–12.

- Goudet, J. (2001) FSTAT, a program to estimate and test gene diversities and fixation indices (version 2.9.3). Available at: http://www2.unil.ch/popgen/softwares/fstat.html (accessed 17 June 2014).
- Graignic, N., Tremblay, F. & Bergeron, Y. (2013) Development of polymorphic nuclear microsatellite markers in sugar maple (*Acer saccharum* Marsh.) using cross-species transfer and SSR-enriched shotgun pyrosequencing. *Conservation Genetics Resources*, 5, 845–848.
- Graignic, N., Tremblay, F. & Bergeron, Y. (2014) Geographical variation in reproductive capacity of sugar maple (*Acer saccharum* Marshall) northern peripheral populations. *Journal of Biogeography*, **41**, 145–157.
- Gravel, D., Beaudet, M. & Messier, C. (2011) Sapling age structure and growth series reveal a shift in recruitment dynamics of sugar maple and American beech over the last 40 years. *Canadian Journal of Forest Research*, **41**, 873–880.
- Guillemette, F., Gauthier, M.-M., Lambert, M.-C. & Bédard, S. (2013) Effets réels décennaux des coupes de jardinage pratiquées de 1995 à 1999 dans un contexte opérationnel. Mémoire de recherche forestière n°168. Direction de la recherche forestière.
- Gunter, L.E., Tuskan, G.A., Gunderson, C.A. & Norby, R.J. (2000) Genetic variation and spatial structure in sugar maple (*Acer saccharum* Marsh.) and implications for predicted global-scale environmental change. *Global Change Biology*, 6, 335–344.
- Guo, Q. (2012) Incorporating latitudinal and central-marginal trends in assessing genetic variation across species ranges. *Molecular Ecology*, **21**, 5396–5403.
- Hampe, A. & Petit, J.R. (2005) Conserving biodiversity under climate change: the rear edge matters. *Ecology Letters*, **8**, 461–467.
- Hamrick, J.L. (2004) Response of forest trees to global environmental changes. *Forest Ecology and Management*, **197**, 323–335.
- Hamrick, J.L. & Godt, M.J.W. (1996) Effects of life history traits on genetic diversity in plant species. *Philosophical Transactions of the Royal Society B: Biological Sciences*, **351**, 1291–1298.
- Hamrick, J.L., Godt, M.J.W. & Sherman-Broyles, S.L. (1992) Factors influencing levels of genetic diversity in woody plant species. *New Forests*, **6**, 95–124.

- Harrell Jr, F.E. (2013) Hmisc: Harrell miscellaneous. R package version 3.12-2. Available at: http://www.cran.r-project.org/web/packages/Hmisc/index.html.
- Hawley, G.J. & Dehayes, D.H. (1994) Genetic diversity and population structure of red spruce (*Picea rubens*). *Canadian Journal of Botany*, **72**, 1778–1786.
- Hegerl, G.C., Zwiers, F.W., Braconnot, P., Gillett, N.P., Luo, Y., Marengo Orsini, J.A., Nicholls, N., Penner, J.E. & Stott, P.A. (2007) Understanding and attributing climate change. *Climate change 2007: the physical science basis*. *Contribution of Working Group I to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change* (ed. by S. Solomon, D. Qin, M. Manning, Z. Chen, M. Marquis, K.B. Averyt, M. Tignor and H.L. Miller), pp.663–745. Cambridge University Press, Cambridge, UK.
- Hett, J.M. & Loucks, O.L. (1971) Sugar maple (*Acer saccharum* Marsh.) seedling mortality. *Journal of Ecology*, **59**, 507–520.
- Hoban, S.M., Borkowski, D.S., Brosi, S.L., Mccleary, T.S., Thompson, L.M., Mclachlan, J.S., Pereira, M.A., Schlarbaum, S.E. & Romero-Severson, J. (2010) Range-wide distribution of genetic diversity in the North American tree juglans cinerea: a product of range shifts, not ecological marginality or recent population decline. *Molecular Ecology*, **19**, 4876–4891.
- Holliday, J.A., Ritland, K. & Aitken, S.N. (2010) Widespread, ecologically relevant genetic markers developed from association mapping of climate-related traits in sitka spruce (*Picea sitchensis*). *New Phytologist*, **188**, 501–514.
- Hosie, R.C. (1980) Arbres indigènes du Canada. Service Canadien des Forêts, Montréal, Québec, CA.
- Houle, G. (1992) Spatial relationship between seed and seedling abundance and mortality in a deciduous forest of north-eastern North America. *Journal of Ecology*, **80**, 99–108.
- Houle, G. (1999) Mast seeding in *Abies balsamea*, *Acer saccharum* and *Betula alleghaniensis* in an old growth, cold temperate forest of north-eastern North America. *Journal of Ecology*, **87**, 413–422.
- Houston, D.B. & Houston, D.R. (1994) Variation in American beech (*Fagus grandifolia* Ehrh.): isozyme analysis of genetic structure in selected stands. *Silvae Genetica*, **43**, 277–284.

- Houston, D.B. & Houston, D.R. (2000) Allozyme genetic diversity among *Fagus* grandifolia trees resistant or susceptible to beech bark disease in natural populations. *Canadian Journal of Forest Research*, **30**, 778–789.
- Howe, G.T., Aitken, S.N., Neale, D.B., Jermstad, K.D., Wheeler, N.C. & Chen, T.H. (2003) From genotype to phenotype: unraveling the complexities of cold adaptation in forest trees. *Canadian Journal of Botany*, **81**, 1247–1266.
- Hsia, J.F. (2009) Postdispersal sugar maple (*Acer saccharum*) seed predation by small mammals in a northern hardwood forest. *American Midland Naturalist*, **162**, 213–223.
- Huang, H., Dane, F. & Kubisiak, T. (1998) Allozyme and RAPD analysis of the genetic diversity and geographic variation in wild populations of the *American chestnut (Fagaceae). American Journal of Botany*, **85**, 1013–1021.
- Hubisz, M.J., Falush, D., Stephens, M. & Pritchard, J.K. (2009) Inferring weak population structure with the assistance of sample group information. *Molecular Ecology Resources*, 9, 1322–1332.
- Hyun, J.O., Rajora, O.P. & Zsuffa, L. (1987) Genetic variation in *Trembling aspen* in Ontario based on isozyme studies. *Canadian Journal of Forest Research*, **17**, 1134–1138.
- Iddrisu, M.N. & Ritland, K. (2004) Genetic variation, population structure, and mating system in bigleaf maple (*Acer macrophyllum* Pursh). *Canadian Journal of Botany*, **82**, 1817–1825.
- Iverson, L.R., Schwartz, M.W. & Prasad, A.M. (2004) How fast and far might tree species migrate in the eastern United States due to climate change? *Global Ecology and Biogeography*, 13, 209–219.
- Iverson, L.R., Prasad, A.M., Matthews, S.N. & Peters, M. (2008) Estimating potential habitat for 134 eastern US tree species under six climate scenarios. *Forest Ecology and Management*, 254, 390–406.
- Jackson, S.T., Webb, R.S., Anderson, K.H., Overpeck, J.T., Webb III, T., Williams, J.W. & Hansen, B.C.S. (2000) Vegetation and environment in eastern North America during the Last Glacial Maximum. *Quaternary Science Reviews*, 19, 489–508.
- Jadwiszczak, K.A., Banaszek, A., Jabłońska, E. Sozinov, O.V. (2011) Could betula humilis have survived the last glaciation at a current margin of its

distribution? Testing the hypothesis of a glacial refugium using nuclear microsatellites. *Plant Systematics and Evolution*, **297**, 147–156.

- Janerette, C.A. (1979) The effects of water soaking on the germination of sugar maple seeds. *Seed Science and Technology*, **7**, 341–346.
- Johnson, W.C. (1988) Estimating dispersibility of *Acer*, *Fraxinus* and *Tilia* in fragmented landscapes from patterns of seedling establishment. *Landscape Ecology*, **1**, 175–187.
- Jump, A.S., Woodward, F.I. & Burke, T. (2003) Cirsium species show disparity in patterns of genetic variation at their range-edge, despite similar patterns of reproduction and isolation. *New Phytologist*, **160**, 359–370.
- Keller, S.R., Levsen, N., Olson, M.S. & Tiffin, P. (2012) Local adaptation in the flowering-time gene network of balsam poplar, *Populus balsamifera* L. *Molecular Biology and Evolution*, 29, 3143–3152.
- Kellman, M. (2004) Sugar maple (*Acer saccharum* Marsh.) establishment in boreal forest: results of a transplantation experiment. *Journal of Biogeography*, **31**, 1515–1522.
- Kelly, D. (1994) The evolutionary ecology of mast seeding. *Trends in Ecology and Evolution*, **9**, 465–470.
- Kelly, D. & Sork, V.L. (2002) Mast seeding in perennial plants: why, how, where? *Annual Review of Ecology and Systematics*, **33**, 427–447.
- Kikuchi, S. & Shibata, M. (2008) Development of polymorphic microsatellite markers in *Acer mono* maxim. *Molecular Ecology Notes*, **8**, 339–341.
- Kikuchi, S., Shibata, M., Tanaka, H., Yoshimaru, H. & Niiyama, K. (2009) Analysis of the disassortative mating pattern in a heterodichogamous plant, *Acer mono* Maxim. using microsatellite markers. *Plant Ecology*, **204**, 43–54.
- Kramer, A.T., Ison, J.L., Ashley, M.V. & Howe, H.F. (2008) The paradox of forest fragmentation genetics. *Conservation Biology*, **22**, 878–885.
- Kremer, A., Ronce, O., Robledo-Arnuncio, J.J., Guillaume, F., Bohrer, G., Nathan, R., Bridle, J.R., Gomulkiewicz, R., Klein, E.K., Ritland, K., Kuparinen, A., Gerber, S. & Schueler, S. (2012) Long-distance gene flow and adaptation of forest trees to rapid climate change. *Ecology Letters*, 15, 378–392.

- Kriebel, H.B. (1957) *Patterns of genetics variation in sugar maple*. Research Bulletin 791. Ohio Agricultural Experiment Station, Wooster, OH.
- Kriebel, H.B. & Gabriel, W.J. (1969) *Genetics of sugar maple*. Research Paper WO-7. USDA Forest Service, Washington, DC.
- Kwit, M.C., Rigg, L.S. & Goldblum, D. (2010) Sugar maple seedling carbon assimilation at the northern limit of its range: the importance of seasonal light. *Canadian Journal of Forest Research*, **40**, 385–393.
- Labelle, C. & Richard, P.J.H. (1984) Histoire de la végétation dans la région de Mont-Saint-Pierre, Gaspésie, Québec. Géographie physique et Quaternaire, 38, 257–274.
- Lamy, S., Bouchard, A. & Simon, J.-P. (1999) Genetic structure, variability, and mating system in eastern white cedar (*Thuja occidentalis*) populations of recent origin in an agricultural landscape in southern Quebec. *Canadian Journal of Forest research*, 29, 1383–1392.
- Langella, O. (1999) Populations version 1.2.31. Population genetic software (individuals or populations distances, phylogenetic trees). Available at: http://bioinformatics.org/~tryphon/populations/ (accessed 7 October 2013).
- Lara-Gomez, G., Gailing, O. & Finkeldey, R. (2005) Genetic variation in isolated Mexican populations of the endemic maple *Acer skutchii* Rehd. *Allgemeine Forst und Jagdzeitung*, **176**, 97–103.
- Lavoie, M. & Richard, P.J.H. (2000) Postglacial water-level changes of a small lake in southern Québec, Canada. *The Holocene*, **10**, 621–634.
- Ledig, F.T. & Korbobo, D.R. (1983) Adaptation of sugar maple populations along altitudinal gradients: photosynthesis, respiration, and specific leaf weight. *American Journal of Botany*, **70**, 256–265.
- Lesser, M.R., Parchman, T.L. & Jackson, S.T. (2013) Development of genetic diversity, differentiation and structure over 500 years in four ponderosa pine populations. *Molecular Ecology*, **22**, 2640–2652.
- Li, Y.-C., Korol, A.B., Fahima, T., Beiles, A. & Nevo, E. (2002) Microsatellites: genomic distribution, putative functions and mutational mechanisms: a review. *Molecular Ecology*, **11**, 2453–2465.

- Lind, J. & Gailing, O. (2013) Genetic structure of *Quercus rubra* L. and *Quercus ellipsoidalis* E. J. Hill populations at gene-based EST-SSR and nuclear SSR markers. *Tree Genetics & Genomes*, **9**, 707–722.
- Little, E.L., Jr. (1971) Conifers and important hardwoods. *Atlas of United States trees*, pp. 1–9, 200 maps. U.S. Department of Agriculture Miscellaneous Publication 1146.
- Liu, K.-B. (1990) Holocene paleoecology of the boreal forest and Great Lakes-St. Lawrence forest in northern ontario. *Ecological Monographs*, **60**, 179–212.
- Lloyd, A.H. & Fastie, C.L. (2003) Recent changes in treeline forest distribution and structure in interior Alaska. *Ecoscience*, **10**, 176–185.
- Logan, K.T. (1965) Growth of tree seedlings as affected by light intensity, I: White birch, yellow birch, sugar maple and silver maple. Department of Forestry Publication 1121. Government of Canada, Department of Forestry, Chalk River, ON.
- Luikart, G. & Cornuet, J.-M. (1998) Empirical evaluation of a test for identifying recently bottlenecked populations from allele frequency data. *Conservation Biology*, **12**, 228–237.
- Lund, S.T., Furnier, G.R. & Mohn, C.A. (1992) Isozyme variation in quaking aspen in Minnesota. *Canadian Journal of Forest Research*, **22**, 521–524.
- Majcen, Z., Richard, Y. & Ménard, M. (1984) Ecologie et dendrométrie dans le sudouest du Québec, étude de douze secteurs forestiers. Mémoire no. 85.
  Gouvernement du Québec, Ministère de l'Energie et des Ressources, Service de la recherche (terre et forêt). Québec, Canada.
- Majcen, Z. (1994) Historique des coupes de jardinage dans les forêts inéquiennes au Québec. *Revue Forestiere Francaise (Nancy)*, **4**, 375–385.
- Malausa, T., Gilles, A., Meglécz, E., Blanquart, H., Duthoy, S., Costedoat, C., Dubut, V., Pech, N., Castagnone-Sereno, P., Délye, C., Feau, N., Frey, P., Gauthier, P., Guillemaud, T., Hazard, L., Le Corre, V., Lung-Escarmant, B., Malé, P.-J.G., Ferreira, S. & Martin, J.-F. (2011) High-throughput microsatellite isolation through 454 GS-FLX Titanium pyrosequencing of enriched DNA libraries. *Molecular Ecology Resources*, 11, 638–644.
- Mariac, C., Scarcelli, N., Pouzadou, J., Barnaud, A., Billot, C., Faye, A., Kougbeadjo, A., Maillol, V., Martin, G., Sabot, F., Santoni, S., Vigouroux, Y. & Couvreur,

T.L.P. (2014) Cost-effective enrichment hybridization capture of chloroplast genomes at deep multiplexing levels for population genetics and phylogeography studies. *Molecular Ecology Resources*, Published online, doi: 10.1111/1755-0998.12258

- Marquardt, P.E. & Epperson, B.K. (2004) Spatial and population genetic structure of microsatellites in white pine. *Molecular Ecology*, 13, 3305–3315.
- Marshall, C. & Brown, J.H. (1975) Optimum sampling strategies for gene conservation. Crop genetic resources for today and tomorrow (eds. O.H. Frenkel, J.G. Hawkes), pp.53–80. Cambridge University press, Cambridge, UK.
- Mazerolle, M. (2011) *AICcmodavg: model selection and multimodel inference based on* (*Q*)*AIC(c)*. R package version 1.19. Available at: http://cran.rproject.org/web/packages/AICcmodavg/index.html.
- McCarragher, S.R., Goldblum, D. & Rigg, L.S. (2011) Geographic variation of germination, growth, and mortality in sugar maple (*Acer saccharum*): common garden and reciprocal dispersal experiments. *Physical Geography*, 32, 1–21.
- McKenney, D.W., Pedlar, J.H., Lawrence, K., Campbell, K. & Hutchinson, M.F. (2007) Beyond traditional hardiness zones: using climate envelopes to map plant range limits. *BioScience*, 57, 929–937.
- McLachlan, J.S., Clark, J.S. & Manos, P.S. (2005) Molecular indicators of tree migration capacity under rapid climate change. *Ecology*, **86**, 2088–2098.
- McMahon, S.M., Harrison, S.P., Armbruster, W.S., Bartlein, P.J., Beale, C.M., Edwards, M.E., Kattge, J., Midgley, G., Morin, X. & Prentice, I.C. (2011) Improving assessment and modelling of climate change impacts on global terrestrial biodiversity. *Trends in Ecology and Evolution*, 26, 249–259.
- Meng, L., Yang, R.U.I., Abbott, R.J., Miehe, G., Hu, T. & Liu, J. (2007) Mitochondrial and chloroplast phylogeography of *Picea crassifolia* Kom. (Pinaceae) in the Qinghai-Tibetan Plateau and adjacent highlands. *Molecular Ecology*, 16, 4128–4137.
- Ministère des Ressources Naturelles de la Faune et des Parcs du Québec (MRNFPQ) (2003) *Manuel d'aménagement forestier*. Availaible at: http://www.mrn.gouv.qc.ca/publications/forets/amenagement/manuel.pdf (accessed 17 June 2014).

- Ministère des Ressources Naturelles du Québec (MRNQ) (2013a) Données numériques écoforestières. Available at: http://www.mrn.gouv.qc.ca/forets/connaissances/connaissances-inventairecartes-donnees.jsp (accessed 30 May 2013).
- Ministère des Ressources Naturelles du Québec (MRNQ) (2013b) Les écosystèmes forestiers exceptionnels: éléments clés de la diversité biologique du Québec. Available at: http://www.mrn.gouv.qc.ca/forets/connaissances/connaissancesecosystemes.jsp (accessed 24 June 2013).
- Ministère des Ressources Naturelles du Québec (MRNQ) (2013c) Système d'Information FORestière par Tesselle (SIFORT) – données et couverture cartographique. Available at: http://www.mrn.gouv.qc.ca/forets/connaissances/connaissances-inventairecartes-sifort.jsp (accessed 30 May 2013).
- Moore, J.-D., Camiré, C. & Ouimet, R. (2000) Effects of liming on the nutrition, vigor, and growth of sugar maple at the Lake Clair Watershed, Québec, Canada. *Canadian Journal of Forest Research*, **30**, 725–732.
- Morin, X. & Thuiller, W. (2009) Comparing niche- and process-based models to reduce prediction uncertainty in species range shifts under climate change. *Ecology*, **90**, 1301–1313.
- Mulcahy, D.L. (1975) Differential mortality among cohort in a population of *Acer* saccharum (Aceraceae) seedlings. American Journal of Botany, **62**, 422–426.
- Namroud, M.-C., Park, A., Tremblay, F. & Bergeron, Y. (2005) Clonal and spatial genetic structures of aspen (*Populus tremuloides* Michx.). *Molecular Ecology*, 14, 2969–2980.
- Namroud, M.-C., Beaulieu, J., Juge, N., Laroche, J. & Bousquet, J. (2008) Scanning the genome for gene single nucleotide polymorphisms involved in adaptive population differentiation in white spruce. *Molecular Ecology*, **17**, 3599– 3613.
- Nanson, A. (2004) *Génétique et amélioration des arbres forestiers*. Les Presses Agronomiques de Gembloux, Gembloux, Belgique. 712 p.
- Nei, M. (1972) Genetic distance between populations. *The American Naturalist*, **106**, 283–292.

- Nei, M., Maruyama, T. & Chakrabordy, R. (1975) The bottleneck effect and genetic variability in populations. *Evolution*, **29**, 1–10.
- Nicholson, S.A., Jon, T.S. & Breisch, A.R. (1979) Structure and succession in the tree stratum at Lake George, New York. *Ecology*, **60**, 1240–1254.
- Nijensohn, S.E., Schaberg, P.G., Hawley, G.J. & Dehayes, D.H. (2005) Genetic subpopulation structuring and its implications in a mature eastern white pine stand. *Canadian Journal of Forest Research*, **35**, 1041–1052.
- Newton, A.C., Allnutt, T.R., Gillies, A.C.M., Lowe, A.J. & Ennos, R.A. (1999) Molecular phylogeography, intraspecific variation and the conservation of tree species. *Trends in Ecology & Evolution*, 14, 140–145.
- Nowacki, G.J. & Abrams, M.D. (2008) The demise of fire and "mesophication" of forests in the eastern United States. *Bioscience*, **58**, 123–138.
- Nybom, H. (2004) Comparison of different nuclear DNA markers for estimating intraspecific genetic diversity in plants. *Molecular Ecology*, **13**, 1143–1155.
- O'Connell, L.M., Mosseler, A. & Rajora, O.P. (2006) Impacts of forest fragmentation on the mating system and genetic diversity of white spruce (*Picea glauca*) at the landscape level. *Heredity*, **97**, 418–426.
- Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., O'Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H. & Wagner, H. (2011) vegan: community ecology package. R package version 1.17-10. Available at: http://cran.rproject.org/web/packages/vegan/index.html.
- Ouimet, R., Duchesne, L., Houle, D. & Arp, P.A. (2001) Critical loads and exceedances of acid deposition and associated forest growth in the northern hardwood and boreal coniferous forests in Québec, Canada. *Water, Air, & Soil Pollution: Focus*, **1**, 119–134.
- Page, R.D.M. (1996) Treeview: an application to display phylogenetic trees on personal computers. *Computer Applications in the Biosciences*, **12**, 357–358.
- Pandey, M., Gailing, O., Fischer, D., Hattemer, H.H. & Finkeldey, R. (2004) Characterization of microsatellite markers in sycamore (*Acer pseudoplatanus* L.). *Molecular Ecology Notes*, 4, 253–255.
- Pandey, M. (2005) Development of microsatellites in sycamore maple (*Acer pseudoplatanus* L.) and their application in population genetics. PhD thesis,

Forestry Science at the Faculty of Forest Sciences and Forest Ecology, Georg-August University of Göttingen, Göttingen, Germany.

- Pandey, M. & Rajora, O. (2012a) Higher fine-scale genetic structure in peripheral than in core populations of a long-lived and mixed-mating conifer eastern white cedar (*Thuja occidentalis* L.). *BMC Evolutionary Biology*, **12**, 48.
- Pandey, M. & Rajora, O.P. (2012b) Genetic diversity and differentiation of core vs. peripheral populations of eastern white cedar, *Thuja occidentalis* (*Cupressaceae*). *American Journal of Botany*, **99**, 690–699.
- Parmesan, C., Gaines, S., Gonzalez, L., Kaufman, D.M., Kingsolver, J., Townsend Peterson, A. & Sagarin, R. (2005) Empirical perspectives on species borders: from traditional biogeography to global change. *Oikos*, **108**, 58–75.
- Paul, V.R., Bergeron, Y. & Tremblay, F. (2014) Does climate control the northern range limit of eastern white cedar (*Thuja occidentalis* L.)? *Plant Ecology*, 215, 181–194.
- Pautasso, M. (2009) Geographical genetics and the conservation of forest trees. *Perspectives in Plant Ecology, Evolution and Systematics*, **11**, 157–189.
- Payette, S., Fortin, M.-J. & Morneau, C. (1996) The recent sugar maple decline in southern Québec : probable causes deduced from tree rings. *Canadian Journal of Forest Research*, 26, 1069–1078.
- Peakall, R.O.D. & Smouse, P.E. (2006) GenAlEx 6: genetic analysis in Excel. Population genetic software for teaching and research. *Molecular Ecology Notes*, 6, 288–295.
- Peñalba, J.V., Smith, L.L., Tonione, M.A., Sass, C., Hykin, S.M., Skipwith, P.L., Mcguire, J.A., Bowie, R.C.K. & Moritz, C. (2014) Sequence capture using pcr-generated probes: A cost-effective method of targeted high-throughput sequencing for nonmodel organisms. *Molecular Ecology Resources*, Published online, doi: 10.1111/1755-0998.12249
- Perry, D.J. & Bousquet, J. (2001) Genetic diversity and mating system of post-fire and post-harvest black spruce: An investigation using codominant sequencetagged-site (sts) markers. *Canadian Journal of Forest Research*, **31**, 32–40.
- Perry, D.J. & Knowles, P. (1989) Allozyme variation in sugar maple at the northern limit of its range in Ontario, Canada. *Canadian Journal of Forest Research*, 19, 509–514.

- Perry, D.J. & Knowles, P. (1991) Spatial genetic structure within three sugar maple (*Acer saccharum* Marsh.) stands. *Heredity*, **66**, 137–142.
- Petit, R.J., Duminil, J., Fineschi, S., Hampe, A., Salvini, D. & Vendramin, G.G. (2005) Comparative organization of chloroplast, mitochondrial and nuclear diversity in plant populations. *Molecular Ecology*, 14, 689–701.
- Pierce, A., Bromer, W. & Rabenold, K. (2008) Decline of *Cornus florida* and forest succession in a *Quercus–Carya* forest. *Plant Ecology*, **195**, 45–53.
- Pilon, C.E., Côté, B. & Fyles, J.W. (1994) Effect of snow removal on leaf water potential, soil moisture, leaf and soil nutrient status and leaf peroxidase activity of sugar maple. *Plant and Soil*, 162, 81–88.
- Pilon, V.J. (2013) Dynamique à long terme des érablières à leur limite nordique de répartition au Québec. Mémoire de maîtrise en biologie, Université Laval, Québec. 65pp.
- Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D. & Team, R.D.C. (2011) *nlme: linear and nonlinear mixed effects models*. R package version 3.1-102. Available at: http://cran.r-project.org/web/packages/nlme/index.html.
- Piry, S., Luikart, G. & Cornuet, J.-M. (1999) Bottleneck: a computer program for detecting recent reductions in the effective population size using allele frequency data. *Journal of Heredity*, **90**, 502–503.
- Prentice, I.C. & Jolly, D. (2000) Mid-Holocene and glacial-maximum vegetation geography of the northern continents and Africa. *Journal of Biogeography*, 27, 507–519.
- Pritchard, J.K., Stephens, M. & Donnelly, P. (2000) Inference of population structure using multilocus genotype data. *Genetics*, **155**, 945–959.
- R Development Core Team (2011) *R: a language and environment for statistical computing*. Version 2.13.1 (2011-07-08). R Foundation for Statistical Computing, Vienna, Austria. Available at: http://cran.r-project.org/.
- Rainville, A. (2007) Forest genetic resource conservation issues in Québec. *In* proceedings of the forum on conservation of forest genetic resources: Challanges, issues, and solutions. 28-29 July 2006 Charlottetown, PEI. Compiled by I.D. Simpson. Information Report M-X-220. Natural Resources Canada, Canadian Forest Service, Atlantic Forest Centre, Frederictown, NB, Canada.

- Rajora, O.P. (1999) Genetic biodiversity impacts of silvicultural practices and phenotypic selection in white spruce. *Theoretical and Applied Genetics*, **99**, 954–961.
- Rajora, O.P., Deverno, L., Mosseler, A. & Innes, D.J. (1998) Genetic diversity and population structure of disjunct newfoundland and central ontario populations of eastern white pine (*Pinus strobus*). *Canadian Journal of Botany*, **76**, 500– 508.
- Rajora, O.P., Rahman, M.H., Buchert, G.P. & Dancik, B.P. (2000) Microsatellite DNA analysis of genetic effects of harvesting in old-growth eastern white pine (*Pinus strobus*) in Ontario, Canada. *Molecular Ecology*, **9**, 339–348.
- Régnière, J. (1996) Generalized approach to landscape-wide seasonal forecasting with temperature-driven simulation models. *Environmental Entomology*, **25**, 869–881.
- Rice, W.R. (1989) Analysing tables of statistical tests. Evolution, 43, 223–225.
- Richard, P. (1980) Histoire postglaciaire de la végétation au sud du lac Abitibi, Ontario et Québec. *Géographie physique et Quaternaire*, **34**, 77–94.
- Richard, P.J.H. (1993) Origine et dynamique postglaciaire de la forêt mixte au Québec. *Review of Palaeobotany and Palynology*, **79**, 31–68.
- Richard, P.J.H. & Grondin, P. (2009) Histoire postglaciaire de la végétation. *Manuel de foresterie. Chapitre 4*, pp. 170-176. Ordre des ingénieurs forestiers du Québec, 2è édition, ouvrage collectif, éditions multimondes, Québec, 1510 p.
- Robichaud, R.L., Glaubitz, J.C., Rhodes, O.E. & Woeste, K. (2010) Genetic consequences of harvest in a mature second-growth stand of black walnut (*Juglans nigra* L.). *Annals of Forest Science*, **67**, 702.
- Robitaille, L. & Pardé, J. (1979) La sylviculture des érables à sucre. *Revue Forestiere Francaise (Nancy)*, **XXXI**, 205–211.
- Rosenberg, N.A. (2004) Distruct: A program for the graphical display of population structure. *Molecular Ecology Notes*, **4**, 137–138.
- Rousset, F. (2008) Genepop'007: a complete re-implementation of the genepop software for Windows and Linux. *Molecular Ecology Resources*, **8**, 103–106.

- Rusanen, M., Vakkari, P. & Blom, A. (2000) Evaluation of the finnish geneconservation strategy for norway maple (acer platanoides l.) in the light of allozyme variation. *Forest Genetics*, 7, 155–165.
- Rusanen, M., Vakkari, P. & Blom, A. (2003) Genetic structure of *Acer platanoides* and *Betula pendula* in northern Europe. *Canadian Journal of Forest Research*, **33**, 1110–1115.
- Saeki, I., Dick, C.W., Barnes, B.V. & Murakami, N. (2011) Comparative phylogeography of red maple (*Acer rubrum* L.) and silver maple (*Acer saccharinum* L.): impacts of habitat specialization, hybridization and glacial history. *Journal of Biogeography*, **38**, 992–1005.
- Salk, T.T., Frelich, L.E., Sugita, S., Calcote, R., Ferrari, J.B. & Montgomery, R.A. (2011) Poor recruitment is changing the structure and species composition of an old-growth hemlock-hardwood forest. *Forest Ecology and Management*, 261, 1998–2006.
- Saucier, J.-P., Grondin, P., Robitaille, A. & Bergeron, J.-F. (2003) Zones de végétation et domaines bioclimatiques du Québec. Publication No. 2003-3015. Ministère des Ressources naturelles, de la Faune et des Parcs (MRNFP), direction des inventaires forestiers. Québec, Canada. Available at: http://www.mrnf.gouv.qc.ca/publications/forets/connaissances/zonevegetation-2003.pdf (accessed 7 August 2012).
- Saunders, D.A., Hobbs, R.J. & Margules, C.R. (1991) Biological consequences of ecosystem fragmentation: a review. *Conservation Biology*, **5**, 18–32.
- Seeb, J.E., Carvalho, G., Hauser, L., Naish, K., Roberts, S. & Seeb, L.W. (2011) Single-nucleotide polymorphism (snp) discovery and applications of snp genotyping in nonmodel organisms. *Molecular Ecology Resources*, 11, 1–8.
- Segarra-Moragues, J., Gleiser, G. & González-Candelas, F. (2008) Isolation and characterization of microsatellite loci in *Acer opalus* (Aceraceae), a sexually-polymorphic tree, through an enriched genomic library. *Conservation Genetics*, **9**, 1059–1062.
- Shaw, J. & Small, R.L. (2005) Chloroplast DNA phylogeny and phylogeography of the North American plums (*Prunus* subgenus *Prunus* section *Prunocerasus*, Rosaceae). *American Journal of Botany*, **92**, 2011–2030.

- Shih, C.Y., Dumbroff, E.B. & Peterson, C.A. (1985) Developmental studies of the stratification–germination process in sugar maple embryos. *Canadian Journal of Botany*, **63**, 903–908.
- Siccama, T.G. (1971) Presettlement and present forest vegetation in northern Vermont with special reference to Chittenden County. *American Midland Naturalist*, **85**, 153–172.
- Simon, J.-P., Payette, Y. & Longpré, M.-H. (1995) Comparative analysis of the genetic composition of canopy and juvenile sugar maple individuals (*Acer* saccharum) in an old-growth forest in southern Québec as related to anthropogenic disturbance. Canadian Journal of Forest Research, 25, 743– 752.
- Skepner (1998) RAPD reveals genetic similarity of *Acer saccharum* and *Acer nigrum*. *Heredity*, **80**, 422–428.
- Skepner, A.P. & Krane, D.E. (1997) cpDNA of *Acer saccharum* and *Acer nigrum* are very similar. *Ohio Journal of science*, **97**, 90–93.
- Soltis, D.E., Morris, A.B., McLachlan, J.S., Manos, P.S. & Soltis, P.S. (2006) Comparative phylogeography of unglaciated eastern North America. *Molecular Ecology*, **15**, 4261–4293.
- Sork, V., Huang, S. & Wiener, E. (1993) Macrogeographic and fine-scale genetic structure in a north American oak species, *Quercus rubra* L. *Annales des sciences forestières*, **50**, 261s–270s.
- Spencer, C.C., Neigel, J.E. & Leberg, P.L. (2000) Experimental evaluation of the usefulness of microsatellite DNA for detecting demographic bottlenecks. *Molecular Ecology*, **9**, 1517–1528.
- Stapley, J., Reger, J., Feulner, P.G.D., Smadja, C., Galindo, J., Ekblom, R., Bennison, C., Ball, A.D., Beckerman, A.P. & Slate, J. (2010) Adaptation genomics: the next generation. *Trends in Ecology & Evolution*, 25, 705–712.
- Sunnucks, P. (2000) Efficient genetic markers for population biology. *Trends in Ecology & Evolution*, **15**, 199–203.
- Taberlet, P., Gielly, L., Pautou, G. & Bouvet, J. (1991) Universal primers for amplification of three non-coding regions of chloroplast DNA. *Plant Molecular Biology*, 17, 1105–1109.

- Tachiki, Y. & Iwasa, Y. (2010) Both seedling banks and specialist seed predators promote the evolution of synchronized and intermittent reproduction (masting) in trees. *Journal of Ecology*, **98**, 1398–1408.
- Takayama, K., Sun, B.-Y. & Stuessy, T.F. (2012) Genetic consequences of anagenetic speciation in *Acer okamotoanum* (Sapindaceae) on Ullung Island, Korea. *Annals of Botany*, **109**, 321–330.
- Takayama, K., Sun, B.-Y. & Stuessy, T.F. (2013) Anagenetic speciation in Ullung island, Korea: genetic diversity and structure in the island endemic species, *Acer takesimense* (Sapindaceae). *Journal of Plant Research*, **126**, 323–333.
- Tremblay, M.F., Bergeron, Y., Lalonde, D. & Mauffette, Y. (2002) The potential effects of sexual reproduction and seedling recruitment on the maintenance of red maple (*Acer rubrum* L.) populations at the northern limit of the species range. *Journal of Biogeography*, 29, 365–373.
- Ueno, S., Klopp, C., Leple, J.C., Derory, J., Noirot, C., Leger, V., Prince, E., Kremer, A., Plomion, C. & Le Provost, G. (2013) Transcriptional profiling of bud dormancy induction and release in oak by next-generation sequencing. *BMC Genomics*, 14, 236.
- Van Tienderen, P.H., De Haan, A.A., Van Der Linden, C.G. & Vosman, B. (2002) Biodiversity assessment using markers for ecologically important traits. *Trends in Ecology & Evolution*, 17, 577–582.
- Vavrek, M.J. (2012) Fossil: Palaeoecological and palaeogeographical analysis tools. R package version 0.3.7. Available at: http://cran.rproject.org/web/packages/fossil/index.html.
- Vranckx, G., Jacquemyn, H., Muys, B. & Honnay, O. (2012) Meta-analysis of susceptibility of woody plants to loss of genetic diversity through habitat fragmentation. *Conservation Biology*, 26, 228–237.
- Vucetich, J.A. & Waite, T.A. (2003) Spatial patterns of demography and genetic processes across the species' range: null hypotheses for landscape conservation genetics. *Conservation Genetics*, **4**, 639–645.
- Walck, J.L., Hidayati, S.N., Dixon, K.W., Thompson, K. & Poschlod, P. (2011) Climate change and plant regeneration from seed. *Global Change Biology*, 17, 2145–2161.

- Wang, B.S.P. (1974) *Tree-seed storage*. Department Publication No. 1335. Environment Canada, Canadian Forestry Service, Petawawa Forest Experiment Station, Chalk River, ON.
- Wang, Z.-S., Liu, H., Wei, N., Xu, W.-X. & An, S.-Q. (2010) Contribution of progeny cohorts to the restoration of genetic diversity in the post-harvest dragon spruce (*Picea asperata*) stands. *Forestry*, 83, 307–314.
- Waters, J.M., Fraser, C.I. & Hewitt, G.M. (2013) Founder takes all: Densitydependent processes structure biodiversity. *Trends in Ecology & Evolution*, 28, 78–85.
- Williams, J.W., Webb, T., Richard, P.H. & Newby, P. (2000) Late quaternary biomes of Canada and the eastern United States. *Journal of Biogeography*, 27, 585– 607.
- Wolfe, K.H., Li, W.-H. & Sharp, P.M. (1987) Rates of nucleotide substitution vary greatly among plant mitochondrial, chloroplast, and nuclear dnas. *Proceedings of the National Academy of Sciences*, **84**, 9054–9058.
- Woodward, F.I. (1987) *Climate and plant distribution*. Cambridge University Press, Cambridge, UK.
- Wright, J.W. (1976) *Introduction to forest genetics*. New York Academic Press, New York, USA. 463p.
- Wyman, J., Bruneau, A. & Tremblay, M.-F. (2003) Microsatellite analysis of genetic diversity in four populations of populus tremuloides in Quebec. *Canadian Journal of Botany*, 81, 360–367.
- Xu, H., Tremblay, F., Bergeron, Y., Paul, V. & Chen, C. (2012) Genetic consequences of fragmentation in "arbor vitae," eastern white cedar (*Thuja* occidentalis L.), toward the northern limit of its distribution range. Ecology and Evolution, 2, 2506–2520.
- Young, A., Boyle, T. & Brown, T. (1996) The population genetic consequences of habitat fragmentation for plants. *Trends in Ecology & Evolution*, **11**, 413–418.
- Young, A., Merriam, H.G. & Warwick, S.I. (1993a) The effects of forest fragmentation on genetic variation in *Acer saccharum* Marsh. (sugar maple) populations. *Heredity*, **71**, 277–289.

- Young, A.G. & Merriam, H.G. (1994) Effects of forest fragmentation on the spatial genetic structure of *Acer saccharum* Marsh. (sugar maple) population. *Heredity*, **72**, 201–208.
- Young, A.G., Warwick, S.I. & Merriam, H.G. (1993b) Genetic variation and structure at three spatial scales for *Acer saccharum* (sugar maple) in Canada and the implications for conservation. *Canadian Journal of Forest research*, 23, 2568–2578.
- Zhao, L., Sun, W. & Yang, J. (2011) Development and characterization of microsatellite markers in the critically endangered species *Acer yangbiense* (Aceraceae). *American Journal of Botany*, **98**, e247–e249.
- Zhu, K., Woodall, C.W. & Clark, J.S. (2012) Failure to migrate: lack of tree range expansion in response to climate change. *Global Change Biology*, **18**, 1042–1052.